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# A density management diagram for Norway spruce in the temperate European montane region

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**Abstract** Norway spruce is one of the most important conifer tree species in Europe, paramount for timber provision, habitat, recreation, and protection of mountain roads and settlements from natural hazards. Although natural Norway spruce forests exhibit diverse structures, even-aged stands can arise after disturbance or as the result of common silvicultural practice, including off-site afforestation. Many even-aged Norway spruce forests face issues such as senescence, insufficient regeneration, mechanical stability, sensitivity to biotic disturbances, and restoration. We propose the use of Density Management Diagrams (DMD), stand-scale graphical models designed to project growth and yield of even-aged forests, as a heuristic tool for assessing the structure and development of even-aged Norway spruce stands. DMDs are predicated on basic tree allometry and the assumption that self-thinning occurs predictably in forest stands. We designed a DMD for Norway spruce in temperate Europe based on wide-ranging forest inventory data. Quantitative relationships between tree- and stand-level variables that describe resistance to selected natural disturbances were superimposed on the DMD. These susceptibility zones were used to demonstrate assessment and possible management actions related to, for example, windfirmness and effectiveness of the protective function against rockfall or avalanches. The Norway spruce DMD provides forest managers and silviculturists a simple, easy-to-use, tool for evaluating stand dynamics and scheduling needed density management actions.

**Keywords** Decision support systems · Natural hazards · *Picea abies* (L.) Karst. · Protective function · Self-thinning · Silviculture

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## Introduction

Norway spruce (*Picea abies* (L.) Karst.) is one of the most important tree species in the mountain ranges of central and southern Europe. Norway spruce stands are important for timber production and provide relevant ecosystem services (Pretzsch et al. 2008). In mountain regions, these forests can provide protection from natural hazards such as avalanches, rockfall, or landslides (Bebi et al. 2001; Mayer and Ott 1991). Norway spruce forests also provide habitat for game, and may harbor endangered fauna or flora (e.g., Nascimbene et al. 2009).

Vast areas of pure, monolayered Norway spruce plantations are common in European montane and lowland landscapes, oftentimes usurping the space of natural forests (Hansen and Spiecker 2004). The species has been introduced far outside its natural range, both in countries where it occurs naturally, for example, in Germany and Norway, and in novel areas such as Denmark, Belgium, and Ireland (Skroppa 2003). Natural and semi-natural Norway spruce forests, on the other hand, are relatively rare (Parviainen et al. 2000; Motta 2002)

and often exhibit multiple structural and compositional attributes depending in part on the disturbance regime (Shohorova et al. 2009). These structures range from sparse, multilayered subalpine stands (Kulakowski et al. 2004; Krumm et al. 2011) to monolayered forests resulting from severe disturbances (Fisher et al. 2002; Angelstam and Kuuluvainen 2004), to uneven-aged mixtures (Svoboda et al. 2010, 2012).

Windstorms, snow loading, and insects are among the most damaging disturbance agents in Norway spruce stands (Klopcic et al. 2009; Svoboda et al. 2012). Increasing susceptibility to natural disturbances (Schlyter et al. 2006; Seidl et al. 2011), in combination with aging stands and increasing demand for enhanced structural complexity and close-to-nature forest structures (Gamborg and Larsen 2003), results in a silvicultural conundrum that cannot be adequately addressed using simple management tools (e.g., yield tables). Given the importance of Norway spruce in managed montane forests of central-southern Europe, it is important to develop ecologically based decision support systems that allow for the development of realistic management scenarios, and enable the comparison of alternative schedules with respect to the evaluation criteria of interest (e.g., volume production, carbon storage, stand stability, structural diversity, nature conservation, and biodiversity).

Density management diagrams (DMD) are empirical models of even-aged stand dynamics (Jack and Long 1996). They reflect fundamental relationships involving tree size, stand density, site occupancy, and self-thinning. Allometric relationships between mean tree size, age, height, and yield are portrayed allowing users to design treatments by plotting both current and desired future stand structure on the DMD. Alternative management strategies that accomplish diverse objectives can be simultaneously compared and their efficacy evaluated at a glance. In this paper, we analyzed data from Norway spruce stands to construct a DMD with wide applicability across montane regions of central-southern Europe. Using specific examples of (1) maximizing volume production, (2) mechanical stability against wind damage, (3) avalanche protective function, and (4) potential resistance to spruce bark beetle (*Ips typographus* L.), we demonstrate the usefulness of the Norway spruce DMD.

## Methods

### Data sources

The data used to develop the Norway spruce DMD (Table 1) covered many regions of central-southern Europe (Fig. 1) and included 5,656 plots. Most areas occupied by temperate European montane forest were represented, with the exception of areas with few pure Norway spruce forests (e.g., Balkans) or countries where forest inventory data were not readily accessible.

1. Data from France were obtained from the French National Forest Inventory (http://www/.ifn.fr/spip/) for the inventory period 2005–2009. The French inventory design implemented three nested fixed-area plots [6, 9, and 15 m radius for trees 7–22.5 cm, 22.6–37.5 cm, and 37.5+ cm in diameter at breast height (DBH), respectively] from which trees per hectare (N) expansion factors were calculated. The French Inventory also included tree height (H) and estimated tree volume (Vidal et al. 2007).
2. Data from the Czech Republic came from two regions, Sumava and Tajga. In the Sumava region, the inventory design was three nested fixed-area plots (3.5, 7, and 12.6 m radius for trees 7–14.9 cm, 15–29.9, and 30+ cm DBH, respectively) and did not include estimates of tree volumes (Čížková et al. 2011). In the Tajga region, the inventory scheme consisted of one 12.5-m-radius fixed-area plot where DBH and H were measured and estimates of volume included for all trees >10 cm DBH.
3. Data from Romania came from the mountain regions of Călimani and Giumalau (Cenuşă 1992). The inventory in these regions used either a 500- or 1,000-$m^2$ fixed-area plot with a lower DBH cutoff of 10 cm. Tree heights were estimated using locally calibrated models and there were no estimates of volume (M. Svoboda—unpublished data).
4. Italian data came from multiple regions and inventory designs. At Aosta and Piemonte (IPLA 2003) fixed-area plots ranging from 8 to 15 m radius were used, depending on overstory density, and the lower DBH cutoff was ~7 cm; species-and site-specific volume equations were provided. At Paneveggio and San Martino (Berretti and Motta 2005) fixed-area plots of 12 m radius with a lower DBH cutoff of 17 cm were used and no estimates of volume were made. At Val Pontebbana (Castagneri et al. 2010) 12-m-radius fixed-area plots were sampled with a lower DBH cutoff of ~7 cm. In Valbona, 400-$m^2$ fixed-area plots were used with a lower DBH cutoff of ~7 cm (Motta et al. 2006). At Burgusio, Lasa, Latemar, Luttago, Meltina, Naturno, Valle Aurina, and for plots of the National Forest Inventory (INFC 2006), variable radius plots (basal area factor = 4 $m^2$ $ha^{-1}$) were employed with a lower DBH cutoff of ~4 cm and volume was not estimated.
5. Bulgarian data referred to forest patches in the Parangalitsa Reserve, including a number of post-disturbance stands (Panayotov et al. 2011). A total of

