

FORUM

Invasions by *Harmonia axyridis* (Pallas) (Coleoptera: Coccinellidae) in the Western Hemisphere: Implications for South AmericaROBERT L. KOCH¹, ROBERT C. VENETTE² AND WILLIAM D. HUTCHISON¹¹Dept. Entomology, Univ. Minnesota, 219 Hodson Hall, 1980 Folwell Avenue, St. Paul, Minnesota, 55108, USA²North Central Research Station, USDA Forest Service, 1561 Lindig Avenue, St. Paul, MN 55108, USA*Neotropical Entomology* 35(4):421-434 (2006)Invasões de *Harmonia axyridis* (Pallas) (Coleoptera: Coccinellidae) no Hemisfério Ocidental: Implicações para a América do Sul

RESUMO - A joaninha, *Harmonia axyridis* (Pallas), nativa da Ásia, foi recentemente detectada na América do Sul depois de ter invadido a América do Norte e Europa. Essa joaninha é um predador voraz, e portanto, popular e eficaz no controle biológico. Infelizmente, *H. axyridis* também está relacionada a impactos nocivos (ex., como peste residencial e de frutas temperadas e ameaça a organismos não-alvos). Para fazer prever os possíveis impactos de *H. axyridis* na América do Sul, a história da sua invasão no Hemisfério Ocidental foi revisada e os vários fatores críticos para futuras invasões (isto é, chegada, estabelecimento e disseminação) em novas áreas da América do Sul foram discutidos. A possibilidade de introduções contínuas de *H. axyridis* na América do Sul parece alta devido a sua popularidade como agente de controle biológico e através de introduções acidentais. Seu estabelecimento também parece possível em extensas regiões da América do Sul. A similaridade climática com a região nativa sugere que o estabelecimento é possível na região sul da América do Sul. Porém, similaridade de hábitat com a região nativa sugere que o estabelecimento seja mais adequado na região norte da América do Sul. Além disso, a disponibilidade da presa não deve ser um fator limitante para o estabelecimento desse predador. Após o estabelecimento, *H. axyridis* pode se disseminar pelo próprio voo e por meios associados ao homem. Concluindo, a invasão de *H. axyridis* em novas áreas do continente sul-americano é provável.

PALAVRAS-CHAVE: Controle biológico, espécie invasiva, espécie exótica, similaridade climática, impacto não-alvo

ABSTRACT - The multicolored Asian lady beetle, *Harmonia axyridis* (Pallas), native to Asia, has recently been detected in South America after successfully invading North America and Europe. This coccinellid is a voracious predator; therefore, it is popular and effective in biological control. Unfortunately, *H. axyridis* also has associated adverse impacts (i.e., as a household pest, pest of fruit production, and threat to non-target organisms). To predict the potential geographic extent of impacts of *H. axyridis* in South America we review the history of its invasion in the Western Hemisphere and address various factors critical to the future invasion (i.e., arrival, establishment, and spread) of new areas of South America. The likelihood of continued introductions (i.e., arrival) of *H. axyridis* to South America seems high, due to its popularity as a biological control agent and through accidental introductions. Establishment also seems likely in broad regions of South America. Climate matching with the native range suggested that much of southern South America may be suitable for establishment. In contrast, habitat matching with the native range suggested that northern South America may be more suitable. In addition, prey availability should not limit establishment of this predator. Once established, *H. axyridis* seems likely to spread by flight and human-assisted means. Overall, the invasion of *H. axyridis* over broad areas in South America seems likely.

KEY WORDS: Biological control, invasive species, exotic species, climate matching, non-target impact

The multicolored Asian lady beetle, *Harmonia axyridis* (Pallas), is a dramatic example of a successful invader. This coccinellid, native to Asia, has become one of the most intensively studied insect predators (Sloggett 2005), with much of this work reviewed by Koch (2003) and Pervez & Omkar (2006). *H. axyridis* is a voracious predator of aphids and other soft-bodied insects and has been utilized in numerous biological control programs. This coccinellid is frequently associated with trees in natural and agricultural settings when prey is available (Table 1). In addition, this semi-arboreal predator also occurs in various herbaceous habitats, including agricultural (reviewed by Koch 2003) and natural (Sebolt & Landis 2004, RLK unpublished data) systems.

Along with beneficial impacts as a biological control agent, *H. axyridis* has shown three general adverse impacts. First, as with many other exotics, *H. axyridis* may threaten native organisms. For example, densities of native predators seem to have decreased as the abundance of *H. axyridis* increased (Colunga-Garcia & Gage 1998, Brown & Miller 1998, Michaud 2002b, Alyokhin & Sewell 2004, Saini 2004), which may be partly due to intraguild predation (e.g., Cottrell & Yeargan 1998, Michaud 2002b, Cottrell 2004, Yasuda *et al.* 2004). In addition, *H. axyridis* may impact populations of the monarch butterfly, *Danaus plexippus* (Koch *et al.* 2004c, *in press b*). Second, *H. axyridis* can be a pest of fruit production (Koch *et al.* 2004a), particularly as a contaminant during wine production (Pickering *et al.* 2004, Galvan *et al.* 2006). Third, *H. axyridis* can be a nuisance to humans. It can become a household pest when it seeks shelter from winter in homes and other structures (Nalepa *et al.* 2004, 2005). Once on or in a home, massive aggregations of *H. axyridis* are a nuisance to homeowners (Nalepa *et al.* 2004, Huelsman & Kovach 2004) and can cause allergic reactions in humans (Ray & Pence 2004). Similarly, *H. axyridis* can form autumn aggregations in bee hives, where it apparently does not harm the bees, but is a nuisance to the bee keepers (Caron 1996).

