

# Benthic Invertebrate Fauna, Small Streams

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## Introduction

Small streams (first- through third-order streams) make up >98% of the total number of stream segments and >86% of stream length in many drainage networks. Small streams occur over a wide array of climates, geology, and biomes, which influence temperature, hydrologic regimes, water chemistry, light, substrate, stream permanence, a basin's terrestrial plant cover, and food base of a given stream. Small streams are generally most abundant in the upper reaches of a basin, but they can also be found throughout the basin and may enter directly into larger rivers. They have maximum interface with the terrestrial environment, and in most temperate and tropical climates they may receive large inputs of terrestrial, or allochthonous, organic matter (e.g., leaves, wood) from the surrounding plant communities. In locations with open canopies such as grasslands and deserts, autochthonous or primary production in the form of algae, or higher aquatic plants, may serve as the main food base. Hence, headwater streams display a diverse fauna, which is often adapted to physical, chemical, and biotic conditions of the region.

## Diversity of Benthic Invertebrates in Small Streams

The benthic invertebrate fauna of small streams is composed primarily of aquatic insects, crustaceans, mollusks, and various other invertebrate taxa. The insect fauna consists primarily of Odonata (dragonflies and damselflies), Ephemeroptera (mayflies), Plecoptera (stoneflies), Megaloptera (alderflies and dobsonflies), Coleoptera (beetles), Trichoptera (caddisflies), occasional Lepidoptera (moths), and Diptera (true flies). Crustaceans (including amphipods, isopods, and crayfish) can also be found in small streams as well as microcrustaceans such as cladocerans, ostracods, and copepods. Other common invertebrates found in small streams include nematodes, oligochaetes, turbellarians, and mollusks such as snails, limpets, and finger-nail clams. Total invertebrate diversity in small streams can be quite high. The Breitenbach, a first-order stream in Germany, contains at least 1004 described invertebrate taxa. At least 293 invertebrate taxa have been found in headwater streams in the southern Appalachian mountains of the United States. Over 182 known

invertebrate taxa have been recorded in a mountain stream on Bougainville Island, Papua New Guinea. Incredibly, there are many headwater invertebrate species that remain undescribed in both isolated and populated regions of the world.

With the great diversity of foods available for consumption by invertebrates (i.e., deposited and retained on substrates, or suspended in the water column), it is not surprising that invertebrates have evolved diverse morphobehavioral mechanisms for exploiting food resources. Their diverse feeding behaviors have been lumped into a broad functional classification scheme, which is based on mechanisms used by invertebrates to acquire foods. These functional groups are as follows: scrapers, animals adapted to graze or scrape materials (periphyton, or attached algae, fine particulate organic matter, and its associated microbiota) from mineral and organic substrates; shredders, organisms that consume large pieces of decomposing vascular plant tissue such as leaf detritus (>1 mm diameter) along with its associated microflora and fauna, or feed directly on living vascular hydrophytes, or gouge decomposing wood; gatherers, animals that feed primarily on deposited fine particulate organic matter (FPOM  $\leq$  1 mm diameter); filterers, animals that have specialized anatomical structures (e.g., setae, mouth brushes, or fans) or silk and silk-like secretions that act as sieves to remove particulate matter from suspension; and predators, those organisms that feed primarily on animal tissue by either engulfing their prey, or piercing prey and sucking body contents.

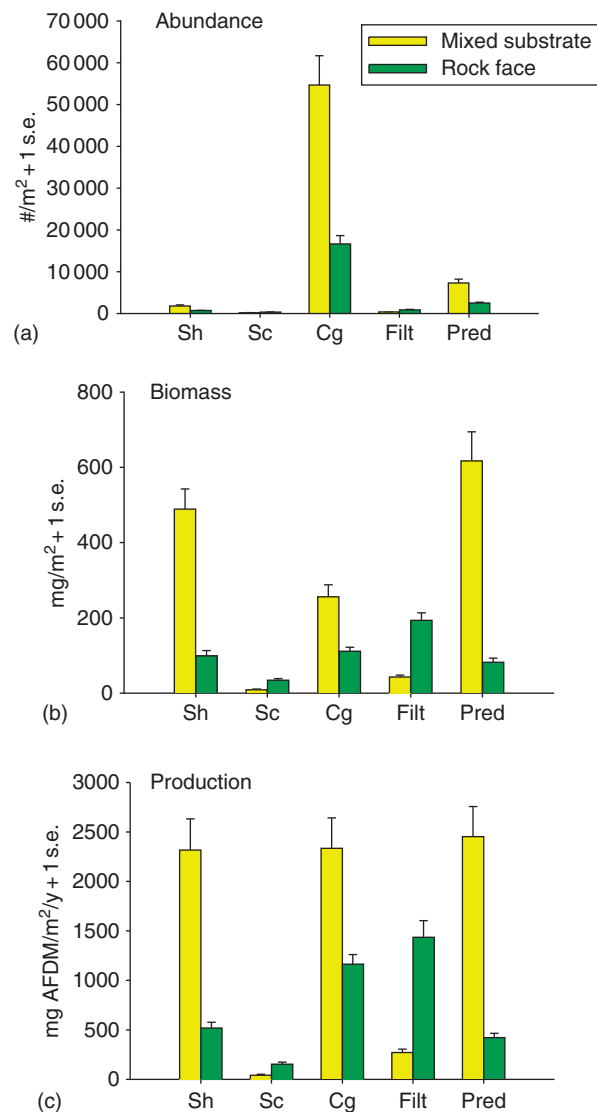
Functional feeding groups refer primarily to modes of feeding and not type of food per se. For example, many filter-feeding insects of high gradient streams are primarily carnivores, whereas scrapers consume quantities of what must be characterized as epilithon, a matrix of polysaccharide exudates, detritus, microflora, and microfauna associated with stone surfaces, and not solely attached algae. Shredders may select those leaves that have been 'microbially conditioned' by colonizing fungi and bacteria. Shredders also ingest attached algal cells, protozoans, and various other components of the fauna during feeding. Some 'shredders' have been shown to grow by harvesting primarily the epixylic biofilm, the matrix of exudates, detritus, microflora, and microfauna found on wood surfaces. Although it appears valid to separate benthic invertebrates according to these mechanisms used to obtain foods, many questions remain concerning

the ultimate sources of protein, carbohydrates, fats, and assimilated energy to each of these functional groups.

### Quantitative Measurements of Benthic Invertebrates in Headwater Streams

Invertebrates are often enumerated by abundances or average numbers per unit area of stream bottom. Other measures include average biomass, or weight, per unit area of stream, or more rarely, secondary production per unit area of stream bottom. Each of these will provide a different picture of the invertebrate community. For example, [Figure 1\(a\)](#) shows abundances per unit area of moss-covered bedrock outcrop and mixed substrates in three headwater streams ( $n=20$  total stream years) at the Coweeta Hydrologic Laboratory in western North Carolina, USA. Note that abundances are dominated by members of the collector-gatherer (Cg), functional group. In contrast, three groups, predators, shredders, and collectors, represent the majority of the biomass in these small streams ([Figure 1\(b\)](#)). Secondary production, which represents the living organic matter or biomass produced by each functional group over a year regardless of its fate, i.e., losses to predation, or other sources of mortality, is fairly evenly distributed between the predator, collector, and shredder functional groups ([Figure 1\(c\)](#)). The integration of production, feeding habits, and bioenergetic data can yield a much better understanding of the role of animal populations in ecosystem function than either abundance or biomass.

Distributional patterns for functional feeding group abundance, biomass, and production in small streams may differ among substrate types ([Figure 1\(a\)–1\(c\)](#)). Collectors and predators dominate abundances on both substrates. For biomass, predators > shredders > collectors dominate the mixed substrates, whereas filterers > collectors > predators contribute most to biomass on bedrock outcrop substrates. Most production is attributed to predators > shredders > collectors in mixed substrates compared to filterers > collectors > predators on bedrock outcrop substrates. Thus, distinct differences exist in functional feeding group production among different substrates within a stream, which correspond to different available food resources. Filterer production predominates in the bedrock habitats with high current velocities that transport FPOM. Collector production is also enhanced by FPOM trapped in the moss on the bedrock outcrops. Conversely, predator, shredder, and collector production are similar in the retentive mixed substrate habitats, which also have the greatest biomass, abundances, and organic matter retention. Scraper abundance, biomass, and production are low for all



**Figure 1** Total abundance, biomass, and production of invertebrates by functional feeding group for mixed substrate and rockface habitats in headwater streams at Coweeta Hydrologic Laboratory in western North Carolina, USA. These data represent 20 total years of abundance, biomass, and production estimates.

habitats as these small streams are heavily shaded year round by dense riparian rhododendron. These data emphasize the influence of local geomorphic processes and riparian linkages on invertebrate productivity in forested headwater streams.