**Table 1** Data sources for the Norway spruce DMD, measurement protocol, and maximum Stand Density Index ($SDI_{max}$ by location for pure, even-aged Norway spruce stands)

| ID | Dataset name | Country | No. of plots | DBH cutoff (cm) | Plot size ($m^2$) | $SDI_{max}$ |
|---|---|---|---|---|---|---|
| 1 | Aosta | Italy | 156 | 7 | 201–707 | 1209 |
| 2 | Piemonte | Italy | 65 | 7 | 201–707 | 1701 |
| 3 | National forest inventory | Italy | 401 | 4 | Relascopic | 1571 |
| 4 | Burgusio | Italy | 91 | 4 | Relascopic | 1080 |
| 5 | Lasa | Italy | 251 | 4 | Relascopic | 1473 |
| 6 | Latemar | Italy | 322 | 4 | Relascopic | 1745 |
| 7 | Luttago | Italy | 72 | 4 | Relascopic | 1007 |
| 8 | Meltina | Italy | 256 | 4 | Relascopic | 1383 |
| 9 | Naturno | Italy | 304 | 4 | Relascopic | 1220 |
| 10 | Valle Aurina | Italy | 155 | 4 | Relascopic | 1493 |
| 11 | Paneveggio | Italy | 91 | 17 | 452 | 1321 |
| 12 | San Martino | Italy | 91 | 17 | 452 | 1278 |
| 13 | Valbona | Italy | 66 | 7 | 400 | 1592 |
| 14 | Val Pontebbana | Italy | 33 | 7 | 452 | 1162 |
| 15 | Tajga | Czech Republic | 78 | 7 | 491 | 755 |
| 16 | Sumava Certovo | Czech Republic | 66 | 7 | 38–499 | 1278 |
| 17 | Sumava NP | Czech Republic | 38 | 7 | 38–499 | 1221 |
| 18 | Sumava large plots | Czech Republic | 15 | 7 | 1,000–2,500 | 1121 |
| 19 | Sumava Trojmezna | Czech Republic | 18 | 7 | 38–499 | 826 |
| 20 | Călimani | Romania | 40 | 10 | 500–1,000 | 1425 |
| 21 | Giumalau | Romania | 41 | 10 | 500–1,000 | 1270 |
| 22 | Baden-Wurttnenberg | Germany | 399 | 7 | Relascopic | 1464 |
| 23 | France 2005 | France | 522 | 7 | 113–707 | 1206 |
| 24 | France 2006 | France | 526 | 7 | 113–707 | 1277 |
| 25 | France 2007 | France | 558 | 7 | 113–707 | 1305 |
| 26 | France 2008 | France | 489 | 7 | 113–707 | 1086 |
| 27 | France 2009 | France | 471 | 7 | 113–707 | 1238 |
| 28 | Parangalitsa | Bulgaria | 227 | 4 | 100 | 2653 |

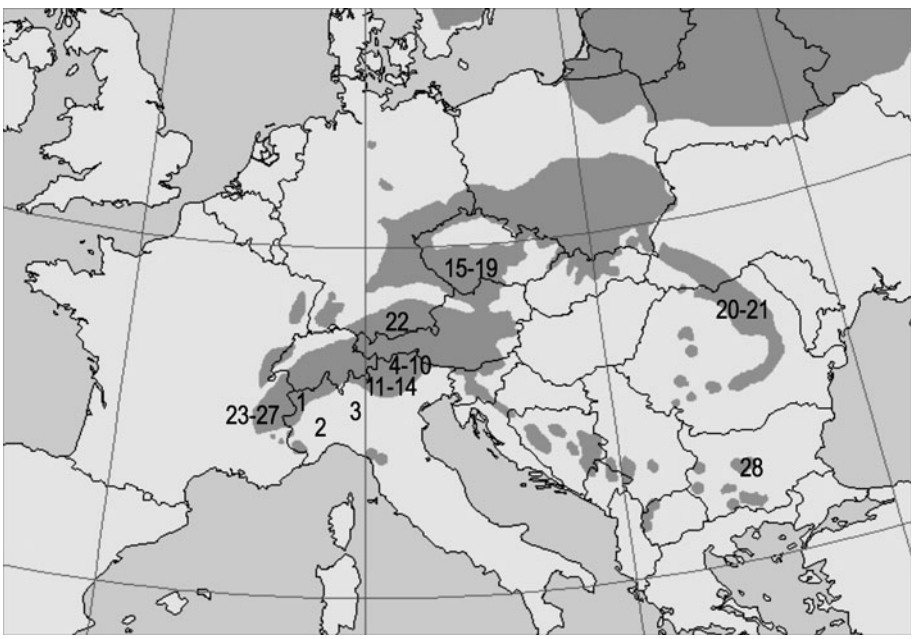


**Fig. 1** Distribution of Norway spruce in central-southern Europe (after Schmidt-Vogt 1977) and ID code for data used for DMD construction. Refer to Table 1 for location names

227 100-m$^2$ plots were sampled with a lower DBH cutoff of 4 cm and no information on H and volume.

6. German data came from the Second National Forest Inventory of Germany (Schmidt and Kandler 2009). Trees with a minimum DBH of 7 cm were selected using variable radius plots with a basal area factor of 4 m$^2$ ha$^{-1}$. The attributes recorded included species, DBH, tree age, and H.

## Size–density relationships

Using the tree-level data, we calculated N, quadratic mean diameter (QMD), basal area, percent basal area of Norway spruce, stand density index (SDI), and stand top height ($HT_{100}$), defined as the average height of the 100 largest (DBH) trees per hectare. SDI was calculated two ways: (1) Reineke ($SDI_p$: Reineke 1933, modified by Long and Daniel 1990),

$$SDI_p = N(QMD/25.4)^{1.605}, \quad (1)$$

and (2) summing the SDI of each $i$-th tree in a stand ($SDI_{sum}$: Shaw 2000),

$$SDI_{sum} = \Sigma_N \left[ N_i (DBH_i/25.4)^{1.605} \right] \quad (2)$$

so that stands with simple structure could be filtered from the data using the $SDI_{sum}$:$SDI_p$ ratio ($SDI_{ratio}$). $SDI_{ratio}$ has been shown to theoretically differentiate even-aged stands, which have strong unimodal diameter distributions ($SDI_{ratio} \geq 0.9$), from uneven-or multi-aged stands, which show increasing skewness in their diameter distribution ($SDI_{ratio} < 0.9$) (Ducey 2009).