The geographic range and impacts (both positive and negative) of *H. axyridis* are expanding rapidly. Currently, this coccinellid is widely established in North America [U.S.A and Canada (Koch 2003)], is established and expanding its range in Europe [France (Hodek & Honik 1996, Lohez 2005), Greece (Katsoyannos *et al.* 1997), Germany (Bathon 2002, Klausnitzer 2002/3) Belgium (Adriaens *et al.* 2003), The Netherlands (Cuppen *et al.* 2004), Switzerland (Klausnitzer 2004), and Britain (Majerus & Roy 2005)], and is established with a restricted distribution in South America [Brazil (Almeida & Silva 2002) and Argentina (Saini 2004)]. Can we use what we have learned from the invasion of *H. axyridis* in North America to predict implications for South America? In this paper, we provide a review of the literature on the history of invasions by this coccinellid in the Western Hemisphere. In addition, we make predictions for the spread and impacts of *H. axyridis* in South America, based on evaluation of the suitability of this continent for the primary components of the invasion process (i.e., arrival, establishment and spread) for *H. axyridis*.

Invasion History

The presumed native distribution of *H. axyridis* extends across southern Siberia from the Altai Mountains to the Pacific Coast, including Korea and Japan, then extends southward to southern China, and includes the Himalayas (Dobzhansky 1933, Chapin 1965, Sasaji 1971, Iablokoff-Khnzorian 1982, Kuznetsov 1997). Numerous color forms (e.g., the *succinea* group and various melanic forms) with a genetic basis (mosaic dominance) occur within this range (Komai 1956). In general, the *succinea* group of *H. axyridis* have an orange-red ground color with or without black spots, whereas numerous melanic forms exist having a black ground color with or without orange-red markings. The relative frequencies of the various forms vary geographically. For instance, the relative frequency of the *succinea* group increases from west to east across the Asian range (Dobzhansky 1933, Komai 1956). West of Lake Baikal in Russia, the *succinea* group is rare, but in eastern Siberia it comprises 80% of the population (Dobzhansky 1933, Komai 1956). Using methods described in Venette and Ragsdale (2004) we determined that the *succinea* group in Asia is generally associated with biomes characterized by Olson *et al.* (2001) as boreal forest, temperate broadleaf and mixed forest, temperate coniferous forest, and tropical-subtropical moist broadleaf forest. The melanic group is most commonly associated with boreal forest. In general, melanic color forms of coccinellids may have advantages over their non-melanic counterparts in colder climates. For instance, because the elytra of melanic forms of *Adalia bipunctata* (L.) have less reflectance than those of non-melanic forms, the melanics are able to attain higher body temperatures relative to ambient temperature than non-melanics when exposed to sunlight. Higher body temperatures during the mating period translate to greater activity, and therefore greater fitness for melanics (Hodek & Honik 1996).

Introduction of *H. axyridis* into non-native regions has been extensive. In the following text, we focus primarily on documented intentional releases. However, undocumented commercial releases, unintentional introductions, and unaided dispersal capacity (i.e., flight) have certainly contributed to the extent of its invasion. Records exist of *H. axyridis* being intentionally introduced to North America (Gordon 1985), Europe (Garcia 1986, Ferran *et al.* 1996, Hodek & Honik 1996, Katsoyannos *et al.* 1997, Trouve *et al.* 1997, Buzzocchi *et al.* 2004), Africa (El-Arnaouty *et al.* 2000, OEPP/EPPO 2002), and South America (Saini 2004). Here, we focus specifically on the invasion history in the Western Hemisphere.

In North America, intentional releases began as early as 1916 in California, with more frequent releases in the U.S.A. and Canada during the 1970's and 1980's (Gordon 1985) into areas with biomes generally classified as temperate broadleaf and mixed forest or temperate coniferous forest (Olson *et al.* 2001). For example, 87,810 individuals were released from 1978-1981 in Byron, Georgia, U.S.A. (Teddars & Schaefer 1994) and 37,852 individuals were released from 1981-1982 in three counties of Washington, U.S.A. (LaMana & Miller 1996). Many of these early releases were targeted

Table 1. Arboreal hosts (trees and shrubs) and associated preys utilized by *H. axyridis*.