### Comparison of Secondary Productivity Measurements from Small Streams

Secondary productivity measures for benthic invertebrates from small streams from various temperate areas around the world are given in [Table 1](#). Total annual productivity is quite variable ranging from

**Table 1** Estimates of secondary production ( $\text{g m}^{-2} \text{ year}^{-1}$ ) for various functional feeding groups, or primary and secondary consumers, for small streams from various regions of the world

| Country | Biome              | Stream                      | Scrapers | Shredders | Collectors | Filterers | Predators | Total production | Source |
|---------|--------------------|-----------------------------|----------|-----------|------------|-----------|-----------|------------------|--------|
| USA, ID | Cool desert        | Douglas Creek*              | 2.65     | 0.64      | 15.28      | 4.20      | 0.45      | 23.22            | 2      |
| USA, ID | Cool desert        | Snively Springs*            | 0.00     | 1.32      | 9.33       | 3.18      | 0.33      | 14.15            | 2      |
| USA, ID | Cool desert        | Rattlesnake Springs*        | 0.00     | 0.17      | 3.62       | 11.80     | 0.77      | 16.36            | 2      |
| Denmark | Deciduous          | Rold Kilde                  | 0.14     | 5.93      | 2.58       | 0.01      | 0.88      | 9.54             | 6      |
| USA, NC | Eastern deciduous  | Coweeta Catchment 53        | 0.09     | 3.23      | 5.78       | 0.56      | 4.10      | 13.77            | 7      |
| USA, NC | Eastern deciduous  | Coweeta Catchment 54        | 0.22     | 3.53      | 3.77       | 0.72      | 3.17      | 11.41            | 7      |
| USA, NC | Eastern deciduous  | Coweeta (1985) Catchment 55 | 0.17     | 2.51      | 2.86       | 0.73      | 2.52      | 8.79             | 7      |
| USA, NC | Eastern deciduous  | Coweeta (1986) Catchment 55 | 0.50     | 2.75      | 3.41       | 0.54      | 3.37      | 10.57            | 7      |
| USA, NC | Eastern deciduous  | Upper Ball Ck               | 0.68     | 1.67      | 2.95       | 0.53      | 1.68      | 7.51             | 5      |
| USA, NC | Eastern deciduous  | Bear Pen Ck                 | 0.68     | 2.02      | 4.45       | 2.32      | 1.66      | 11.12            | 12     |
| USA, VA | Eastern deciduous  | Buzzard's Branch            | 0.11     | 2.84      | 6.21       | 0.98      | 3.78      | 13.92            | 10     |
| USA, VA | Eastern deciduous  | Collier's Ck                | 0.11     | 1.69      | 0.95       | 1.28      | 1.58      | 5.60             | 10     |
| USA, NH | Eastern deciduous  | Bear Brook                  | 0.74     | 1.45      | 0.61       | 0.43      | 0.94      | 4.17             | 3      |
| USA, ME | Eastern deciduous  | Goosefare Brook             | 0.16     | 11.78     | 5.28       | 3.96      | 5.20      | 27.35            | 13     |
| USA, ME | Eastern deciduous  | West Bear Brook             | <0.00    | 0.86      | 0.21       | 0.22      | 0.37      | 1.66             | 1      |
| USA, ME | Eastern deciduous  | East Bear Brook             | <0.00    | 0.80      | 0.20       | 0.27      | 0.41      | 1.68             | 1      |
| USA, KS | Tall grass prairie | Kings Creek                 | 3.80     | 4.50      | 6.00       | 1.70      | 3.60      | 19.60            | 11     |
| Germany | Deciduous forest   | Steina (1986)               | 4.63     | 5.10      | 4.53       | 2.43      | 2.33      | 19.02            | 8      |
| Germany | Deciduous forest   | Steina (1987)               | 7.96     | 2.38      | 3.84       | 3.72      | 2.93      | 20.83            | 8      |

Continued

**Table 1** Continued

| Country                                                                                                            | Biome             | Stream         | Scrapers                                       | Shredders | Collectors | Filterers                                         | Predators | Total production | Source |
|--------------------------------------------------------------------------------------------------------------------|-------------------|----------------|------------------------------------------------|-----------|------------|---------------------------------------------------|-----------|------------------|--------|
| The following studies calculated production based on primary and secondary consumers rather than functional groups |                   |                |                                                |           |            |                                                   |           |                  |        |
|                                                                                                                    |                   |                | Total primary consumer production <sup>a</sup> |           |            | Total secondary consumers production <sup>b</sup> |           | Total production |        |
| USA, MA                                                                                                            | Eastern deciduous | Factory Brook* | 4.00                                           |           |            | 0.56                                              |           | 4.56             | 9      |
| New Zealand                                                                                                        | Tussock Grass     | Sutton Stream  | 13.35                                          |           |            | 2.54                                              |           | 15.89            | 4      |

Note that data for those streams marked with an asterisk were in dry mass (DM), whereas others were in ash-free dry mass (AFDM) DM values are ~10–20% greater than AFDM.

<sup>a</sup>Primary consumers includes scraper, shredder, collector, and filterer functional feeding groups.

<sup>b</sup>Secondary consumers = predators.

#### Sources

1. Chadwick MA and Huryn AD (2005) Response of stream macroinvertebrate production to atmospheric nitrogen deposition and channel drying. *Limnology and Oceanography* 50: 228–236.
2. Gaines WL, Cushing CE, and Smith SD (1992) Secondary production estimates of benthic insects in three cold desert streams. *Great Basin Naturalist* 52: 11–24.
3. Hall RO, Likens GE, and Malcom JM (2001) Trophic basis of invertebrate production in 2 streams at the Hubbard Brook Experimental Forest. *Journal of the North American Benthological Society* 20: 432–447.
4. Huryn AD (1996) An appraisal of the Allen paradox in a New Zealand trout stream. *Limnology and Oceanography* 41: 243–252.
5. Huryn AD and Wallace JB (1987) Local geomorphology as a determinant of macrofaunal production in a mountain stream. *Ecology* 68: 1932–1942.
6. Iversen TM (1988) Secondary production and trophic relationships in a spring invertebrate community. *Limnology and Oceanography* 33: 582–592.
7. Lugthart JG and Wallace JB (1992) Effects of disturbance on benthic functional structure and production in mountain streams. *Journal of the North American Benthological Society* 11: 138–164.
8. Meyer EI and Poepperl R (2003) Secondary production of invertebrates in a Central European mountain stream (Steina, Black Forest, Germany) *Archiv für Hydrobiologie*, 158: 25–42.
9. Neves RJ (1979) Secondary production of epilithic fauna in a woodland stream. *American Midland Naturalist* 102: 209–224.
10. Smock LA, Gladden JE, Riekenberg JL, Smith LC, and Black CR (1992) Lotic macroinvertebrate production in three dimensions: Channel surface, hyporheic, and floodplain environments. *Ecology* 73: 876–886.
11. Stagliano DM and Whiles MR (2001) Macroinvertebrate production and trophic structure in a tallgrass prairie headwater stream. *Journal of the North American Benthological Society* 21: 97–113.
12. Wohl DL, Wallace JB, and Meyer JL (1995) Benthic macroinvertebrate community structure, function and production with respect to habitat type, reach and drainage basin in the southern Appalachians (USA) *Freshwater Biology* 34: 447–464.
13. Woodcock TS and Huryn AD (2007) The response of macroinvertebrate production to a pollution gradient in a headwater stream. *Freshwater Biology* 52: 177–196.

$<2$  to  $>27 \text{ g m}^{-2} \text{ year}^{-1}$  for streams in a variety of landscapes (Figures 2–10). The following is an indication of the average secondary production (percentage of total) in invertebrates: collectors (31.2%, range 11.9–65.9%) > shredders (avg. = 27.0%, range



**Figure 2** A steep-gradient headwater stream in the Appalachians of western NC. Note the dense rhododendron understory, which shades the stream. (Photo by S. L. Eggert.)

1–62.2%) > predators (avg. = 19.0%, range 1.9–31.9%) > filterers (avg. = 15.1%, range 0.1–72.1%) > scrapers (avg. = 7.4%, range = 0.0–38.2%). Scrapper production as a percent of total production was greatest in the Steina, Germany (deciduous forest), followed by Kings Creek, a tall grass prairie stream in Kansas, and Bear Brook, NH (deciduous forest). The differences in scraper production in small streams in the eastern deciduous forest are striking. In Bear Brook NH, scraper production comprised >17% of total production, compared with 0.7–9% for the Coweeta streams in NC, which are heavily shaded by dense riparian rhododendron. With the exception of three cold desert streams in southeastern Washington, USA, shredder production was always greater than 10% of the total production. Percent collector–filterer production was lowest in a Danish Spring and highest in a cool desert stream. These data demonstrate that invertebrate production in small streams can be quite variable among various temperate regions. Given the usefulness of this integrative measure for comparing small stream functioning in natural and disturbed environments, additional efforts at quantifying total secondary production in small streams are badly needed.

### Factors that Influence Invertebrates in Small Streams

Small stream invertebrates are influenced by physical, chemical, and biological factors (Table 2). Physical factors include climate, (e.g., temperature and



**Figure 3** Headwater stream draining catchment 53 at the Coweeta Hydrologic Laboratory in western North Carolina during autumn. Note large amounts of leaves and woody debris in the stream channel. (Photo by S. L. Eggert.)