Before estimating the self-thinning boundary, all data were filtered for Norway spruce composition ≥80 % (determined by percent basal area) and even-agedness ($SDI_{ratio} \geq 0.9$), which resulted in 1609 plots. In order to filter for fully stocked stands, we used a binning method (Bi and Turvey 1997) (200 N bins) from which maximum observations of $SDI_{sum}$ were extracted, and the maximum self-thinning line was fit by ordinary least-squares (OLS) regression. We assessed whether a lower DBH cutoff of 4, 7, 10, or 17 cm had any effect on $SDI_{max}$ (Curtis 2010) and/or the slope of the self-thinning line fitted using the respetive thresholds. Moreover, since differing self-thinning slopes are reported in the literature, both between- and within-tree species (including Norway spruce: Sterba 1987; Hynynen 1993; Monserud et al. 2005; Pretzsch and Biber 2005; Pretzsch 2006; Schütz and Zingg 2010; Charru et al. 2011), we tested whether Reineke's (1933) suggested slope of −1.605 was statistically different from that of our linear fit. Subsequently, we shifted the OLS line to cross the point of maximum stocking. $SDI_{max}$ indicates maximum growing space occupancy (Yoda et al. 1963), so that plots falling above the line should be exceedingly rare. Therefore, we assumed the 98th percentile of the $SDI_{sum}$ frequency distribution appropriately characterized the maximum attainable SDI. Finally, we juxtaposed lines on the DMD to describe relative stand density (percent of $SDI_{max}$) following the recommendations of Long (1985). That is, 25 % of $SDI_{max}$ represents crown closure, 35 % of $SDI_{max}$ indicates the beginning of individual-tree growth reduction due to inter-tree competition, and 60 % of $SDI_{max}$ the onset of severe competition.

We tested for the existence of a Mature Stand Boundary (MSB) in the maximum self-thinning limit (Shaw and Long 2007) by fitting the following three-parameter function:

$$QMD = a(N_{max} + b)^c, \quad (3)$$

where $N_{max}$ are observations of maximum N for each 0.01 class of $Log_{10}$ QMD. Only plots with QMD ≥ 15 cm were used, because stands in the smaller size classes are not needed to establish the MSB. Subsequently, we shifted the curve developed in Eq. 3 so that the maximum SDI value on the curve was asymptotic to the $SDI_{max}$ on the DMD.

## Top height and volume

When included on a DMD, $HT_{100}$ can be used with local site index curves to assess the temporal development of a particular stand (Jack and Long 1996). Using plot data that included observations of $HT_{100}$, we modeled QMD as a function of $HT_{100}$, attenuated by an inverse logarithmic function of tree density:

$$QMD = HT_{100}(b_1 - b_2 \ln N) \quad (4)$$

To generate stand-level volume (VOL) isolines on the DMD, we modeled VOL as a power function of QMD and N (Eq. 5a), then rewrote the equation as QMD = $f$(VOL), where VOL is total standing volume (m$^3$ ha$^{-1}$) for plot data with volume observations:

$$VOL = c_1 + c_2 N\ QMD^{c_3} \quad (5a)$$

$$QMD = \left[ (c_1 + c_2 N)^{-1} VOL \right]^{(1/c_3)} \quad (5b)$$

We plotted $HT_{100}$ and VOL isolines on the DMD for ranges of 20–50 m, and 200–1200 m$^3$ ha$^{-1}$, respectively. Different inventories may have used different equations for tree or stand volume, generating idiosyncrasies when pooling all volume data in one model. However, because we were missing inventory-specific volume equations, we used original data as much as possible, acknowledging that

DMD isolines merely represent average conditions across the entire dataset.

All models were assessed for parameter significance and goodness-of-fit by computing adjusted $R^2$ and root mean square error (RMSE). We determined that both models had little or no bias by inspecting residual plots over the predictor variables, elevation when available, SDI, basal area, region, and whether the plot had a lower DBH cutoff of 4, 7, 10, or 17 cm.

### Disturbances and site index

To illustrate the advantages of the DMD in designing silvicultural strategies to maximize resistance to disturbances and protection from natural hazards, we superimposed "susceptibility zones" on the diagram, which encapsulate combinations of size and density that (1) fulfill an effective protection against avalanche release; and (2) result in a low risk of wind damage. Thresholds for (1) were summarized as follows (after Berretti et al. 2006; Gauquelin and Courbaud 2006):

(a) Basal area $\geq$25 $m^2$ $ha^{-1}$ when QMD = 25 cm, and $\geq$7.5 $m^2$ $ha^{-1}$ when QMD = 10 cm for effective snowpack stabilization if slope is steeper than 35°;

(b) Live crown ratio $\geq$60 % in trees or cluster of trees supporting the stability of the stand. We relaxed this requirement to $\geq$33 %, representing a minimal acceptable level of individual-tree vigor that should be ensured with a relative SDI <60 % of $SDI_{max}$ (Long 1985);

(c) H/DBH ratio <80 in dominant trees. H/DBH ratio cannot be read directly off the DMD. However, assuming that DBH is normally distributed in a stand and that dominant diameter (DD) is equivalent to the 90th percentile of such distribution ($Z$ value = +1.64), DD can be computed by

$$DD = 1.64\sigma_{DBH} + QMD, \quad (6a)$$

where $\sigma_{DBH}$ is the standard deviation of the DBH distribution in the stand. In order to represent risk zones on the DMD, we assumed that $\sigma_{DBH} = 0.3$ QMD and solved Eq. 6a for QMD:

$$QMD = 0.67\ DD, \quad (6b)$$

to be substituted in $HT_{100}$/QMD ratio from Eq. 4 and constrained to $\leq$0.8. This allowed the influence of smaller, suppressed trees to be removed so that only the slenderness of dominant trees was considered (Castedo-Dorado et al. 2009);

(d) Gap size $\leq$1.5 times tree height (i.e., in order to avoid tree-free patches prone to dangerous snow gliding). If square spacing is assumed, a Mean nearest neighbor distance (MNND) can be computed as the square root of the reciprocal of N. We introduced a multiplier to account for clumped patterns, that is, the ratio between maximum and observed nearest neighbor index (NNI). NNI ranges from 0, when trees are highly clumped, to 2.1491, when trees are arranged along a hexagonal grid (Clark and Evans 1954):

$$MNND = \left(2.1491\ NNI^{-1}\right) \times 100 N^{-0.5}, \quad (7)$$

subsequently constrained to $\leq$1.5$HT_{100}$ and used to back calculate critical N–$HT_{100}$ combinations.

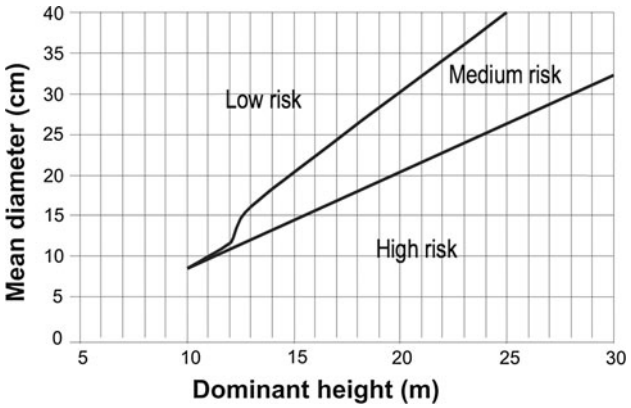


**Fig. 2** Wind stability zones for even-aged coniferous stands based upon $HT_{100}$ and QMD (after Riou–Nivert 2001)

While the DMD can be used to assess avalanche hazard related to stand structure, other predisposing conditions (e.g., weather, topography, characteristics of snowpack, and terrain ruggedness) must be evaluated independently.