Host	Reported prey	Reference
<i>Abies procera</i> Rehder, noble fir	<i>Cinara</i> sp. (Aphididae)	LaMana & Miller 1996
<i>Acacia</i> spp., acacia	<i>Psylla uncatoides</i> (Ferris & Klyver) (Psyllidae)	Leeper & Beardsley 1974
<i>Acer negundo</i> L., boxelder	<i>Periphyllus negundinis</i> (Thomas) (Aphididae)	Koch & Hutchison 2003
<i>Acer saccharum</i> Marsh., sugar maple	<i>Periphyllus testudinaceae</i> (Fernie), <i>Drepanaphis idahoensis</i> Smith & Dilley, <i>Drepanosiphum platanoides</i> (Schränk) (Aphididae)	LaMana & Miller 1996
<i>Betula pendula</i> Roth, European white birch	<i>Callipterinella calipterus</i> (Hartig), <i>Euceraphis betulae</i> (Kalterbach) (Aphididae)	LaMana & Miller 1996
<i>Carya illinoensis</i> (Wangenh.) K. Koch, pecan	<i>Monellia caryella</i> (Fitch.), <i>Monelliopsis pecanis</i> Bissell (Aphididae), <i>Melanocallis caryaefoliae</i> (Davis)	Tedders & Schaefer 1994
<i>Carya</i> sp.	<i>Monellia caryella</i> (Fitch.)	Saini 2004
<i>Castanea crenata</i> Sieb. & Zucc., chestnut	<i>Diaspidiotus</i> (=Comstockaspis) <i>macroporatus</i> (Takagi) (Diaspididae)	Choi <i>et al.</i> 1995a
<i>Citrus</i> spp., citrus	<i>Diaphorina citri</i> Kuwayama (Psyllidae), <i>Toxoptera citricida</i> (Kirkaldy) (Aphididae)	Michaud 1999, 2002a
<i>Fagus sylvatica</i> L., European beech	<i>Phyllaphis fagi</i> (L.) (Aphididae)	LaMana & Miller 1996
<i>Hibiscus syriacus</i> L., hibiscus	<i>Aphis gossypii</i> Glover (Aphididae)	Kindlmann <i>et al.</i> 2000
<i>Juglans regia</i> L., English walnut	<i>Chromaphis juglandicola</i> (Kaltenbach) (Aphididae)	Li 1992
<i>Lagerstroemia indica</i> L.	<i>Tinocallis kahawaluokalani</i> (Kirkaldy) (Aphididae)	Almeida & Silva 2002
<i>Lagerstroemia</i> sp., crape myrtle	<i>Tinocallis kahawaluokalani</i> (Kirkaldy) (Aphididae)	Chapin & Brou 1991
<i>Liriodendron tulipifera</i> L., tuliptree	<i>Illinoia liriodendri</i> Monell (Aphididae)	LaMana & Miller 1996
<i>Magnolia macrophylla</i> Michaux, magnolia	Not specified	Tedders & Schaefer 1994
<i>Malus</i> sp., apple	<i>Aphis spiraeicola</i> Patch (Aphididae)	Brown & Miller 1998
<i>Malus</i> sp., dwarf apple	<i>Aphis pomi</i> DeGeer (Aphididae)	Coderre <i>et al.</i> 1995
<i>Malus</i> sp., crab apple	<i>Aphis spiraeicola</i> Patch (Aphididae)	Chapin & Brou 1991
<i>Pinus densiflora</i> Siebold & Zucc., Japanese red pine	<i>Thecodiplosis japonensis</i> Uchida & Intuye (Cecidomyiidae), <i>Matsucoccus matsumurae</i> (Margarodidae)	Miura <i>et al.</i> 1986 McClure 1986b
<i>Pinus massonia</i> Lamb., Chinese red pine	<i>Matsucoccus matsumurae</i> (Kuwana) (Margarodidae)	Chai 1999
<i>Pinus resinosa</i> Ait., red pine	<i>Matsucoccus resinosa</i> Bean & Goodwin (Margarodidae)	McClure 1986a, 1987
<i>Pinus taeda</i> L., loblolly pine	<i>Eulachnus agilis</i> (Kaltenbach) (Aphididae)	Tedders & Schaefer 1994
<i>Pinus thunbergiana</i> Franco, Japanese black pine	<i>Matsucoccus thunbergiana</i> Miller & Park, <i>Matsucoccus matsumurae</i> (Margarodidae)	Choi <i>et al.</i> 1995b McClure 1986b
<i>Pinus</i> spp.	<i>Eulachnus agilis</i> (Kaltenbach) (Aphididae), <i>Cinara atlantica</i> (Wilson), <i>Cinara pinovora</i> (Wilson) (Aphididae)	Tedders & Schaefer 1994 Almeida & Silva 2002

Continue

Table 1. Continuation

Host	Reported preys	Reference
<i>Podocarpus</i> sp.	<i>Neophyllaphis podocarpi</i> Takahashi (Aphididae)	Tedders & Schaefer 1994
<i>Prunus persica</i> (L.) Batsch, peach	<i>Hyalopterus pruni</i> (Geoffroy), <i>Myzus varians</i> Davidson (Aphididae)	Osawa 2000
<i>Prunus</i> sp., plum	<i>Hyalopterus pruni</i> (Geoffrey) (Aphididae)	LaMana & Miller 1996
<i>Quercus rubra</i> L., northern red oak	<i>Myzocallus occultus</i> Richards (Aphididae)	LaMana & Miller 1996
<i>Rhamnus</i> sp., buckthorn	<i>Aphis glycines</i> Matsumura (Aphididae)	Hesler <i>et al.</i> 2004
<i>Salix sieboldiana</i> Blume, willow	<i>Aphis farinosa yanagicola</i> Matsumura, <i>Tuberolachnus salignus</i> (Gmellin) (Aphididae)	Osawa 2000
<i>Salix koriyanagi</i> Kimura, willow	<i>Chaitophorus horii</i> Takashashi, <i>Tuberolachnus salignus</i> (Gmellin) (Aphididae)	Osawa 2000
<i>Salix</i> sp., willow	<i>Tuberolachnus salignus</i> (Gmellin) (Aphididae)	LaMana & Miller 1996
<i>Sambucus sieboldiana</i> Blume, Japanese elderberry	<i>Aulacorthum magnoliae</i> (Essig & Kuwana) (Aphididae)	Osawa 2000
<i>Spirea thunbergii</i> Sieb. ex Bl.	<i>Aphis spiraecola</i> Patch (Aphididae)	Osawa 2000
<i>Spirea blumei</i> G. Don	<i>Aphis spiraecola</i> Patch (Aphididae)	Osawa 2000
<i>Tilia americana</i> L., American basswood	<i>Eucalypterus tiliae</i> (L.) (Aphididae)	LaMana & Miller 1996
<i>Tsuga</i> spp., hemlock	<i>Adelges tsugae</i> Annand (Adelgidae)	Wallace & Hain 2000
<i>Ulmus americana</i> L., American elm	<i>Tinocallis ulmifolii</i> (Monell) (Aphididae)	Hesler 2003
<i>Zanthoxylum bungeanum</i> Maxim., Bunge prickly-ash	<i>Phenacoccus azaleae</i> Kuwana (Pseudococcidae)	Xie <i>et al.</i> 2004

at the pecan aphid complex, *Melanocallis caryaefoliae* (Davis), *Monellia caryella* (Fitch), and *Monelliopsos pecanis* Bissel (Tedders & Schaeffer 1994) and the pear psylla, *Cacopsylla pyricola* Foerster (LaMana & Miller 1996). In addition, *H. axyridis* was released for biological control purposes in the Mexican states of Chihuahua (Quiñones *et al.* 2001), Colima, and Yucatán, (S.H. Tarango Rivero, personal communication). Aside from intentional introductions, it has been argued that accidental sea-port introductions may have played a role in the arrival of this species to North America (Day *et al.* 1994).