**Figure 4** Headwater stream draining watershed 6 in the Hubbard Brook Experimental Forest in central New Hampshire, USA. (Photo by R. O. Hall.)



**Figure 5** Intermittent stream in the Huron Mountains of Michigan's upper peninsula, USA. (Photo by S. L. Eggert.)

precipitation), hydrology, and geology. Hydrology and the frequency of flooding or drying can also influence community structure and productivity. For example, the lowest annual productivity shown in [Table 1](#) occurs in two intermittent streams, East Bear Brook and West Bear Brook in Maine, USA. Geology influences both stream substratum and chemistry. Substratum and the proportions of eroding and depositional substratum within a given stream can have an important effect on invertebrate functional distribution and production. The faster flowing erosional reaches in small streams are often

dominated by filter-feeders and scrapers, whereas depositional reaches with greater amounts of retained organic matter are often dominated by shredders and collectors. Stream chemistries can be strikingly different from those in nearby streams if they have different underlying geologies. For example, in the southern Appalachians, streams draining limestone regions such as the ridge and valley province have higher nutrients, conductivity, pH, and primary production than those draining the crystalline Appalachians. Geology and climate also influence the vegetation of catchments, including the abundance and type of



**Figure 6** Bison grazing in the riparian area of the headwaters of King's Creek, at the Konza Prairie LTER in Kansas, USA. (Photo by Walter Dodd.)



**Figure 8** Rattlesnake Creek, a cold desert stream in eastern Washington, USA. (Photo by C. E. Cushing.)



**Figure 7** Douglas Creek, a cold desert stream in eastern Washington, USA. (Photo by C. E. Cushing.)

riparian vegetation. Streams that are open, and which receive large solar inputs compared with those draining dense forested catchments, may have a very different food base (autochthonous) compared with those receiving primarily allochthonous inputs from forested catchments. Depending on the food base, small streams may display large differences in functional group abundance, biomass, and production.



**Figure 9** A spring-fed tundra stream in Alaska, USA. (Photo by A. D. Huryn.)



**Figure 10** A stream draining tall tussock grass on South Island, New Zealand. (Photo by A. D. Huryn.)

### Ecological Roles of Invertebrates in Small Streams

Feeding activities of invertebrates in small streams link headwater streams to larger rivers downstream by altering resource quantity, size, and shape (Table 2). For example, the shredding of leaf detritus and woody debris by shredders in headwater streams increases the rate of coarse organic matter breakdown to fine organic matter, which is transported by the current to downstream reaches. Scrapers, through their feeding activities and dislodging of epilithon, can enhance the movement of downstream organic particles. Heavy grazing by scrapers results in periphyton mats with adnate, or closely attached forms of diatoms that are less susceptible to scouring during disturbances such as large storms and also promotes nutrient turnover in periphyton communities. Thus, both shredding and grazing activities may result in a consistent, prolonged release of materials to downstream reaches, in contrast to large storms that induce pulsed massive export over short time intervals. The role of gatherers in FPOM transport has been implicated in an Idaho stream, which exhibited continuous deposition and resuspension as particles moved downstream. In montane Puerto Rican streams, feeding activities of atyid shrimp reduce organic matter accrual on benthic substrates. Other invertebrate gatherers such as *Ptychoptera*

(Diptera: Ptychopteridae) and sericostomatid (Trichoptera) larvae may transfer fine organic matter buried in depositional areas to substratum surfaces as feces.

Filter-feeding stream invertebrates enhance retention of organic matter and nutrients. Certain invertebrates can transport superficial organic matter to deeper sediments, which reduces downstream transport. However, direct removal of transported material by filter-feeders has received the bulk of attention and has been shown to have variable effects on retention depending on size of stream, abundance of filterers, and taxon-specific differences in feeding, (e.g., feeding on extremely fine organic particles or drifting invertebrates). Studies using radioactive tracers in Alaskan streams have also suggested that particles generated by invertebrate scrapers such as baetid mayflies and chironomid larvae were instrumental in supplying fine particles to downstream black flies. Microfilterers such as the caddisfly family Philopotamidae, and black flies (Diptera: Simuliidae), and bivalve mollusks increase particle sizes by ingesting minute particles and egesting compacted fecal particles larger than those originally consumed. Such microfilterers perform two very important functions in streams. First, they remove FPOM from suspension (which would otherwise pass unused through the stream segment) and second, they defecate larger particles, which are available to deposit-feeding detritivores.

Predators can play numerous roles at scales ranging from individuals to ecosystems and invertebrate predators in small streams are no exception. Predators can influence export and retention of energy and nutrients through their effects on the standing stocks of other functional groups. Other mechanisms include decreasing rates of nutrient cycling by immobilizing nutrients in long-lived predator taxa versus short-lived prey. Besides direct consumption, foraging by invertebrate predators can enhance invertebrate drift and suspended FPOM, which also increases export of nutrients. Invertebrate predators can enhance retention of organic matter by retarding breakdown rates of leaf litter as well as subsequent generation of FPOM. Predaceous stoneflies and caddisflies can significantly decrease the rate of leaf litter processing by reducing shredder populations in leaf packs. Invertebrate predators can also increase the rate of downstream movement of organisms and sediment. Many stream invertebrates exhibit different responses to fish and invertebrate predators, and the local impact of invertebrate predation on benthic prey may exceed the impact of fish predators. In the presence of fish, invertebrate prey often reduces movement and seeks refuge in the substrate. In contrast, invertebrate

**Table 2** Factors influencing benthic macroinvertebrates and effects of invertebrates on stream processes in headwater streams

| <i>Factor</i>       | <i>Effect on benthic macroinvertebrates and stream processes</i>                                                                                                                                                                                                                                                                                                                                                                                                 | <i>Source</i>             |
|---------------------|------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|---------------------------|
| Physical            |                                                                                                                                                                                                                                                                                                                                                                                                                                                                  |                           |
| Water temperature   | Lower water temperature resulted in higher total taxa richness and Ephemeroptera, Plecoptera and Trichoptera richness; snail production regulated by thermal regime; water temperature positively related to growth rates of invertebrates                                                                                                                                                                                                                       | 5, 20, 21, 47             |
| Hydrology           | Functional feeding group production determined in part by hydrology; current velocity affected invertebrate movement and drift; channel drying altered organic matter standing crops, and invertebrate abundance and production; invertebrate abundance and diversity declined as a result of severe flow diversions                                                                                                                                             | 4, 19, 22, 35, 41         |
| Geology             | Higher alkalinity resulted in greater benthic invertebrate abundance, biomass, drift biomass, and organic matter standing crop; higher snail biomass and caddisfly production                                                                                                                                                                                                                                                                                    | 20, 23, 24, 27            |
| Substrate           | Invertebrate taxa show distinct substrate preferences; functional feeding group production varied with factors associated with stream geomorphology; invertebrate diversity and abundance increases with substrate stability and presence of detritus                                                                                                                                                                                                            | 9, 22, 30                 |
| Chemical            |                                                                                                                                                                                                                                                                                                                                                                                                                                                                  |                           |
| pH                  | Low pH resulted in decreased taxa richness, loss of sensitive taxa (Mollusca, Crustaceans, Ephemeroptera); increased drift immediately after acidification; decreased emergence; long-term decreased abundance and drift; reduced Ephemeroptera growth; reduced leaf breakdown rates; increased detritus standing crop                                                                                                                                           | 2, 11, 13, 16, 17, 39, 45 |
| Conductivity        | Loss or reduction of Ephemeroptera at conductivities >400 $\mu\text{S}/\text{cm}$ ; reduced Ephemeroptera, Plecoptera, and Trichoptera diversity with increased conductivity; replacement of sensitive Ephemeropteran taxa with tolerant Dipteran taxa with increasing conductivity                                                                                                                                                                              | 14, 32, 38                |
| Nutrients           | Increase in abundance, biomass and production of invertebrates; increased growth rates for short-lived invertebrates; increased growth and abundance of limnephilid caddisflies in the presence of salmon carcasses                                                                                                                                                                                                                                              | 7, 8, 51                  |
| Biological          |                                                                                                                                                                                                                                                                                                                                                                                                                                                                  |                           |
| Riparian vegetation | Invertebrate distribution, abundance, biomass, production, diversity, and growth rates, as well as leaf breakdown rates, were strongly related to riparian vegetation composition                                                                                                                                                                                                                                                                                | 1, 15, 29, 34, 42, 46, 52 |
| Competition         | Evidence of interspecific competition: between snail and caddisfly grazers, between caddisfly and mayfly grazers, between net-veined midges and blackflies, and between caddisflies and filterers; intra- and interspecific competition between snails; intraspecific competition for Trichopteran and Ephemeropteran shredders. Competition resulted in varied responses with regard to survivorship, growth rates, colonization of habitat, and feeding rates. | 3, 6, 12, 18, 25, 26, 33  |
| Predation           | Predation on shredders resulted in reduced leaf breakdown and FPOM generation; predation on scrapers resulted in increase periphyton biomass; long-lived predators retain nitrogen; predation causes downstream movement of inorganic material; predators cause prey to drift or seek refuge in the substrate                                                                                                                                                    | 26, 31, 37, 43, 44, 54    |