Thresholds for windfirmness followed those by Riou-Nivert (2001), who established low, medium, and high wind risk zones for conifer species, based on the relationship between QMD and $HT_{100}$ (Fig. 2). Mitchell (2000) suggested that such general zones of stability exist for uniform stands of all temperate zone conifers.

An appropriate site index (SI) curve allows the estimates of $HT_{100}$ on the DMD to be a surrogate for time (Drew and Flewelling 1979). SI estimates were not included in the raw data. In order to provide SI curves applicable to even-aged, pure Norway spruce stands across temperate Europe, we fitted a modified Richards' model of height growth (Sterba 1976) to yield tables from Eisacktal, South Tyrol (Moser 1991), which exhibited a wide range of fertility classes (i.e., $HT_{100}$: 7.9–45.8 m at age 100). All statistics were performed in the R environment version 2.14.1 (R Development Core Team 2011).

## Results

Twenty-nine percent of the original Norway spruce data set, that is, 1609 of 5656 inventory plots (Table 2) were

**Table 2** Summary statistics for pure, even-aged Norway spruce stands ($SDI_{ratio} \geq 0.9$, PRCPA: Norway spruce on total basal area ≥80 %)

| Variable | Unit | No. of plots | Min | Max | Mean | Std. error |
|---|---|---|---|---|---|---|
| N | trees $ha^{-1}$ | 1609 | 14 | 5058 | 564.1 | 13.03 |
| QMD | cm | 1609 | 7.8 | 115.0 | 34.8 | 0.31 |
| $HT_{100}$ | m | 876 | 4.2 | 46.0 | 24.1 | 0.23 |
| VOL | $m^3$ $ha^{-1}$ | 505 | 0.8 | 1163.6 | 316.4 | 9.69 |
| BA | $m^2$ $ha^{-1}$ | 1609 | 0.4 | 130.0 | 40.3 | 0.50 |
| PRCPA | % | 1609 | 0.8 | 1.0 | 1.0 | 0.002 |
| $SDI_{sum}$ | – | 1609 | 14 | 2057 | 705.0 | 8.45 |
| $SDI_{ratio}$ | – | 1609 | 0.9 | 1.0 | 1.0 | 0.001 |
| Age | Years | 669 | 8.0 | 338.0 | 108.5 | 2.40 |
| Elevation | m. a.s.l. | 748 | 82 | 2,230 | 1240.6 | 16.26 |

used to fit a maximum size–density relationship characterizing montane Norway spruce in central-southern Europe. Slope of the self-thinning line was −1.497 (adjusted $R^2 = 0.94$); the 95 % confidence interval of the regression slope coefficient (−1.671 to −1.324) included Reineke's value of −1.605.

$SDI_{max}$ was 1461 (Fig. 3); coefficient of variation between the 28 regions was 26 %, mean = 1334.28, and SD = 345.39 (Table 1). Subsetting by different DBH cutoff values did not change our results with respect to the significance of −1.605, except for the 17 cm cutoff that produced a non-significant regression slope likely due to limited sample size (Table 3). However, the lowest DBH cutoff (4 cm) produced the highest $SDI_{max}$. Parameters of the MSB (Eq. 3) were $a = 3330.105$, $b = 185.158$, and $c = -0.0656$ (adjusted $R^2 = 0.96$).

Top height and volume equations were statistically significant (Table 4). Some bias was revealed in residual plots over observed volume (Fig. 4); however, these occurred in poorly stocked stands (i.e., <50 $m^3$ $ha^{-1}$) and do not constitute a concern for using the DMD in practice. The QMD-$HT_{100}$ model exhibited some high regional bias (Table 5). A 95 % confidence envelope about the mean of QMD residuals included zero in 7 out of 14 sites for the $HT_{100}$ model (Eq. 4), and 8 out of 10 sites for the VOL model (Eq. 5b).

## Discussion

### DMD characteristics

Maximum SDI for Norway spruce in montane forests of central-southern Europe was 1461, which was intermediate in the range of previous regional estimates (Pretzsch 2005—Germany: $SDI_{max} = 1609$; Monserud et al. 2005—Austria: $SDI_{max} = 1571$; Sterba 1981—Austria: $SDI_{max} = 1547$; Castagneri et al. 2008—NE Italy: $SDI_{max} = 1380$), independent of the DBH measurement cutoff. Consistent with previous studies (Shaw and Long 2007), we detected a convex pattern to the self-thinning limit at high tree size–low density combinations, that is, a mature stand boundary (MSB). The most commonly suggested explanation for this process is the decline of the so-called "self-tolerance" (Zeide 1985), by which growing space resulting from the death of very large trees can not be promptly reclaimed by conspecific neighboring trees, lowering the limit of possible size–density combinations. Maintaining stand size–density below the MSB is crucial for management as combinations above the line are ecologically improbable (DeRose et al. 2008).

DMDs that cover widely distributed species (e.g., Long and Shaw 2005) are indicative of average growth patterns and allometric relationships of monospecific stands. We assumed that allometric equations, when portrayed on the DMD, were invariant across all sites (Weiner 2004). Conditions under which the self-thinning boundary may shift include, at the local scale, genetic differences (Buford and Burkhardt 1987), and severe resource deficiencies, e.g., in tree line environments (Körner 2003). However, despite deviations at certain localities (Table 5), our allometric models should be robust, in that the high number of plots used for calibration should average out local peculiarities.

Previous research has observed disparities in mortality of Norway spruce stands located on different elevations and aspects (Krumm et al. 2012). However, we consider these to be an effect of the different rates at which stands progress along their trajectories of development in size–density space. Differences in topography, temperature, light, and soil fertility affect growth rates and, in turn, the rate of mortality during the stem-exclusion phase (Aulitzky 1984; Schönenberger 2001). In other words, a Norway spruce stand on a high-quality site will reach the boundary more quickly than on a lower-quality site, while complying to the same overarching, species-specific, self-thinning boundary (Jack and Long 1996). This constancy is fundamental to the general utility of DMD and allows the use of site index curves to determine the time required to attain particular stand structural characteristics. When using the DMD to portray stands at a specific location, managers should choose the appropriate dominant height curve, in order to account for differences in local productivity.