Established populations of *H. axyridis* were first detected in North America in 1988 in southeastern Louisiana, U.S.A. (Chapin & Brou 1991). The range of *H. axyridis* then expanded rapidly in North America, but spread did not radiate uniformly from the point of first detection (Fig. 1). The beetle population seems to have moved first eastward then to the north. We conjecture that the arrival of *H. axyridis* in Oregon in 1991 resulted from accidental or intentional releases, not from its unaided dispersal from the southeastern U.S.A. For instance, this predator was commercially available as a biological control agent in North America (Heimpel & Lundgren 2000) and was intentionally redistributed within the continent for biological control purposes, with well documented intentional movement of *H. axyridis* from the southeastern U.S.A. to California and New Mexico, U.S.A in 1992 and 1993 and Texas, U.S.A. in

1994 (Tedders & Schaefer 1994). It is interesting to note that the extensive spread of *H. axyridis* in North America has occurred despite relatively narrow genetic differentiation of the invading populations (Krafsur *et al.* 1997).

Currently, *H. axyridis* is nearly ubiquitous in the eastern U.S.A. and southern portions of the eastern Canadian provinces where beetle abundance can be remarkably high (Fig. 1). Likewise, in western North America, this species is distributed along the northern Pacific coast of the U.S.A. and southern coastal area of British Columbia, Canada (Fig. 1). We are currently unaware of *H. axyridis* being present in Montana, (M. Ivie, personal communication), Wyoming (S. Shaw, personal communication), Arizona (D.N. Byrne, personal communication), Hawaii, U.S.A. (F. Howarth & M.G. Wright, personal communications), or Saskatchewan, Canada (J. Acorn, personal communication). The beetle has been detected in Alberta, Canada but is not known to be established there (J. Acorn, personal communication). In addition, *H. axyridis* is not known to occur in New Mexico, U.S.A, except for a small, irrigated area near the city of Las Cruces (J. Ellington, personal communication). *Harmonia axyridis* is also present in the following Mexican cities where no intentional releases have been documented: Monticello, Estado de México; Zaragoza, Coahuila; Guadalajara, Jalisco; Cuernavaca, Morelos; and Huejotzingo, Puebla, (S.H. Tarango, personal communication). Densities and impacts of *H. axyridis* in Mexico have yet to be reported. Interestingly,