Continued

**Table 2** Continued

| <i>Factor</i>                   | <i>Effect on benthic macroinvertebrates and stream processes</i>                                                                                                                          | <i>Source</i>      |
|---------------------------------|-------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|--------------------|
| Altered resource size and shape | Grazing caddisflies increase algal turnover; shredders convert large organic matter to FPOM; microfilterers increase particle size                                                        | 28, 36, 55         |
| Material transport              | Gatherers feeding on and egesting FPOM enhance downstream transport; invertebrates transfer FPOM to surface as feces; shrimps reduce organic matter accrual on substrates through feeding | 10, 40, 48, 53, 55 |
| Material retention              | Incorporation of labeled nitrogen by scrapers and filters downstream; microfilterers retain FPOM; filterers reduce downstream transport of particulate organic matter                     | 37, 49, 50         |

## Sources

1. Aguiar FC, Ferreira MT, and Pinto P (2002) Relative influence of environmental variables on macroinvertebrate assemblages from an Iberian basin. *Journal of the North American Benthological Society* 21: 43–53.
2. Bernard DP, Neill WE, and Rowe L (1990) Impact of mild experimental acidification on short term invertebrate drift in a sensitive British Columbia stream. *Hydrobiologia* 203: 63–72.
3. Boyero L and Pearson RG (2006) Intraspecific interference in a tropical stream shredder guild. *Marine and Freshwater Research* 57: 201–206.
4. Chadwick MA and Hurny AD (2007) Role of habitat in determining macroinvertebrate production in an intermittent-stream system. *Freshwater Biology* 52: 240–251.
5. Collier KJ (1995) Environmental factors affecting the taxonomic composition of aquatic macroinvertebrate communities in lowland waterways of Northland, New Zealand. *New Zealand Journal of Marine and Freshwater Research* 29: 453–465.
6. Cross WF and Benke AC (2002) Intra- and interspecific competition among coexisting lotic snails. *Oikos* 96: 251–264.
7. Cross WF, Johnson BR, Wallace JB, and Rosemond AD (2005) Contrasting response of detritivores to long-term nutrient enrichment. *Limnology and Oceanography* 50: 1730–1739.
8. Cross WF, Wallace JB, Rosemond AD, and Eggert SL (2006) Whole-system nutrient enrichment increases secondary production in a detritus-based ecosystem. *Ecology* 87: 1556–1565.
9. Cummins KW and Lauff GH (1969) The influence of substrate particle size on the microdistribution of stream macrobenthos. *Hydrobiologia* 34: 145–181.
10. Cushing CE, Minshall GW, and Newbold JD (1993) Transport dynamics of fine particulate organic matter in two Idaho streams. *Limnology and Oceanography* 38: 1101–1115.
11. Dangles O, Gessner MO, Guerold F, and Chauvet E (2004) Impacts of stream acidification on litter breakdown: Implications for assessing ecosystem functioning. *Journal of Applied Ecology* 41: 365–378.
12. Dudley TL, D'Antonio CM, and Cooper SD (1990) Mechanisms and consequences of interspecific competition between two stream insects. *Journal of Animal Ecology* 59: 849–866.
13. Fiance SB (1978) Effects of pH on the biology and distribution of *Ephemerella funeralis* (Ephemeroptera) *Oikos* 31: 332–339.
14. Garcia-Criado F, Tomé A, Vega FJ, and Antolín C (1999) Performance of some diversity and biotic indices in rivers affected by coal mining in northwestern Spain. *Hydrobiologia* 394: 209–217.
15. Grubbs SA and Cummins KW (1996) Linkages between riparian forest composition and shredder voltinism. *Archiv für Hydrobiologie* 137: 39–58.
16. Guerold F, Boudot JP, Jacquemin G, Vein D, Merlet D, and Rouiller J (2000) Macroinvertebrate community loss as a result of headwater stream acidification in the Vosges Mountains (N-E France). *Biodiversity and Conservation* 9: 767–783.
17. Hall RJ, Likens GE, Fiance SB, and Hendrey GR (1980) Experimental acidification of a stream in the Hubbard Brook experimental forest, New Hampshire. *Ecology* 61: 976–989.
18. Hill WR, Weber SC, and Stewart AJ (1992) Food limitation of two grazers: Quantity, quality, and size-specificity. *Journal of the North American Benthological Society* 11: 420–432.
19. Hoffman AL, Olden JD, Monroe JB, et al. (2006) Current velocity and habitat patchiness shape stream herbivore movement. *Oikos* 115: 358–368.
20. Hurny AD, Benke AC, and Ward GM (1995) Direct and indirect effects on geology on the distribution, biomass, and production of the freshwater snail *Elimia*. *Journal of the North American Benthological Society* 14: 519–534.
21. Hurny AD and Wallace JB (1985) Life history and production of *Goerita semata* Ross (Trichoptera: Limnephilidae) in the southern Appalachian Mountains. *Canadian Journal of Zoology* 63: 2604–2611.
22. Hurny AD and Wallace JB (1987) Local geomorphology as a determinant of macrofaunal production in a mountain stream. *Ecology* 68: 1932–1942.
23. Jin HS and Ward GM (2007) Life history and secondary production of *Glossosoma nigror* Banks (Trichoptera: Glossosomatidae) in two Alabama streams with different geology. *Hydrobiologia* 575: 245–258.
24. Koetsier P, Minshall GW, and Robinson CT (1996) Benthos and macroinvertebrate drift in six streams differing in alkalinity. *Hydrobiologia* 317: 41–49.

25. Kohler SL (1992) Competition and the structure of a benthic stream community. *Ecological Monographs* 62: 165–188.
26. Kohler SL and Wiley MJ (1997) Pathogen outbreaks reveal large-scale effects of competition in stream communities. *Ecology* 78: 2164–2176.
27. Krueger CC and Waters TF (1983) Annual production of macroinvertebrates in three streams of different water quality. *Ecology* 64: 840–850.
28. Lamberti GA and Resh VH (1983) Stream periphyton and insect herbivores: An experimental study of grazing by a caddisfly population. *Ecology* 64: 1124–1135.
29. Lecerf A, Dobson M, Dang CK, and Chauvet E (2005) Riparian plant species loss alters trophic dynamics in detritus-based stream ecosystems. *Oecologia* 146: 432–442.
30. Mackay RJ and Kalff J (1969) Seasonal variation in standing crop and species diversity of insect communities in a small Quebec stream. *Ecology* 50: 101–109.
31. Malmqvist B (1993) Interactions in stream leaf packs: Effects of a stonefly predator on detritivores and organic matter processing. *Oikos* 66: 454–462.
32. Marqués MJ, Martínez-Conde E, and Rovira JV (2003) Effects of zinc and lead mining on the benthic macroinvertebrates of a fluvial ecosystem. *Water, Air, and Soil Pollution* 148: 363–388.
33. McAuliffe JR (1984) Resource depression by a stream herbivore: Effects on distributions and abundances of other grazers. *Oikos* 42: 327–333.
34. Milner AM and Gloyne-Phillips IT (2005) The role of riparian vegetation and woody debris in the development of macroinvertebrate assemblages in streams. *River Research and Applications* 21: 403–420.
35. Minshall GW and Winger PV (1968) The effect of reduction in stream flow on invertebrate drift. *Ecology* 49: 580–582.
36. Petersen RC and Cummins KW (1974) Leaf processing in a woodland stream. *Freshwater Biology* 4: 343–368.
37. Peterson BJ, Bahr M, and Kling GW (1997) A tracer investigation of nitrogen cycling in a pristine tundra river. *Canadian Journal of Fisheries & Aquatic Sciences* 54: 2361–2367.
38. Pond GJ and McMurray SE (2002) A macroinvertebrate bioassessment index for headwater streams in the eastern coalfield region, Kentucky. Kentucky Department for Environmental Protection, Division of Water, Frankfort, KY.
39. Pretty JL, Gibberson DJ, and Dobson M (2005) Resource dynamics and detritivore production in an acid stream. *Freshwater Biology* 50: 578–591.
40. Pringle CM, Blake GA, Covich AP, Buzby KM, and Finley A (1993) Effects of omnivorous shrimp in a montane tropical stream: Sediment removal, disturbance of sessile invertebrates and enhancement of understory algal biomass. *Oecologia* 93: 1–11.
41. Rader RB and Belish TA (1999) Influence of mild to severe flow alterations on invertebrates in three mountain streams. *Regulated Rivers: Research & Management* 15: 353–363.
42. Ross HH (1963) Stream communities and terrestrial biomes. *Archiv für Hydrobiologie* 59: 235–242.
43. Sih A and Wooster D (1994) Prey behavior, prey dispersal, and predator impacts on stream prey. *Ecology* 75: 1199–1207.
44. Statzner B, Fuchs U, and Higler LWG. (1996) Sand erosion by mobile predaceous stream insects: Implications for ecology and hydrology. *Water Resources Research* 32: 2279–2287.
45. Sutcliffe DW and Carrick TR (1973) Studies on mountain streams in the English Lake District. I. pH, calcium and the distribution of invertebrates in the River Duddon. *Freshwater Biology* 3: 437–462.
46. Synder CD, Young JA, Lemarié DP, and Smith DR (2002) Influence of eastern hemlock (*Tsuga canadensis*) forests on aquatic invertebrate assemblages in headwater streams. *Canadian Journal of Fisheries & Aquatic Sciences* 59: 262–275.
47. Vannote RL and Sweeney BW (1980) Geographic analysis of thermal equilibria: A conceptual model for evaluating the effect of natural and modified thermal regimes on aquatic insect communities. *The American Naturalist* 115: 667–695.
48. Wagner R (1991) The influence of the diel activity pattern of the larvae of *Sericostoma personatum* (Kirby and Spence) (Trichoptera) on organic matter distribution in stream-bed sediments – A laboratory study. *Hydrobiologia* 224: 65–70.
49. Wallace JB and Webster JR (1996) The role of macroinvertebrates in stream ecosystem function. *Annual Review of Entomology* 41: 115–139.
50. Wallace JB, Webster JR, and Woodall WR (1977) The role of filter-feeders in flowing waters. *Archiv für Hydrobiologie* 79: 506–532.
51. Walter JK, Bilby RE, and Fransen BR (2006) Effects of Pacific salmon spawning and carcass availability on the caddisfly *Ecclisomyia conspersa* (Trichoptera: Limnephilidae) *Freshwater Biology* 51: 1211–1218.
52. Whiles MR and Wallace JB (1997) Leaf litter decomposition and macroinvertebrate communities in headwater streams draining pine and hardwood catchments. *Hydrobiologia* 353: 107–119.
53. Wolf B, Zwick P, and Marxsen J (1997) Feeding ecology of the freshwater detritivore *Ptychoptera paludosa* (Diptera, Nematocera). *Freshwater Biology* 38: 375–386.
54. Wooster D (1994) Predator impacts on stream benthic prey. *Oecologia* 99: 7–15.
55. Wotton RS, Malmqvist B, Muotka T, and Larsson K (1998) Fecal pellets from a dense aggregation of suspension-feeders in a stream: An example of ecosystem engineering. *Limnology and Oceanography* 43: 719–725.