### Application of the DMD

The DMD is depicted in log(QMD)-log(Density) space with a superimposed self-thinning line and $HT_{100}$ and VOL isolines (Fig. 5). Application of the DMD proceeds as

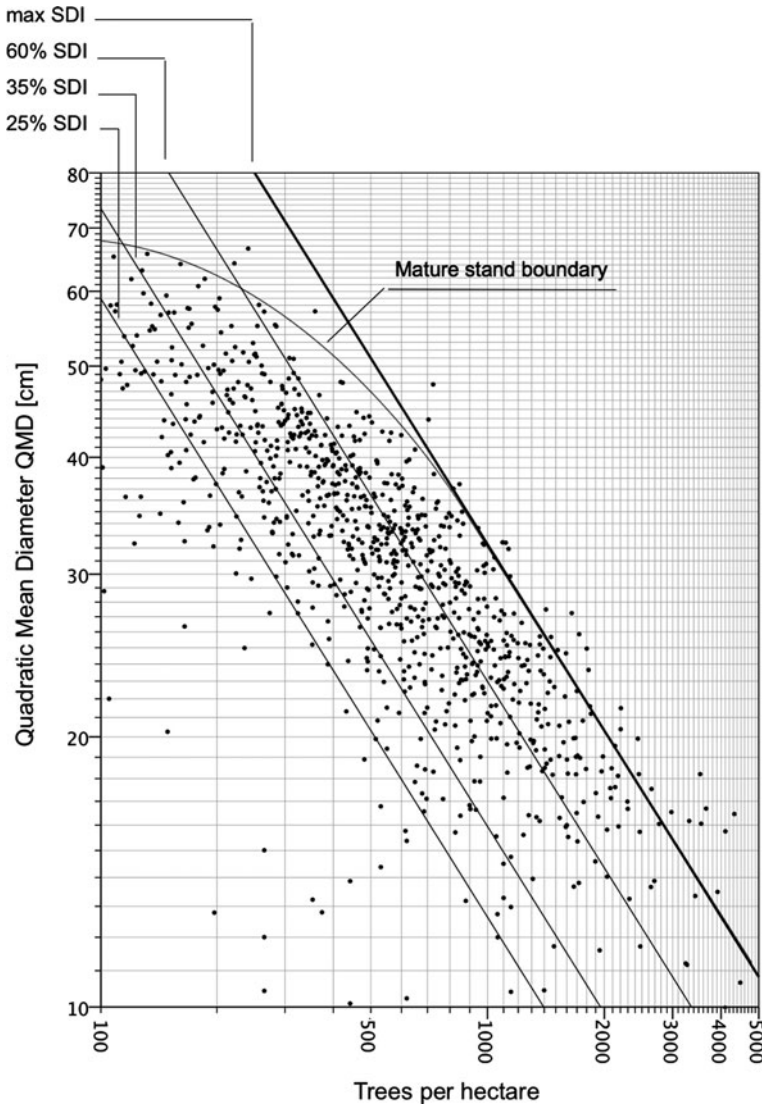


**Fig. 3** Selected Norway spruce stands in size–density space, relative SDI lines, and mature stand boundary

**Table 3** $SDI_{max}$ and self-thinning slope fitted using different DBH cutoff values

| DBH cutoff (in.) | $SDI_{max}$ | Slope | 95 % min | 95 % max | p | Adjusted $R^2$ | No. of plots |
|---|---|---|---|---|---|---|---|
| 0 | 1461 | −1.50 | −1.67 | −1.32 | 0.00 | 0.94 | 1609 |
| 4 | 1587 | −1.61 | −1.85 | −1.36 | 0.00 | 0.90 | 633 |
| 7.5 | 1287 | −1.53 | −1.95 | −1.10 | 0.00 | 0.82 | 635 |
| 10 | 1447 | −1.52 | −1.83 | −1.20 | 0.00 | 0.91 | 250 |
| 17 | 1355 | −1.87 | −3.77 | 0.04 | 0.053 | 0.56 | 91 |

follows: (1) identify starting conditions on the DMD (i.e., current stand structure); (2) identify target stand structure at end of rotation (EOR) and track the likely trajectory of unmanaged stand development (i.e., asymptotic to the self-thinning boundary); (3) ascertain the need for stand density regulation, e.g., to prevent the onset of competition-related mortality ($\sim$60 % $SDI_{max}$) and represent the planned entries on the DMD; (4) assess time to reach EOR by tracking the starting and ending $HT_{100}$ on SI curves (Fig. 6).

*Maximize volume production*

When the goal is timber production, one can use the DMD for minimizing the time required to reach EOR at a desired

**Table 4** Model fit and parameters for Eqs. 4 and 5b ($HT_{100}$ in m, QMD in cm, VOL in $m^3$ $ha^{-1}$)

| Parameter | Estimate | Std. error | 95 % min | 95 % max | Adjusted $R^2$ | No. of plots |
|---|---|---|---|---|---|---|
| $QMD = HT_{100}(b_1 - b_2 \ln N)$ | | | | | | |
| $b_1$ | 3.148 | 0.056 | 3.038 | 3.259 | 0.663 | 1491 |
| $b_2$ | 0.297 | 0.009 | 0.278 | 0.315 | | |
| $VOL = c_1 + c_2 N\ QMD^{c_3}$ | | | | | | |
| $c_1$ | −25.795 | 5.238 | −36.087 | −15.503 | 0.937 | 505 |
| $c_2$ | $1.79 \times 10^{-4}$ | $1.6 \times 10^{-5}$ | $1.46 \times 10^{-4}$ | $2.11 \times 10^{-4}$ | | |
| $c_3$ | 2.432 | 0.025 | 2.383 | 2.480 | | |