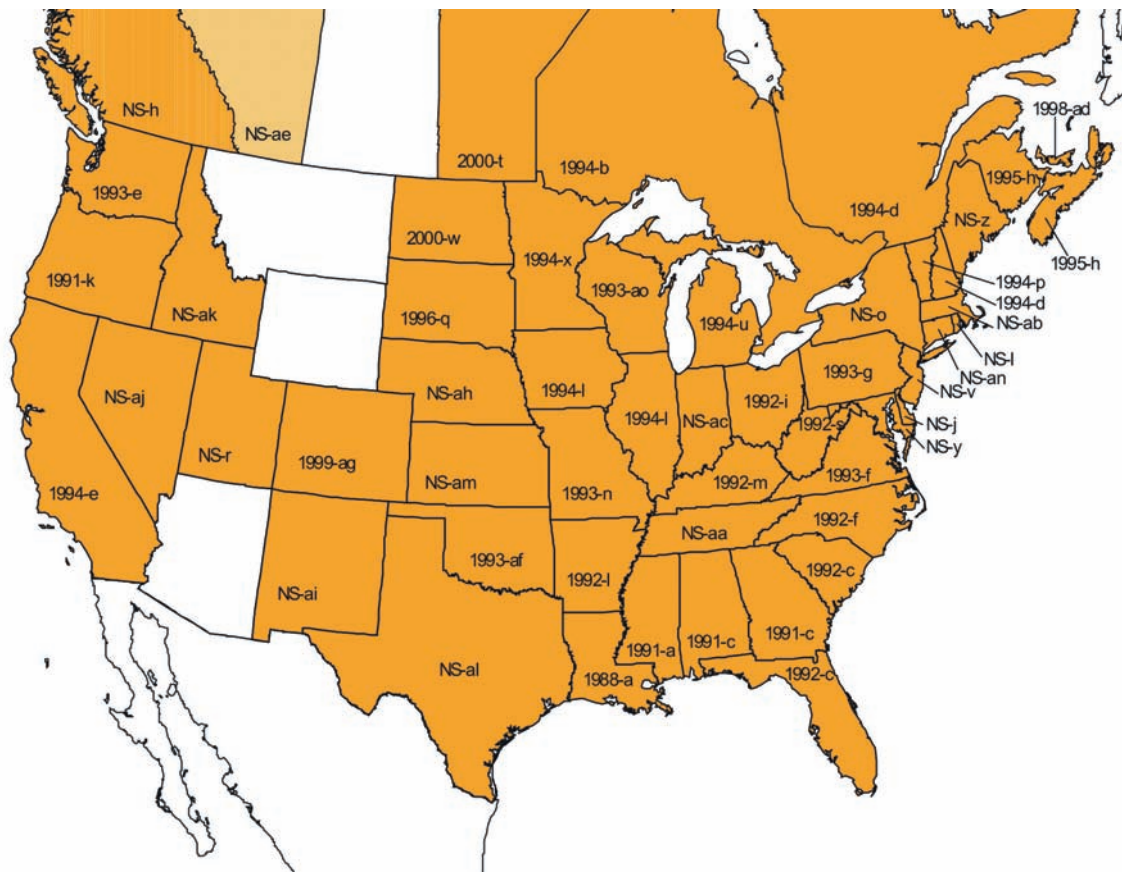


Figure 1. Year of first detection of *H. axyridis* in Canadian provinces and US states. In Alberta, *H. axyridis* has been detected but is not known to be established. (NS-Not specified; a-Chapin & Brou 1991; b-Ker 1994; c-Tedders & Schaffer 1994; d-Corderre *et al.* 1995; e-Driestadt *et al.* 1995; f-Kidd *et al.* 1995; g-Wheeler 1995; h-Hoebeker & Wheeler 1996; i-Horn 1996; j-Hough-Goldstein *et al.* 1996; k-LaMana & Miller 1996; l-Kafsur *et al.* 1997; m-Cottrell & Yeargan 1998; n-Barrett & Bailey 2000; o-Rossini *et al.* 2000; p-Bell 2001; q-Hesler *et al.* 2001; r-Kalaskar & Evans 2001; s-McCutcheon & Scott 2001; t-Wise *et al.* 2001; u-Ellis *et al.* 2002; v-Atanassov *et al.* 2003; w-Fauske *et al.* 2003; x-Koch & Hutchison 2003; y-Malinowski 2003; z-Alyokhin & Sewell 2004; aa-Bost & Hale 2004; ab-Butin *et al.* 2004; ac-Rutledge *et al.* 2004; ad-Majka & McCorquodale 2006; ae-J. Acorn, personal communication; af-P. Bolin, pers. comm.; ag-W. Cranshaw, pers. comm.; ah-A. Cunningham, pers. comm.; ai-J. Ellington, pers. comm.; aj-J. Knight, pers. comm.; ak-D. Johnson, pers. comm.; al-T.-X. Liu, pers. comm.; am-J.P. Michaud, pers. comm.; an-M.K. Oliver, pers. comm.; and ao-P. Pellitteri, pers. comm.).

the *succinea* group appears to predominate throughout North America (Chapin & Brou 1991; Tedders & Schaefer 1994; Dreistadt *et al.* 1995; Kidd *et al.* 1995; LaMana & Miller 1996; Hesler *et al.* 2001, 2004; Wise *et al.* 2001) even though initial colonies of this insect maintained by the United States Department of Agriculture contained the *succinea* and melanic groups (Kafsur *et al.* 1997) and both groups were available commercially (e.g., Koch *et al.* 2004b).

McCorquodale (1998) estimated the rate of *H. axyridis* range expansion in eastern North America (Abitata Springs, LA, U.S.A. to Sydney, NS, Canada) to be 442 km per year. This rate reflects both the innate dispersal capability and human mediated movement of the insect. However, Hesler *et al.* (2004) suggest that the rate of range expansion for this semi-arboreal species was considerably lower in South Dakota, U.S.A. perhaps because of the relative paucity of

wooded areas in South Dakota and other parts of the central North America compared with eastern North America. Our discussions with other cooperators support the general observation that the westward spread of *H. axyridis* into the Great Plains of the U.S.A. has been slower than the rate of spread along the eastern U.S.A. To determine whether spread might be faster in areas of North America that were more bioclimatically similar to the native Asian range than in areas that were less similar, we first identified the biomes in each state of the U.S.A. or Canadian province and classified each state/province as similar or dissimilar. A state/province was considered similar when >50% of its area was covered with one or more biomes like those found within the native range of *H. axyridis*. We then compared the year of first detection among states/provinces classified as similar with those classified as dissimilar. In general, first detection

occurred earlier (between 1991 and 1995) in states/provinces with biomes similar to those where *H. axyridis* occurs in Asia than in states/provinces with biomes that are dissimilar (first detected between 1993 and 2000) (Kruskal-Wallis Chi-square = 10.2; df = 1; P = 0.001). However, any estimates of range expansion by *H. axyridis* in North America are likely confounded by the numerous releases of this beetle at multiple locations (McCorquodale 1998) and varying levels of sampling intensity that can affect the date of first detection.

In South America, *H. axyridis* was intentionally introduced into the state of Mendoza, Argentina in the late 1990's (Saini 2004). Shortly thereafter, the *succinea* form of the species was recovered in the state of Buenos Aires in northern Argentina where it was associated with *Monellia caryella* (Fitch) on pecan, *Carya illinoensis* (Wangenh.) K. Koch, in 2001 (Saini 2004), and in the city of Curitiba in southern Brazil feeding on *Tinocallis kahawaluokalani* (Kirkaldy) on crape myrtle, *Lagerstroemia indica* L., and *Cinara atlantica* (Wilson) and *C. pinivora* (Wilson) on pine, *Pinus* spp., in 2002 (Almeida & Silva 2002). Curitiba is located in an area of Brazil with a biome classified as tropical and subtropical moist broadleaf forest, which is also found within the native range of *H. axyridis*. In Argentina, both Mendoza and Buenos Aires are predominantly temperate grasslands, savannahs, and shrublands, a biome different from those within the native range of *H. axyridis*, but the same as that of the adventive range in central North America.

Implications for South America

Questions, developed by the United States National Research Council, specific to various factors affecting each stage of the invasion process (i.e., arrival, establishment, and spread) (Mack *et al.* 2002) can be used to begin evaluating the likelihood of *H. axyridis* invading South America. In this paper, we provide responses to what we considered the most pertinent questions (Table 2).