predators have the ability to search in sites similar to those being used by their prey, and the latter may respond by actively entering the water column and drifting downstream. Foraging by invertebrate predators can also influence the downstream movement of inorganic material through their physical activities. Several studies have suggested that the foraging activities increase erosion and downstream transport of sand and fine sediments. Furthermore, some specialized parasites (a subcategory of predators) of scraping *Glossosoma* caddisflies have been shown to influence periphyton biomass in Michigan streams.

### Anthropogenic and Natural Disturbances to Invertebrate Productivity in Small Streams

Many unique fauna are found in small streams. Unfortunately, invertebrate fauna in these streams are under assault by anthropogenic and natural disturbances such as invasive species, agriculture, development, logging, mining, recreational activities, global climate change, and wildfires (Table 3). Macroinvertebrate communities and productivity can be altered, which can affect higher trophic levels (e.g., fish production) and other stream processes (e.g., organic matter processing).

#### Invasive Species

Invasive species within riparian habitats can have lasting effects on headwater stream functioning because of the tight linkage between riparian forests and stream processes. Macroinvertebrates abundance and diversity in small streams can be altered by changes in microclimate, energy availability, and habitat that results from loss of tree species within the riparian forest. Outbreaks of terrestrial invaders such as the balsam and hemlock woolly adelgids and the gypsy moth result in losses of some riparian tree species, pulses of slow decaying wood inputs, increases in other tree species, and increases of pesticides used to control invading pests. Indirectly, these changes can affect headwater stream functioning through reductions in the survival, growth, and emergence of macroinvertebrate shredders and detrital processing. The effects of terrestrial invasive species on stream ecosystems are expected to increase in the future. Nonnative scale and fungal diseases such as dogwood anthracnose and beech bark disease have invaded forests of the eastern United States, and the fungus causing butternut canker is beginning to spread rapidly. The fungus causing chestnut blight eliminated the American chestnut from eastern forests and resulted in decreased leaf litter processing, decreased quality of

litter inputs, and decreased invertebrate growth rates in headwater streams. The input of dead chestnut logs into streams also facilitated the retention of sediment and served to stabilize stream channels. Few examples of exotic aquatic species invading small streams have been documented in the literature. One species that successfully invaded first- and second-order streams, *Gammarus pulex*, resulted in spatial and temporal reductions in macroinvertebrate diversity.

#### Agriculture

The filling of former wetlands and headwater streams for agriculture has greatly reduced surface water area worldwide. As an example, 96.6% of the original surface water area of the Kävlinge River catchment in Sweden has been lost due to channelization and drainage of streams for agriculture over a 141-year period. Along with the loss of small streams, intensive agriculture results in excessive nitrate levels in stream water. Overfertilization of agricultural land in low-order sections of river networks affects downstream river reaches. It has been estimated that agricultural sources in Illinois contribute 10–15% of nitrogen and phosphorus loads to the Mississippi River. Nutrient enrichment in small streams can stimulate primary production and higher trophic levels such as scrapers that feed on the abundant periphyton. In detritus-based streams, increased nutrients can lead to increases in microbial production on organic matter, which improves the quality of the food resource for shredder invertebrates. With higher food quality, macroinvertebrate production, particularly those taxa with short life cycles, can increase dramatically in nutrient-enriched streams.

Shifts in the invertebrate community associated with increased sedimentation have been observed in headwater reaches of agriculturally impacted streams. As the percent fine sediment increases, there is usually a shift from clinging and crawling taxa to burrowers. Insecticide runoff from agricultural fields into headwater streams can have more deleterious impacts on macroinvertebrate communities. Pesticides introduced into headwater streams can result in the loss of invertebrate species, cause shifts in functional production of invertebrates, and negatively impact ecosystem processes such as leaf litter breakdown and FPOM export.

#### Urbanization and Roads

Urban growth scenarios predict substantial increases in population and growth for many regions of the world. The replacement of forested land and riparian habitats with impervious surfaces such as roads,

**Table 3** Examples of disturbances and their effects on benthic invertebrates in headwater streams

| <i>Disturbance</i>                                                                                                                                               | <i>Effect on invertebrates and stream function</i>                                                                                                                                                                                               | <i>Source</i>                    |
|------------------------------------------------------------------------------------------------------------------------------------------------------------------|--------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|----------------------------------|
| Invasive species                                                                                                                                                 | Accelerated detritus processing                                                                                                                                                                                                                  | 27                               |
| Gypsy moth defoliation                                                                                                                                           | Reduction in alpha and gamma diversity of invertebrates and changes in trophic composition; pesticide inputs caused decline in invertebrate emergence; increased inputs of slow decaying wood; higher hydrologic variability                     | 19, 20, 51                       |
| Decline in eastern hemlock forests due to hemlock woolly adelgid                                                                                                 | Decrease in leaf litter processing, quality of litter inputs, and invertebrate growth rates; increase in wood inputs and sediment stabilization                                                                                                  | 50, 57                           |
| Loss of American chestnut trees as a result of chestnut blight                                                                                                   | Increased predation on native invertebrates                                                                                                                                                                                                      | 28, 29                           |
| Invasion of <i>G. pulex</i>                                                                                                                                      | Reduced drainage density of stream network                                                                                                                                                                                                       | 59                               |
| Agriculture                                                                                                                                                      | Increased nitrogen alter food resources for invertebrates; increase in abundance, biomass and production of invertebrates; increased growth rates for short-lived invertebrates                                                                  | 13, 14, 16                       |
| Filling and tiling of streams                                                                                                                                    | Decline in Ephemeropteran, Plecopteran, and Trichopteran, and Coleopteran taxa; increases in chironomids, oligochaetes and molluscs                                                                                                              | 2, 6, 7, 24, 39, 44, 53          |
| Overapplication of fertilizer                                                                                                                                    | Decline in invertebrate abundance, biomass, and production; loss of species; shifts in functional structure; decline in organic matter export and leaf breakdown rates                                                                           | 15, 31, 32, 33                   |
| Sediment runoff                                                                                                                                                  | Decline in Ephemeroptera, Plecoptera, and Trichoptera                                                                                                                                                                                            | 24, 44                           |
| Insecticide runoff                                                                                                                                               | Decline in invertebrate diversity; decline in Ephemeroptera, Plecoptera, and Trichoptera; increase in number of pollution-tolerant taxa; decline in invertebrate production; decline in leaf breakdown rates                                     | 4, 9, 30, 39, 40, 48, 58, 60     |
| Increased water temperature                                                                                                                                      | Reduced adult caddisfly diversity and abundance above culverts                                                                                                                                                                                   | 5                                |
| Urbanization and roads                                                                                                                                           | Shift from allochthonous to autochthonous energy; increase in abundance, biomass, and production of taxa with short life cycles; leaf litter breakdown altered; significant reduction in invertebrate production with decline in detrital inputs | 1, 3, 21, 23, 41, 42, 52, 55, 56 |
| Altered hydrology and geomorphology, increased bacterial populations and turbidity, increases of pesticide, herbicide, and fertilizer runoff; decline in habitat | Decline in total richness and abundance of all invertebrate taxa                                                                                                                                                                                 | 22, 54                           |
| Increased number of culverts                                                                                                                                     | Reductions in abundances of sensitive invertebrate taxa; increase in tolerant taxa; decline in species diversity; increased drift; reduced community respiration; reduced secondary production                                                   | 8, 11, 12, 18, 34, 45, 49        |
| Forestry practices                                                                                                                                               | Elimination of all biota in buried streams; downstream declines in Ephemeroptera richness, decline in abundances of Ephemeroptera, Odonata, Coleoptera; decline in scraper and shredder abundance                                                | 25, 43                           |
| Increased stream temperature, discharge, nutrients, and primary production; reduced organic matter inputs                                                        | Altered organic matter standing crops, and invertebrate abundance and production; shifts from large-bodied, long-lived taxa to small-bodied, short-lived taxa                                                                                    | 10, 17                           |
| Sediment runoff from logging roads                                                                                                                               |                                                                                                                                                                                                                                                  |                                  |
| Mining                                                                                                                                                           |                                                                                                                                                                                                                                                  |                                  |
| Acid mine drainage and metal uptake                                                                                                                              |                                                                                                                                                                                                                                                  |                                  |
| Mountaintop mining/Valley fill                                                                                                                                   |                                                                                                                                                                                                                                                  |                                  |
| Burial of headwater streams; increased sedimentation, conductivity, and metals                                                                                   |                                                                                                                                                                                                                                                  |                                  |
| Global climate change                                                                                                                                            |                                                                                                                                                                                                                                                  |                                  |
| Channel drying                                                                                                                                                   |                                                                                                                                                                                                                                                  |                                  |