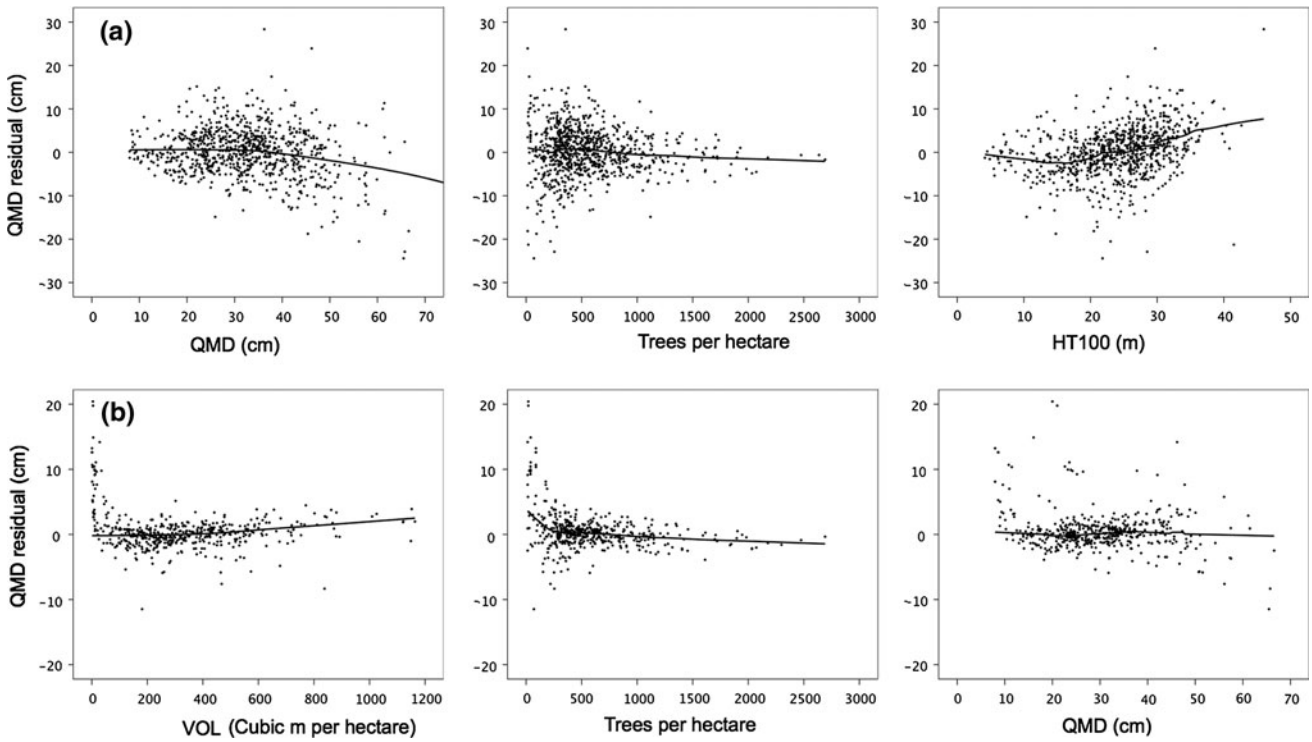


**Fig. 4** Residual plots from $HT_{100}$ (**a**) and VOL (**b**) models (Eqs. 4 and 5b). *Black lines* represent loess fit

mean stem diameter. In addition, by using the HT isolines in combination with site-specific potential productivity, one can incorporate future revenue and future costs into the density management regime. For example, if the desired EOR QMD was 40 cm, and the current stand has ~2600 trees per hectare (see Fig. 5), a thinning would be necessary to forestall density-dependent mortality when relative SDI approaches 60 %. This could be achieved by pre-commercially thinning the stand to ~400 trees per hectare. This would drive stand development on a trajectory to meet the desired EOR of 40 cm at approximately the same time maximum stand growth is achieved (relative SDI = 60 %). Both the timing and volume of the pre-commercial thinning, or any subsequent commercial thinnings, could be estimated using the HT and VOL isolines, respectively, and the return or cost associated with that treatment discounted to today's values to compare management alternatives. Similar to a volume-based regime, by using appropriate biomass conversion factors, and assuming a carbon conversion factor of 0.5, one could plan a density management regime to maximize aboveground carbon sequestration for a particular stand.

*Mechanical stability against wind damage*

Windstorms are the most destructive disturbance agent in temperate European forests (judged by the volume of timber damaged: Schelhaas et al. 2003), often causing extensive damage in Norway spruce, and in particular in structurally homogeneous stands (Schmidt-Vogt et al. 1987).

**Table 5** Confidence intervals for QMD mean bias (predicted–observed) from the $HT_{100}$ and VOL models (Eqs. 4 and 5b), by location

| Location | Mean bias QMD–$HT_{100}$ (cm) | | Mean bias QMD–VOL (cm) | |
|---|---|---|---|---|
| | Lower | Upper | Lower | Upper |
| Aosta | 1.31 | 3.17 | −0.14 | 0.37 |
| Piemonte | −5.49 | 1.88 | −1.82 | −0.37 |
| Italy | −0.30 | 1.33 | – | – |
| Valbona | −3.67 | −1.23 | −0.39 | 1.00 |
| Val Pontebbana | −3.52 | 0.68 | −0.62 | 1.05 |
| Tajga | 2.34 | 3.57 | 0.74 | 2.82 |
| Sumava NP | −2.99 | −1.10 | – | – |
| Călimani | 3.29 | 5.22 | – | – |
| Giumalau | 5.41 | 8.14 | – | – |
| France 2005 | −3.57 | −0.13 | −0.43 | 1.78 |
| France 2006 | −3.05 | −0.47 | −0.52 | 1.06 |
| France 2007 | −3.46 | 0.77 | −0.48 | 2.24 |
| France 2008 | −1.22 | 2.80 | −0.04 | 2.67 |
| France 2009 | −2.51 | 0.50 | −0.76 | 0.89 |

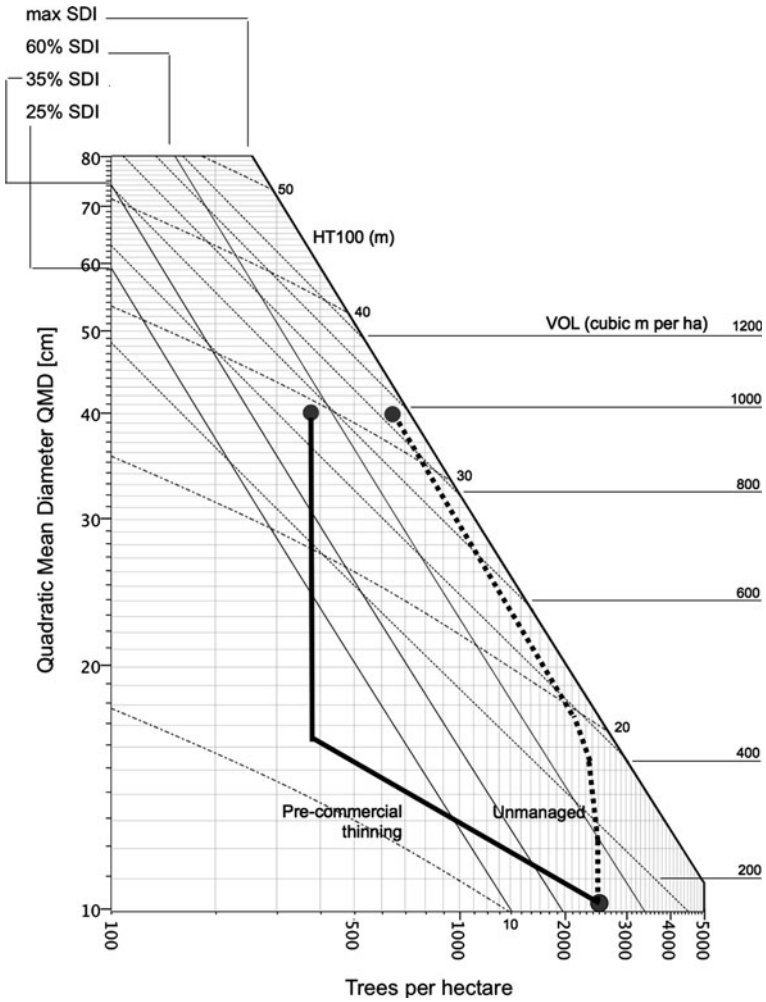


**Fig. 5** DMD for Norway spruce in the central-southern European montane ecoregion, and working example of stand trajectories for unmanaged and a pre-commercial thinning alternative (starting stand conditions: N = 2500 trees per hectare, QMD = 10 cm; end of rotation: QMD = 40 cm). Competition-related mortality onsets at 60 % $SDI_{max}$. Target QMD is reached in 70 years in the working example, as opposed to 90 years in the unmanaged alternative, on a medium fertility site (SI = 23.6 m, see Fig. 6)