Arrival. Since *H. axyridis* has already arrived in South America this stage of the invasion process may seem trivial. However, its importance lies in the potential for additional arrival events by this species in uninvaded areas. The

likelihood of continued arrival of *H. axyridis* to South America is high for two reasons. First, continued intentional introduction seem probable, unless regulatory intervention is imposed. This insect has been repeatedly introduced and/or redistributed intentionally for biological control in North America, Europe, Africa, and South America. Although we are unaware of additional documented releases in the South America, it seems likely, given the popularity of this predator, that more releases for biological control purposes were made or could be made. Second, the likelihood of *H. axyridis* being accidentally introduced into South America also seems high, partly due to the large and expanding overall geographic range of this beetle. With the relatively recent entrance and rapid expansion into the global market by countries like Brazil, *H. axyridis* could accidentally arrive in South America through seaports or other points of entry (e.g., Day *et al.* 1994). *H. axyridis* has been detected on cargo crossing the Atlantic (Roy *et al.* 2005). Therefore, as exports to the South America from countries experiencing large populations of *H. axyridis* continue to increase, the potential for *H. axyridis* accidentally entering South America via cargo shipments will also increase.

Establishment. Outside of its expansive native range, *H. axyridis* has proven capable of establishing populations under a broad range of environmental conditions, particularly in North America. Climate plays an important role in determining the geographic range of ectothermic organisms. Therefore, climate matching between native and potential areas is often used to evaluate the risk of establishment for exotic species (e.g., Baker 2002). Climatic similarities between South America and locations in Asia were measured using climate matching software, CLIMEX (v2, Hearne Scientific, Melbourne), and 0.5°-gridded climate data from the Climate Research Unit (Norwich, UK). This software has been used to predict the potential distribution of numerous organisms (e.g., Sutherst *et al.* 1989, Venette & Hutchison 1999, Venette & Ragsdale 2004). Climate data were from 1961-1990. Because the *succinea* form of *H. axyridis*, predominates in South America (Almeida & Silva 2002, Saini 2004) and North America (see above), we focused on this form of *H. axyridis*

Table 2. Questions relevant to the invasion of *H. axyridis* in South America (revised from Mack *et al.* 2002).

Arrival

- Has the arthropod been recently intercepted in South America?
- Does the arthropod have a wide geographic range (proportional to likelihood of transport)?
- Does the arthropod, at some times, have high population densities in its native or current range?

Establishment

- Is there a history of establishment in similar habitats elsewhere outside the native range?
- Is the climate similar between the current geographic range and potential destinations?
- Are potential hosts spatially and temporally available?

Spread

- Does the arthropod have an effective means of dispersal (natural or human-assisted)?
-

for our climate-based predictions. Consequently, we restricted comparisons to the eastern extent of the native range of *H. axyridis*, where the *succinea* form is more abundant. Five locations were identified that spanned the reported latitudinal and longitudinal range of the *succinea* group of *H. axyridis* in Asia: Chita, Russia; Vladivostok, Russia; Kofu, Japan; Fang'ao, China; and Gongshan, China. Climate matching was performed for the entire year based on average monthly maximum temperature, minimum temperature, total precipitation, and precipitation pattern, all weighted equally. Similarity was expressed as a proportion, with 1.00 reflecting a perfect match between two locations. We also used the biome matching method described in Venette and Ragsdale (2004) to identify which biomes in South America might provide suitable habitat based on the distribution of the *succinea* group in Asia.

South America is climatically diverse, and the degree of similarity to the native range of *H. axyridis* varies considerably depending on where comparisons are made. The climate in Asia associated with the northern, native range of the beetle was most similar to the higher elevations in the mountainous terrain of Chile, Bolivia and Peru (Fig. 2A, B). The southerly range of the beetle is most similar to southern Brazil and northern Argentina (Fig. 2C, D, E). None of the locations in Asia closely matched the climate of northern Brazil (Fig. 2). In general, the southern half of South America appears to have a high degree of climatic similarity and might be assumed to be suitable for establishment for *H. axyridis*. Likewise, the comparison of biomes suggests that large areas of South America should provide suitable habitat (Fig. 3), but the most suitable areas are different from those areas highlighted by the climate analysis. In fact, the biome analysis suggests northern Brazil with tropical and subtropical moist broadleaf forests should provide a suitable habitat. How do we reconcile these conflicting predictions?

A poor climatic match does not necessarily mean an area is not suitable for establishment. For example, average high temperatures near Gongshan, China vary between 13-23°C during a year, and average low temperatures vary between 0-15°C. In contrast, average high temperatures near Manaus, in northern Brazil only fluctuate between 30-33°C during a year, and low temperatures vary between 23-24°C. Based on maximum temperature and minimum temperature, similarity indices between the two locations were only 0.16 and 0.11, respectively, out of 1.00. When precipitation is factored, the composite match index climbs to 0.32 out of 1.00. On all measures, the climates of the two locations seem dissimilar. However, *H. axyridis* begins to develop between ~8°C (Soares *et al.* 2003) and ~11°C (LaMana & Miller 1998). Development is most rapid at ~30°C (Schanderl *et al.* 1985). Likewise, females are mated more quickly at 30°C than at cooler temperatures (Stathas *et al.* 2001). Metabolic patterns indicate that temperatures of 35°C begin to stress second instars and adults (Acar *et al.* 2004). Mortality of these stages is high at 40°C, but all other stages survive (Acar *et al.* 2004). Thus, although temperatures in equatorial Brazil are considerably warmer than in Gongshan, China, these temperatures do not seem sufficiently extreme

to preclude establishment. In addition, no evidence suggests *H. axyridis* has obligatory diapause or requires cold exposure to mature reproductively.