Continued

**Table 3** Continued

| <i>Disturbance</i>                                                                                                         | <i>Effect on invertebrates and stream function</i>                                                                                                                             | <i>Source</i>          |
|----------------------------------------------------------------------------------------------------------------------------|--------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|------------------------|
| Increased water temperature                                                                                                | Decline in total invertebrate densities, faster growth rates, reduced size at maturity, and altered sex ratios of some taxa                                                    | 26                     |
| Wildfire                                                                                                                   |                                                                                                                                                                                |                        |
| Intense heating, altered water chemistry, food resources, hydrologic runoff patterns, vegetative cover, sediment transport | Shift in functional feeding groups that parallel changes in food resources; shift toward short-lived, trophic generalists; decline in invertebrate abundance and taxa richness | 35, 36, 37, 38, 46, 47 |
| Recreational activities                                                                                                    |                                                                                                                                                                                |                        |
| Streamside camping, fishing, swimming, rafting, gold mining                                                                | Localized decline in abundance of scraper limnephilid caddisfly                                                                                                                | 61                     |

## Sources

1. Baillie BR, Collier KJ, and Nagels J (2005) Effect of forest harvesting and woody debris removal on two Northland streams, New Zealand. *New Zealand Journal of Marine and Freshwater Research* 39: 1–15.
2. Barton DR and Farmer MED (1997) The effects of conservation tillage practice on benthic invertebrate communities in headwater streams in southwestern Ontario, Canada. *Environmental Pollution* 96: 207–215.
3. Benfield EF, Webster JR, Tank JL, and Hutchens JJ (2001) Long-term patterns in leaf breakdown in response to watershed logging. *International Review of Hydrobiology* 86: 467–474.
4. Blakely TJ and Harding JS (2005) Longitudinal patterns in benthic communities in an urban stream under restoration. *New Zealand Journal of Marine and Freshwater Research* 39: 17–28.
5. Blakely TJ, Harding JS, McIntosh AR, and Winterbourn MJ (2005) Barriers to the recovery of aquatic insect communities in urban streams. *Freshwater Biology* 51: 1634–1645.
6. Braccia A and Voshell, JR Jr. (2005) Environmental factors accounting for benthic macroinvertebrate assemblage structure at the sample scale in streams subjected to a gradient of cattle grazing. *Hydrobiologia* 573: 55–73.
7. Braccia A and Voshell JR Jr. (2007). Benthic macroinvertebrate responses to increasing levels of cattle grazing in Blue Ridge Mountain Streams, Virginia, USA. *Environmental Monitoring and Assessment* 131: 185–200.
8. Carlisle DM and Clements WH (2003) Growth and secondary production of aquatic insects along a gradient of Zn contamination in Rocky Mountain streams. *Journal of the North American Benthological Society* 22: 582–597.
9. Chadwick MA, Dobberfuhl DR, Benke AC, Huryn AD, Suberkropp K, and Thiele JE (2006) Urbanization affects stream ecosystem function by altering hydrology, chemistry, and biotic richness. *Ecological Applications* 16: 1796–1807.
10. Chadwick MA and Huryn AD (2007) Role of habitat in determining macroinvertebrate production in an intermittent-stream system. *Freshwater Biology* 52: 240–251.
11. Clements WH (2004) Small-scale experiments support causal relationships between metal contamination and macroinvertebrate community responses. *Ecological Applications* 14: 954–967.
12. Clements WH, Carlisle DM, Lazorchak JM, Johnson PC (2000) Heavy metals structure benthic communities in Colorado mountain streams. *Ecological Applications* 10: 626–638.
13. Cross WF, Johnson BR, Wallace JB, and Rosemond AD (2005) Contrasting response of detritivores to long-term nutrient enrichment. *Limnology and Oceanography* 50: 1730–1739.
14. Cross WF, Wallace JB, Rosemond AD, and Eggert SL (2006) Whole-system nutrient enrichment increases secondary production in a detritus-based ecosystem. *Ecology* 87: 1556–1565.
15. Cuffney TF, Wallace JB, and Lughart GJ (1990) Experimental evidence quantifying the role of benthic invertebrates in organic matter dynamics of headwater streams. *Freshwater Biology* 23: 281–299.
16. David MB and Gentry LE (2000) Anthropogenic inputs of nitrogen and phosphorus and riverine export for Illinois, USA. *Journal of Environmental Quality* 29: 494–508.
17. Dewson ZS, James ABW, and Death RG (2007) Invertebrate responses to short-term water abstraction in small New Zealand streams. *Freshwater Biology* 52: 357–369.
18. Dills G and Rogers DT (1974) Macroinvertebrate community structure as an indicator of acid mine pollution. *Environmental Pollution* 6: 239–261.
19. Ellison AM, Bank MS, Clinton BD *et al.* (2005) Loss of foundation species: Consequences for the structure and dynamics of forested ecosystems. *Frontiers in Ecology and the Environment* 3: 479–486.
20. Griffith MB, Barrows EM, and Perry SA (1996) Effects of aerial application of diflubenzuron on emergence and flight of adult aquatic insects. *Journal of Economic Entomology* 89: 442–446.
21. Griffith MB and Perry SA (1991) Leaf pack processing in 2 Appalachian Mountain streams draining catchments with different management histories. *Hydrobiologia* 220: 247–254.
22. Gowns IO and Davis JA (1994) Effects of forestry activities (clearfelling) on stream macroinvertebrate fauna in south-western Australia. *Australian Journal of Marine and Freshwater Research* 45: 963–975.
23. Gurtz ME and Wallace JB (1984) Substrate-mediated response of stream invertebrates to disturbance. *Ecology* 65: 1556–1569.
24. Harding JS, Young RG, Hayes JW, Shearer KA, and Stark JD (1999) Changes in agricultural intensity and river health along a river continuum. *Freshwater Biology* 42: 345–357.
25. Hartman KJ, Kaller MD, Howell JW, and Sweka JA (2005) How much do valley fills influence headwater streams? *Hydrobiologia* 532: 91–102.
26. Hogg ID and Williams D Dudley (1996) Response of stream invertebrates to a global-warming thermal regime: An ecosystem-level manipulation. *Ecology* 77: 395–407.
27. Hutchens JJ and Benfield EF (2000) Effects of forest defoliation by the gypsy moth on detritus processing in southern Appalachian streams. *American Midland Naturalist* 143: 397–404.