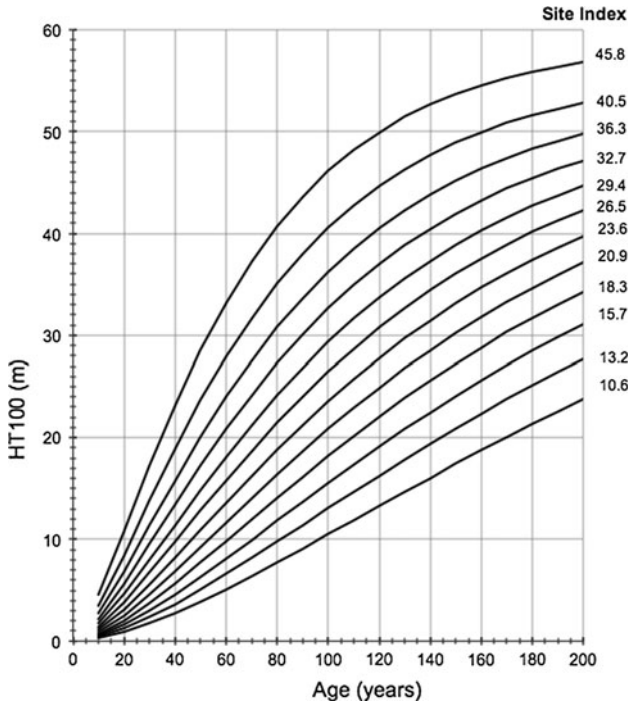


**Fig. 6** Site index curves from Eisacktal (South Tyrol) yield tables

Tree damage begins at wind speeds of 15 m $s^{-1}$ and can be catastrophic at 25 m $s^{-1}$ (Zajaczkowski 1991). Susceptibility is higher for slender trees (e.g., Rottmann 1986; Thomasius 1988; Riou-Nivert 2001; Dobbertin 2002) and with short, broad crowns (Schütz et al. 2006), a condition created through stand dynamics characterized by intense inter-tree competition. When risk zones for wind damage are superimposed on the DMD (Fig. 7), two types of management action are supported: (1) the ability to assess current conditions relative to risk, and (2) the possibility of projecting the effect of interventions which aim to maintain or drive stand structures into low-risk areas as long or quickly as possible. The second management approach is depicted in the example of an unmanaged stand trajectory portrayed in Fig. 7. Among structural attributes, a threshold of $\sim$1800 trees $ha^{-1}$ strongly differentiates high and medium susceptibility to wind damage. By contrast, the threshold to low susceptibility is mainly determined by tree slenderness, where "safe" values are typically encountered in low-density stands. From such results, we conclude that the typical even-aged Norway spruce stand (either natural or planted) is characterized by a medium risk of wind damage.

First glance at our Norway spruce stand plotted on the DMD might indicate that a heavy thinning may effectively lower stand susceptibility to wind damage, but in dense stands, it may result in sudden isolation of trees with high height-to-diameter ratio, and hence, increase the probability of damage by breakage or uprooting (Thomasius 1980).

While uneven-aged stands are acknowledged to have higher resistance to wind (e.g., Shorohova et al. 2008), they cannot be accurately represented on the diagram. Additional limitations of DMD are that (a) they cannot track risk factors unrelated to stand structure, for example, soil (trees are much more vulnerable to wind damage on shallow or wet soils) or weather, and (b) they cannot track the long-term influence of climate change on either autogenic, or allogenic growth factors.

### *Avalanche and rockfall protective function*

Because Norway spruce predominates in the upper montane and subalpine belt, it can be quite effective against the release of avalanches (although not on their transit), provided that stands meet given structure and density standards (Motta and Haudemand 2000). Like windfirmess, required stand structures can be represented as risk zones on the DMD (Fig. 8). Although individual-tree resistance parameters are similar to those required for windfirmness, effective stand structures differ because open stands with thicker trees are more prone to avalanche release due to the presence of tree-free gaps (Meyer-Grass and Schneebeli 1992; Bebi et al. 2009). By experimenting with different management regimes on the DMD (Fig. 8), we concluded that Norway spruce stands could remain within a low-risk zone for as long as 60 years, provided that site index is not too high, such as most subalpine stands (e.g., 25.2 m on average for stands at elevations >1700 m on the Eastern Alps, data from Cantiani et al. 2000). Even for high potential productivity, the low-risk period could extend up to 30 years, which would allow for spatial planning of silvicultural interventions in avalanche-prone catchments, with a goal to maintain some proportion of Norway spruce stands in the catchment as active protection forests. Boundaries for the low-risk zone could be extended by relaxing the tree slenderness or competitive status requirements. However, this would come at the expense of individual vitality and stand-scale resistance.

When the degree of tree clumping is high, it is very difficult to contrast the presence of gaps large enough to trigger potentially hazardous snow movements. Management can mitigate the tendency for large gap creation at lower elevations. For example, simulations by Cordonnier et al. (2008) suggest that by creating small gaps every 20 years, uneven-aged structure can be initiated, thereby increasing the protective function of mountain Norway spruce stands in the western Alps. In subalpine forests, which exhibit clumped spatial arrangements (Motta and Lingua 2005), stabilization of avalanche channels has to be pursued by alternative means or structures. Similar considerations could be made for rockfall, albeit using different thresholds on the DMD (Vacchiano et al. 2008).