Eco-climatic predictions of a species potential distribution can be based on a solid understanding of the known distribution of the species or biological parameters specific to that species (Baker 2002). However, for *H. axyridis*, only coarse descriptions of the native distribution have been provided. Further work is needed to provide a more detailed understanding of this species' native range. In addition, more work is needed to develop parameters describing this species' response to various environmental conditions to allow for more detailed modeling of its population dynamics and potential distribution in South America.

The likelihood of establishment also depends on the availability of prey. Native aphid species are relatively sparse in South America (Dixon 1985; Blackman & Eastop 1994, 2000). However, this fact will likely have little impact on the probability of *H. axyridis* establishment for two reasons. First, exotic aphid species have colonized plants and become pests in South America (Blackman & Eastop 2000). Second, although *H. axyridis* is primarily aphidophagous, it will feed on numerous other insect prey (e.g., Tetranychidae, Psyllidae, Coccoidea, Chrysomelidae, Curculionidae, and Lepidoptera) and plant material (e.g., damaged fruit, pollen, and nectar) (Koch 2003). Therefore, the potential lack of aphid prey in South America may be compensated for by the increased relative numbers of other prey, such as Psyllidae (Blackman & Eastop 1994), Aleocharidae, and Coccoidea (Dixon 1985).

Spread. *H. axyridis* possesses exceptional means of dispersal. As reviewed above, because of its value for biological control, this predator will often be collected and redistributed or mass-reared and released into locations it has not yet invaded. In addition, its tendency to aggregate in concealed locations on prominent objects prior to overwintering may increase the likelihood of *H. axyridis* going undetected and being unintentionally transported via plane, train, truck, or ship to new locations. Furthermore, coccinellids have potential to fly great distances (Hodek & Honik 1996). Based on the relatively slow spread of *H. axyridis* through the temperate grasslands, savannahs, and shrublands of the central U.S.A. (see above), it seems likely that its spread in Argentina and other areas similar to the central U.S.A. may experience slightly slower rates of range expansion than areas with biomes like those found in the Asian native range and eastern and western North America.

Conclusions

Overall, the invasion of *H. axyridis* over broad areas in South America seems likely. Once established, eradication of exotic, invasive species is difficult to impossible. Because of this, efforts should be focused not on elimination of widespread established populations of *H. axyridis*, but on learning how to slow the spread and cope with the presence of this exotic organism. To slow the spread of *H. axyridis*,