28. Kelly DW and Dick JTA (2005) Effects of environment and an introduced invertebrate species on the structure of benthic macroinvertebrate species at the catchment level. *Archiv für Hydrobiologie* 164: 69–88.
29. Kelly DW, Dick JTA, Montgomery WI, and MacNeil C (2003) Differences in composition of macroinvertebrate communities with invasive and native *Gammarus* spp. (Crustacea: Amphipoda) *Freshwater Biology* 48: 306–315.
30. Kemp SJ and Spotila JR (1997) Effects of urbanization on brown trout (*Salmo trutta*), other fishes and macroinvertebrates in Valley Creek, Valley Forge, Pennsylvania. *American Midland Naturalist* 138: 55–68.
31. Liess M and Schulz R (1999) Linking insecticide contamination and population response in an agricultural stream. *Environmental Toxicology and Chemistry* 18: 1948–1955.
32. Lughart GJ and Wallace JB (1992) Effects of disturbance on benthic functional structure and production in mountain streams. *Journal of the North American Benthological Society* 11: 138–164.
33. Lughart GJ, Wallace JB, and Huryn AD (1990) Secondary production of chironomid communities in insecticide-treated and untreated headwater streams. *Freshwater Biology* 24: 417–427.
34. Malmqvist B and Hoffsten P (1999) Influence of drainage from old mine deposits on benthic macroinvertebrate communities in central Swedish streams. *Water Research* 33: 2415–2423.
35. Mihuc TB and Minshall GW (2005) The trophic basis of reference and post-fire stream food webs 10 years after wildfire in Yellowstone National Park. *Aquatic Sciences* 67: 541–548.
36. Minshall GW (2003) Responses of stream benthic macroinvertebrates to fire. *Forest Ecology and Management* 178: 155–161.
37. Minshall GW, Brock JT, Andrews DA, and Robinson CT (2001) Water quality, substratum and biotic responses of five central Idaho (USA) streams during the first year following the Mortar Creek fire. *International Journal of Wildland Fire* 10: 185–199.
38. Minshall GW, Royer TV, and Robinson CT (2001) Response of the Cache Creek macroinvertebrates during the first 10 years following disturbance by the 1988 Yellowstone wildfires. *Canadian Journal of Fisheries & Aquatic Sciences* 58: 1077–1088.
39. Moore AA and Palmer MA (2005) Invertebrate biodiversity in agricultural and urban headwater streams: Implications for conservation and management. *Ecological Applications* 15: 1169–1177.
40. Morse CC, Huryn AD, and Cronan C (2003) Impervious surface area as a predictor of the effects of urbanization on stream insect communities in Maine, USA. *Environmental Monitoring and Assessment* 89: 95–127.
41. Newbold JD, Erman DC, and Roby KB (1980) Effects of logging on macroinvertebrates in streams with and without buffer strips. *Canadian Journal of Fisheries & Aquatic Sciences* 37: 1076–1085.
42. Noel DS, Martin CW, and Federer CA (1986) Effects of forest clearcutting in New England on stream macroinvertebrates and periphyton. *Environmental Management* 10: 661–670.
43. Pond GJ and McMurray SE (2002) A macroinvertebrate bioassessment index for headwater streams in the eastern coalfield region, Kentucky. Kentucky Department for Environmental Protection, Division of Water, Frankfort, KY.
44. Quinn JM, Williamson RB, Smith RK, and Vickers ML (1992) Effects of riparian grazing and channelisation on streams in Southland, New Zealand. 2. Benthic invertebrates. *New Zealand Journal of Marine and Freshwater Research* 26: 259–273.
45. Roback SS and Richardson JW (1969) The effects of acid mine drainage on aquatic insects. *Proceedings of the Academy of Natural Sciences Philadelphia* 121: 81–107.
46. Robinson CT, Uehlinger U, and Minshall GW (2005) Functional characteristics of wilderness streams twenty years following wildfire. *Western North American Naturalist* 65: 1–10.
47. Roby KB and Azuma DL (1995) Changes in a reach of a northern California stream following wildfire. *Environmental Management* 19: 591–600.
48. Roy AH, Rosemond AD, Paul MJ, Leigh DS, and Wallace JB (2003) Stream macroinvertebrate response to catchment urbanisation (Georgia, USA). *Freshwater Biology* 48: 329–346.
49. Short TM, Black JA, and Birge WJ (1990) Effects of acid-mine drainage on the chemical and biological character of an alkaline headwater stream. *Archives of Environmental Contamination and Toxicology* 19: 241–248.
50. Smock LA and MacGregor CM (1988) Impact of the American chestnut blight on aquatic shredding macroinvertebrates. *Journal of the North American Benthological Society* 7: 212–221.
51. Snyder CD, Young JA, Lemarié DP, and Smith DR (2002) Influence of eastern hemlock (*Tsuga canadensis*) forests on aquatic invertebrate assemblages in headwater streams. *Canadian Journal of Fisheries and Aquatic Sciences* 59: 262–275.
52. Stone MK and Wallace JB (1998) Long-term recovery of a mountain stream from clear-cut logging: The effects of forest succession on benthic invertebrate community structure. *Freshwater Biology* 39: 141–169.
53. Stone ML, Whiles MR, Webber JA, Willard KWJ, and Reeve JD (2005) Macroinvertebrate communities in agriculturally impacted southern Illinois streams: Patterns with riparian vegetation, water quality, and in-stream habitat quality. *Journal of Environmental Quality* 34: 907–917.
54. Tebo LB Jr. (1955) Effects of siltation, resulting from improper logging, on the bottom fauna of a small trout stream in the southern Appalachians. *Progressive Fish-Culturist* 17: 64–70.
55. Wallace JB, Eggert SL, Meyer JL, and Webster JR (1999) Effects of resource limitation on a detrital-based ecosystem. *Ecological Monographs* 69: 409–442.
56. Wallace JB and Gurtz ME (1986) Response of *Baetis* mayflies (Ephemeroptera) to catchment logging. *American Midland Naturalist* 115: 25–41.
57. Wallace JB, Webster JR, Eggert SL, Meyer JL, and Siler ER (2001) Large woody debris in a headwater stream: Long-term legacies of forest disturbance. *International Review of Hydrobiology* 86: 501–513.
58. Whiting ER and Clifford HF (1983) Invertebrates and urban runoff in a small northern stream, Edmonton, Alberta, Canada. *Hydrobiologia* 102: 73–80.
59. Wolf P (1956) *Utdikad Civilisation. (Drained Civilization)*. Malmö, Sweden: Gleerups.
60. Woodcock TS and Huryn AD (2007) The response of macroinvertebrate production to a pollution gradient in a headwater stream. *Freshwater Biology* 52: 177–196.
61. Wright KK and Li JL (1998) Effects of recreational activities on the distribution of *Dicosmoecus gilvipes* in a mountain stream. *Journal of the North American Benthological Society* 17: 535–543.

rooftops, and lawns alters the hydrology and geomorphology of streams. Increases in surface runoff associated with storm flow lead to declines in water quality, increases in bacterial populations and turbidity, and increases of pesticide, herbicide, and fertilizer runoff into nearby streams. Sediment runoff from construction sites and erosion due to downcutting result in habitat loss for aquatic life. Measurable aquatic degradation occurs with 6–10% impervious area and has been long been associated with decreased water quality of nearby streams. Urbanization in watersheds containing small streams usually results in less diverse invertebrate communities consisting of pollution-tolerant species. Culverts have been shown to act as barriers to upstream migration of adult caddisflies. Since many stressors associated with urbanization act synergistically, it is difficult to separate cause and effect of individual stressors on the invertebrate communities in small urban streams.

### Forestry Practices

Logging results in changes in stream temperature regimes, increased discharge and altered hydrology, increased nutrient export and increased solar radiation and primary production, increased sediment export, and changes in dissolved organic matter derived from the terrestrial ecosystem. These changes are accompanied by substantial changes in the energy base of headwater streams, with a shift from allochthonous detritus to autochthonous production. The physical and energy base changes can lead to large changes in macroinvertebrate community structure. An experimental long-term reduction of organic matter inputs to a small stream in the southern Appalachians resulted in a significant decline in total invertebrate production. Invertebrate taxa with short life cycles and the ability to exploit increases in primary production greatly increase in abundance, biomass, and productivity. Studies in the central and southern Appalachians show that long-term patterns of leaf litter breakdown can be altered for many years following logging. However, depending upon the extent of terrestrial succession, invertebrate assemblages can revert back toward their prelogged condition.

### Mining

Mining has severe consequences for benthic invertebrates in small streams worldwide. Effects of mining on macroinvertebrates in small streams are caused by acid mine drainage, sediments, or burial of the streams themselves. Acid mine drainage and the associated problems of heavy metal contamination usually results in reductions of numbers of sensitive

taxa in the orders of Ephemeroptera (particularly those of the family Heptageniidae), Plecoptera, Trichoptera, Megaloptera, Odonata, and Diptera and an overall decline in species diversity. Some studies have shown that functional measures of benthic invertebrates such as drift and community respiration are also negatively affected by mining impacts. In recent years, the practice of mountaintop removal and valley fill mining has resulted in the filling and permanent burial of at least 1450 km of small streams in the Appalachian Mountains. The burial of multiple small streams destroys all aquatic life in these streams and results in declines of sensitive invertebrate taxa immediately below valley fills. The cumulative effects of burying multiple headwater streams on the water quality in downstream rivers should be evaluated.

### Recreational Activities

Little information regarding the effects of recreational activities (e.g., horseback riding, cycling, all terrain vehicle use (ATV)) on small streams has been reported in the primary literature. One study suggested that populations of *Dicosmeocus gilvipes*, a scraping limnephilid caddisfly, in a fifth-order stream were affected by localized disturbances associated with multiple recreational activities such as gold mining, streamside camping, swimming, and fishing. With growing public demand for access to undeveloped land harboring networks of small streams for recreational activities such as off-highway vehicle use, there is an urgent need for more research examining the impacts of such use and ways to mitigate potential negative effects.