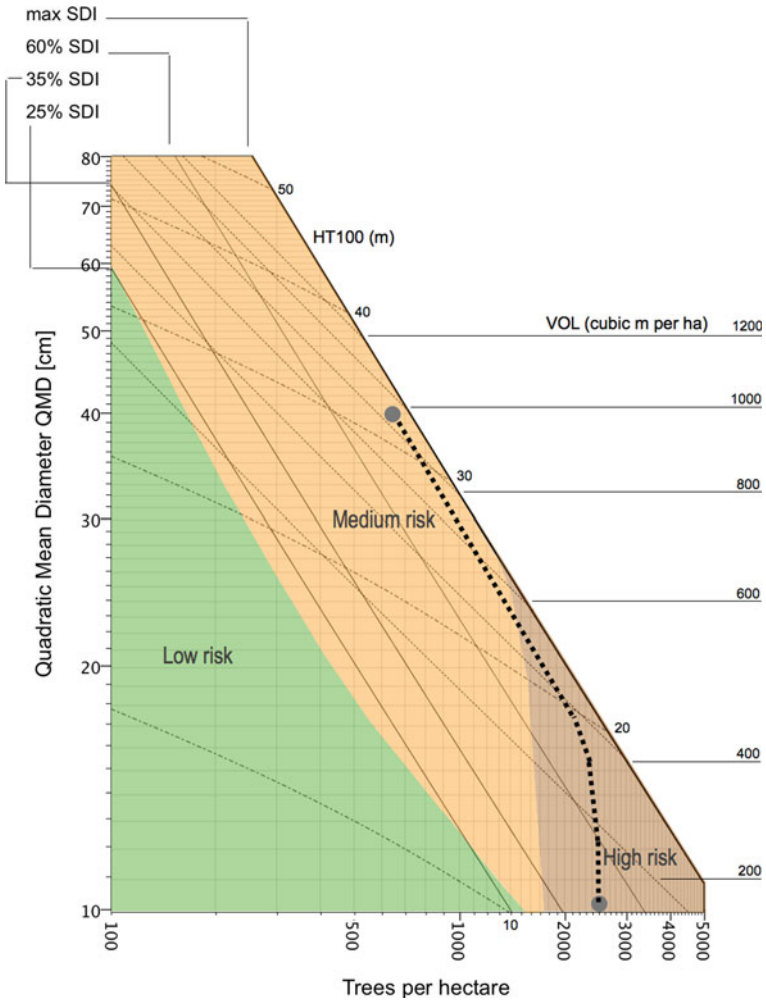


**Fig. 7** DMD and risk zones for windfirmness of Norway spruce stands. Starting stand conditions, EOR, and unmanaged stand trajectory as in working example for Fig. 5

An overlay of low-risk zones from Figs. 7 and 8 demonstrates potential conflicting management goals or desired conditions that cannot be simultaneously maximized.

*Resistance to spruce bark beetle*

In central-southern Europe, spruce bark beetle outbreaks are a part of the natural disturbance regimes of Norway spruce forests (Svoboda et al. 2012). However, mortality induced by bark beetle may severely alter structure and functionality of stands that are managed for important ecosystem services, such as protection from geological hazards (Amman 2006) or water quality (Huber 2005). Outbreaks are primarily triggered by climate and abundance of infestation source such as recent deadwood; droughts, windthrow, or pollution may decrease tree vigor and increase susceptibility, although evidence is still contradictory to this extent (Baier 1996; Dutilleul et al. 2000; Wermelinger 2004). Norway spruce trees have recently been found to be potentially more resistant to spruce bark beetle when the density of foliage or foliage packing is high (Jakuš et al. 2011), presumably as a result of the inability of adults to reach the stem. This suggests Norway spruce trees that maintain longer crowns *throughout stand development* are more likely to resist spruce bark beetle infestation. Although the DMD was developed using stand-level data, it is relatively easy to visualize stand–density combinations necessary to maintain long crowns. Assuming that full canopy closure in Norway spruce stands occurs at 25–35 % $SDI_{max}$ (Long 1985), we would seek to maintain stands on average below that level when portrayed on the DMD. While it may be possible to enhance individual-tree growth and potentially resist the beetle under this regime, it would come at the expense of stand-level growth and would almost certainly result in low-quality logs by the EOR because of large lower branches. This shows that trade-offs associated with management goals must be considered. Fortunately, they can be simultaneously portrayed on the DMD.

The ability of Norway spruce stands to meet various management objectives can be assessed on the DMD provided that associated requirements can be expressed by

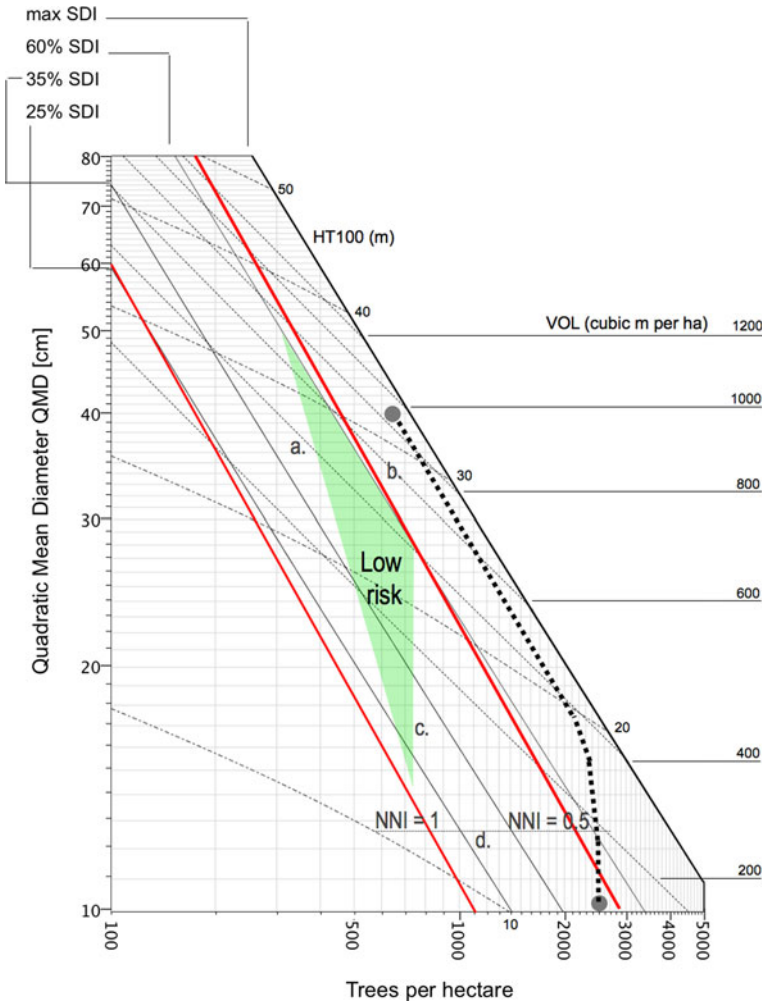


**Fig. 8** Low-risk zone for avalanche release hazard (slope = 35°). Low-risk boundaries express: *a* minimum basal area, *b* SDI for minimum crown ratio, *c* maximum $HT_{100}$/DD ratio. *Red lines:* maximum gap size for *d* NNI = 0.5 (clumped tree spatial pattern) and NNI = 1 (random pattern) according to Eq. 7. Starting stand conditions, EOR, and unmanaged stand trajectory as in working example for Fig. 5

average (or distributional) stand parameters. Possibilities include habitat quality for ungulates (Smith and Long 1987) and birds (Shaw and Long 2007). For example, the DMD can be used to project which density regime would promote tree growth of the dominant cohort and speed up the creation of future veteran trees that will serve as habitat when alive or standing dead, or to estimate the time necessary for conversion from monocultures to mixed natural forest by using the MSB to manage for time required to form stable canopy gaps.

## Conclusion

The proposed DMD represents a marked improvement in Norway spruce density management over conventional approaches, because it characterizes ecological processes that drive growth and mortality. Statistical results for the stand-scale DMD suggest it is adequately robust for use over the geographic area covered by our analysis. The DMD allows the silviculturist to graphically display current stand conditions, and to project stand development after treatment with respect to density-dependent mortality and susceptibility to natural hazards or disturbance. Multiple management scenarios can be simultaneously portrayed on the DMD to assess which EOR goals in terms of tree size, density, volume, and ecosystem services can be met, how much time is required to meet them, and how long they can be maintained by management.

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