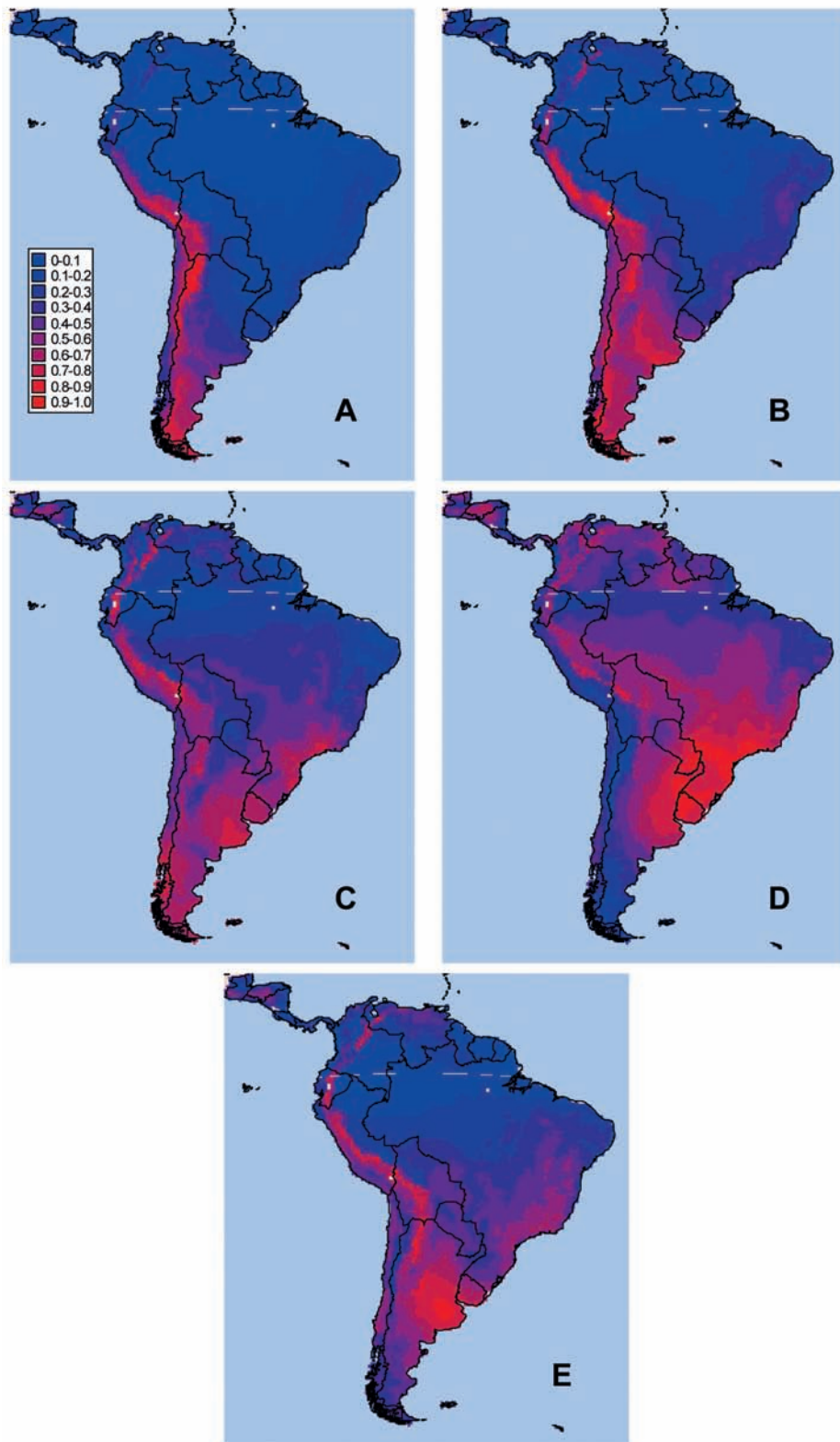


Figure 2. Climatic similarity of South America and southern Central America to Asian cities with *succinea* forms of *H. axyridis*: A) Chita, Russia [52.0°N, 113.6°E]; B) Vladivostok, Russia [43.1°N, 131.9°E]; C) Kofu, Japan [35.7°N, 138.6°E]; D) Fang'ao, China [27.9°N, 120.6°E]; and E) Gongshan, China [25.8°N, 103.2°E].

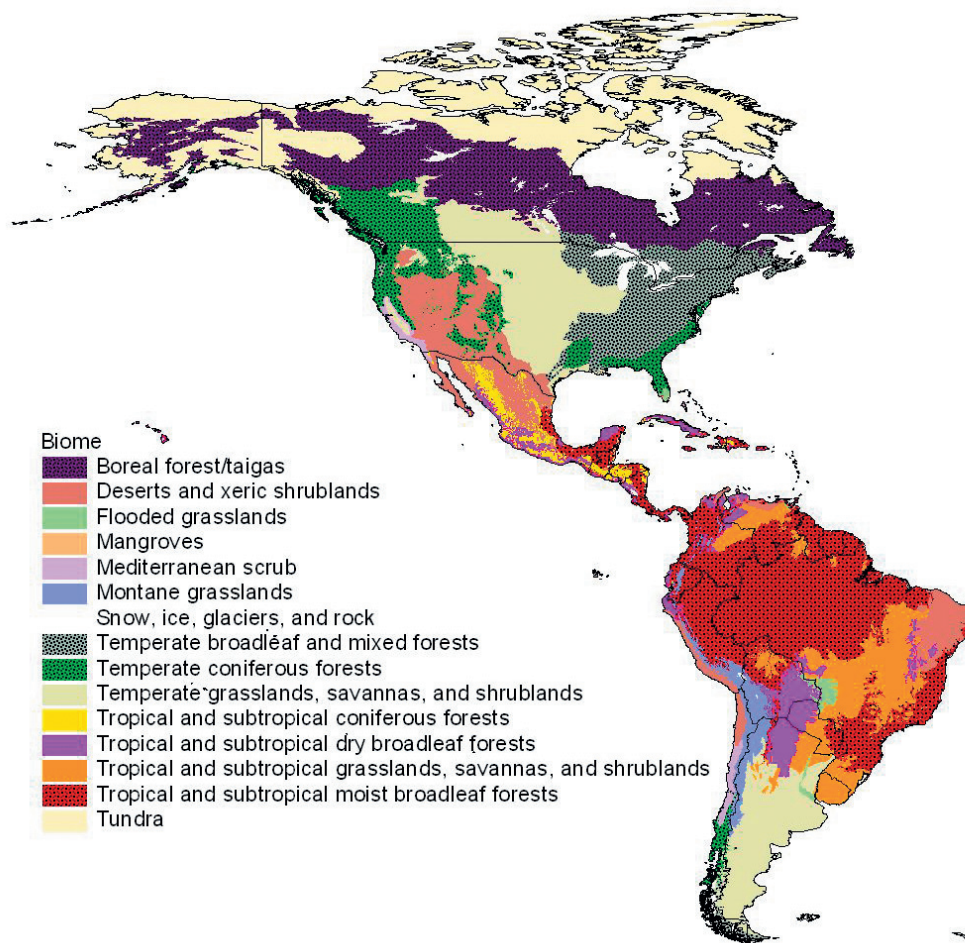


Figure 3. Biomes of the Western Hemisphere as defined by Olson *et al.* (2001). Dots indicate biomes occupied by *H. axyridis* in Asia.

we must explore techniques to minimize the likelihood of further human aided dispersal of *H. axyridis*. This could include increased sampling efforts at points of entry into areas where *H. axyridis* is not yet established. Furthermore, we should consider putting an end to the mass rearing and release of this natural enemy for biological control purposes into areas where it has not yet established. However, because *H. axyridis* will likely become permanently established in many of the areas it has invaded, we must continue to advance our knowledge on how to reap benefits in situations where *H. axyridis* is a potential biological control agent and mitigate its effects in situations where it is a potential pest (e.g., as a household pest, pest of fruit production and threat to non-target organisms). For instance, further work is needed to promote the potential biological control offered by this predator and to begin incorporating it into existing integrated pest management programs (e.g., Musser & Shelton 2003; Musser *et al.* 2004; Galvan *et al.* 2005a, 2005b, *in press*; Koch *et al.* *in press a*). In contrast, continued work is needed to aid in management of this pest in household (e.g., Nalepa *et al.* 2004, 2005) and fruit

production (e.g., Koch *et al.* 2004a, Pickering *et al.* 2004, Galvan *et al.* 2006) situations. In addition, caution should be taken if trying to promote populations of this predator for biological control, so that impacts on non-target organisms can be minimized. Furthermore, if action is taken to suppress populations of *H. axyridis*, what non-target impacts could result from such intervention?

The multiple invasions of *H. axyridis* across the globe provide us with a unique opportunity to examine invasion biology. Essentially, we have two replications of the invasion process on continental scales in North America and Europe, and a third replication of this invasion process beginning in South America. As a scientific community we could benefit greatly from more coordinated research efforts among countries within continents and among continents.

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