### Global Climate Change

Consequences of global climate change on invertebrates in small streams will vary greatly spatially and temporally, thus making it difficult to predict potential effects. Generally, precipitation and evaporation are expected to become more variable over time. Some regions of the world will become wetter, while others will become drier, affecting runoff patterns. Increased temperatures as a result of global climate change will reduce snow cover and also affect hydrologic patterns in small streams. Shifts in hydrologic patterns (e.g., flooding, drying) will impact transport of nutrients, organic matter, and habitats available for colonization by benthic invertebrates. Changes in riparian vegetation may alter the quality and quantity of detrital inputs to headwater streams, thereby altering ecosystem processes (e.g., production, respiration, organic matter breakdown)

within small stream reaches and longitudinally-linked downstream reaches, as well as invertebrate life histories and species composition. There is some evidence in the literature that the timing and duration of small stream channel drying results in altered organic matter standing crops, and invertebrate production. Furthermore, extended channel drying results in shifts from large-bodied, long-lived taxa to small-bodied, short-lived taxa. An experimental manipulation of thermal regime (2.5–3.0 °C increase in water temperature) in a small stream near Ontario, Canada, resulted in a reduction of total invertebrate densities, faster growth rates, reduced size at maturity, and altered sex ratios of some invertebrates.

### Wildfire

Invertebrates in small streams are more susceptible to fire disturbance than those in larger streams. Intense heating, severely altered water chemistry, and the smothering of food resources by ash in smaller streams can kill invertebrates directly. Over longer time periods, changes in hydrologic runoff patterns, vegetative cover, channel morphology, and sediment transport also affect invertebrates in fire impacted streams. Changes in food resources over time result in changes in the functional characteristics of the macroinvertebrate community. Initially, scraper densities increase following a fire because of increased primary productivity associated with canopy opening and increased available nutrients. As transportable organic matter levels increase in the stream, abundances of collectors increase. Shredder populations are usually the last to recover since they depend on detrital inputs from the riparian habitat. Recovery of macroinvertebrate communities in intact, normally functioning small streams prior to fire usually occurs quickly (5–10 years) following fire disturbance and parallels the regeneration of the terrestrial vegetation. Short-lived invertebrate taxa that are trophic generalists, and have wide habitat preferences generally recover quicker.

### Summary and Knowledge Gaps

The functional contributions of benthic invertebrates to small streams are well known. Hundreds of invertebrate species may be found in a small stream. Since headwater streams make up such a large proportion of total stream length in river networks, total invertebrate production in small stream segments may exceed that in large rivers. Invertebrates also represent an important link between terrestrial and aquatic ecosystems due to the close proximity of the two

systems. A variety of environmental factors influence the types and productivity of invertebrates in small streams. Benthic invertebrates are also good indicators of the health of small streams. Human and natural disturbances alter typical macroinvertebrate assemblages in small streams which may have indirect effects on higher trophic levels and small stream processes.

In the last two decades scientists have begun to study macroinvertebrate communities and the ecological processes affected by invertebrates along longitudinal reaches spanning multiple stream orders. However, little information is known about the quantitative and qualitative contribution of headwater benthic fauna to the functioning of downstream ecosystems. In some cases, entire benthic invertebrate communities are being destroyed by burial or stream piping before the true diversity of organisms found in small streams is known. Furthermore, with an increasing number of disturbances that are large scale in magnitude, it is critical that scientists become better able to predict threshold levels of disturbance within headwaters of river networks such that downstream water quality and ecosystem functions are not irrevocably damaged.

See *also*: Acidification; Agriculture; Aquatic Insects – Ecology, Feeding, and Life History; Aquatic Insects, Classification; Aquatic Plants: A General Introduction; Aquatic Plants and Attached Algae; Benthic Invertebrate Fauna, Small Streams; Benthic Invertebrate Fauna; Bioassessment of Aquatic Ecosystems; Biodiversity of Aquatic Ecosystems; Biological Interactions in River Ecosystems; Climate and Rivers; Coarse Woody Debris in Lakes and Streams; Comparative Primary Production; Conservation of Aquatic Ecosystems; Decapoda; Deforestation and Nutrient Loading to Fresh Waters; Diptera (Biting Flies); Diptera (Non-Biting Flies); Ecology and Role of Headwater Streams; Effects of Recreation and Commercial Shipping; Ephemeroptera (Mayflies); Fires; Flatworms (Turbellarians); Floods; Fluvial Export; Gas Exchange at the Air-water Interface; Gastrotricha; Geomorphology of Streams and Rivers; Hemiptera (True Bugs); Hydrachnida (Water Mites); Invasive Species; Isopoda (Aquatic Sowbugs); Littoral Zone; Megaloptera (Alderflies, Dobsonflies); Mercury Pollution in Remote Freshwaters; Microbial Food Webs; Mollusca; Mosses (Bryophytes); Natural Organic Matter; Nematoda; Nematomorpha (Horsehair Worms); Neuston in Aquatic Ecosystems; Odonata (Dragonflies and Damselflies); Plecoptera (Stoneflies); Regulators of Biotic Processes in Stream and River Ecosystems; Restoration Ecology of Rivers; Restoration of Acidic Drainage; Riparian Zones; Streams; Subterranean Aquatic Ecosystems – Groundwater Ecology; Tardigrada (Water Bears); Trichoptera (Caddisflies); Trophic Dynamics in Aquatic Ecosystems; Urban Aquatic Ecosystems.

## Further Reading

- Allan JD (1995) *Stream Ecology: Structure and Function of Running Waters*. New York, NY: Chapman & Hall.
- Cushing CE, Cummins KW, and Minshall GW (2006) *River and Stream Ecosystems of the World*. Berkeley, CA: University of California Press.
- Freeman MC, Pringle CM, and Jackson CR (2007) Hydrologic connectivity and the contribution of stream headwaters to ecological integrity at regional scales. *Journal of the American Water Resources Association* 43: 5–14.
- Heino J, Muotka T, Mykrä H, Paavola R, Hämäläinen H, and Koskenniemi E (2003) Defining macroinvertebrate assemblage types of headwater streams: Implications for bioassessment and conservation. *Ecological Applications* 13: 842–852.
- Hurn AD and Wallace JB (2000) Life history and production of stream insects. *Annual Review of Entomology* 45: 83–110.
- Merritt RW, Cummins KW, and Berg MB (2007) *An Introduction to the Aquatic Insects of North America*, 4th edn. Dubuque, IA: Kendall/Hunt.
- Merritt RW and Cummins KW (2006) Trophic relationships of macroinvertebrates. In: Hauer FR and Lamberti GA (eds.) *Methods in Stream Ecology*, 2nd edn., pp. 585–609. UK: Elsevier.
- Meyer JL, Strayer DL, Wallace JB, Eggert SL, Helfman GS, and Leonard NE (2007) The contribution of headwater streams to biodiversity in river networks. *Journal of the American Water Resources Association* 43: 86–103.
- Meyer JL and Wallace JB (2001) Lost linkages and lotic ecology: Rediscovering small streams. In: Press MC, Huntly NJ, and Levin S (eds.) *Ecology: Achievement and Challenge*, pp. 295–317. Oxford, UK: Blackwell Science.
- Nadeau TL and Rains MC (2007) Hydrological connectivity between headwater streams and downstream waters: How science can inform policy. *Journal of the American Water Resources Association* 43: 118–133.
- Resh VH and Rosenberg DM (1984) *The Ecology of Aquatic Insects*. New York: Praeger.
- Thorp JH and Covich AP (eds.) (2001) *Ecology and Classification of North American Freshwater Invertebrates*, 2nd edn. San Diego, CA: Academic Press.
- Vannote RL, Minshall GW, Cummins KW, Sedell JR, and Cushing CE (1980) The river continuum concept. *Canadian Journal of Fisheries and Aquatic Sciences* 37: 130–137.
- Wallace JB and Merritt RW (1980) Filter-feeding ecology of aquatic insects. *Annual Review of Entomology* 25: 103–132.
- Wallace JB and Webster JR (1996) The role of macroinvertebrates in stream ecosystem function. *Annual Review of Entomology* 41: 115–139.
- Wipfli MS, Richardson JS, and Naiman RJ (2007) Ecological linkages between headwaters and downstream ecosystems: Transport of organic matter, invertebrates, and wood down headwater channels. *Journal of the American Water Resources Association* 43: 72–85.

## Relevant Websites

- <http://www.americanrivers.org/site/DocServer/WhereRiversAreBorn1.pdf?docID=182> – American Rivers: Where Rivers are Born.
- <http://www.benthos.org> – North American Benthological Society.
- <http://www.tolweb.org/tree/phylogeny.html> – Tree of Life.
- <http://www.epa.state.oh.us/dsw/wqs/> – Ohio EPA, Division of Surface Water, Ohio Primary Headwater Habitat Streams.
- <http://www.stroudcenter.org/index.htm> – Stroud Water Research Center.