

Estimates of carbon stored in harvested wood products from United States Forest Service Southwestern Region, 1909-2012



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Abstract

Global forests capture and store significant amounts of carbon through photosynthesis. When carbon is removed from forests through harvest, a portion of the harvested carbon is stored in wood products, often for many decades. The United States Forest Service (USFS) and other agencies are interested in accurately accounting for carbon flux associated with harvested wood products (HWP) to meet greenhouse gas monitoring commitments and climate change adaptation and mitigation objectives. National-level forest carbon accounting has been in place for over a decade, but there is an increasing need for accounting at the scale of smaller administrative units, including USFS National Forest System regions and individual National Forests. This paper uses the Intergovernmental Panel on Climate Change (IPCC) production accounting approach to estimate HWP carbon storage 1909 to 2012 for the USFS Southwestern Region, (R3). For the Southwestern Region as a whole, carbon stocks in the HWP pool were growing at an average rate of 0.11 million megagrams of carbon (MgC) per year in a steady increase from its origin until its peak cumulative storage of 9.7 million MgC in 1994. Net positive flux into the HWP pool over this period correlates closely with harvest levels. In the years between the late 1960s and 1989 timber harvest continued at high but notably volatile levels. Within this net growth period, a significant crash which bottomed in 1982 rebounded to new peak of 0.76 million MgC in 1989. Carbon entering the HWP declined steeply thereafter beginning in 1990, when harvests ultimately fell over the next six years to less than 1/10th of their peak levels. Since 1995, emissions from HWP at solid waste disposal sites have exceeded additions from harvesting, resulting in a decline in the total amount of carbon stored in the HWP pool. The Southwestern Region's HWP pool is now in a period of negative net annual stock change because the decay of products harvested between 1909 and 2012 exceeds additions of carbon to the HWP pool through harvest. Together with estimates of ecosystem carbon, which are also being developed through the Forest Management Carbon Framework (ForCaMF), Regional level estimates of HWP carbon flux can be used to inform management decisions and guide climate change adaptation and mitigation efforts by the agency. Though our emphasis is on the Southwestern Region as a whole, this accounting method can be applied more broadly at smaller land management units, such as National Forests.

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Cover: (L-R) Photo #1 of timber loading on Coconino National Forest, Arizona (National Archives); Photo #2 ripping boards for pallet manufacture; Photo #3 end product, softwood pallets. Photo's #2 and #3 provided by Dave Cummings.

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Background

Recent estimates of net annual storage (flux) indicate that the world's forests are an important carbon sink, removing more carbon dioxide (CO₂) from the atmosphere through photosynthesis than they emit through combustion and decay (Pan et al. 2011). The forest sector of the United States (US) currently stores about 45 billion megagrams of carbon (MgC), or the equivalent of about 24 years of total US emissions at the 2010 rate (US EPA 2012).

Nationally, net additions to ecosystem and harvested wood products (HWP) pools have been estimated at 251.4 million MgC yr⁻¹ (US EPA 2012), with US forests offsetting about 13.5% of the country's annual fossil fuel emissions. About 5.5% of total US forest sector carbon stocks and 7.1% of the annual flux is attributable to carbon in HWP. Increasing social and managerial interest in mitigating rising atmospheric CO₂ concentrations and the resulting impacts on climate has focused attention on the ecosystem service of forest carbon storage, including storage in HWP.

As defined by the Intergovernmental Panel on Climate Change (IPCC), HWP are products made from wood including lumber, panels, paper, paperboard, and wood used for fuel (Skog 2008). The HWP carbon pool includes both products in use and products that have been discarded to solid waste disposal sites (SWDS). Additions to the HWP pool are made through harvesting, and emissions result from decay and combustion of wood products. Forest management can affect the quantity of carbon stored in both ecosystems and forest products over time, and management activities in the US frequently include silvicultural treatments that produce HWP. Credible information on forest ecosystem and HWP carbon stocks and fluxes can inform forest managers and the public of the tradeoffs between carbon storage and other forest management objectives, and between the short and long-term carbon consequences of alternative forest management strategies (Ryan et al. 2010, McKinley et al. 2011, Galik and Jackson 2009). Though the HWP fraction of the pool is small compared to ecosystem carbon, it is an important component of national level carbon accounting and reporting.

There is growing interest among forest managers in monitoring and managing forests for sequestration of carbon as an ecosystem service. For example, during 2010, the US Forest Service (USFS) developed a climate change scorecard that will be completed annually for each of the 155 National Forests and grasslands managed by the agency (USFS 2011). The scorecard includes four categories of scored elements: organizational capacity, engagement, adaptation, and mitigation and sustainable consumption. Elements under mitigation and sustainable consumption direct individual National Forests to develop a baseline assessment of carbon stocks, as well as an assessment of the influence of disturbance and management activities on these stocks. These assessments are meant to guide mitigation actions and monitoring. Managers are expected to begin integrating carbon stewardship with management of their forest for traditional multiple uses and other ecosystem services (USFS 2011). Consequently, these requirements necessitate robust and accessible monitoring systems that provide quantitative metrics to gauge progress.

HWP carbon monitoring systems have been implemented at the national level (US EPA 2012, Skog 2008, IPCC 2006, Smith et al. 2006). Robust inventory-based methods for estimating carbon stocks and flux in forest ecosystems are well established in the US and several tools are available to forest managers (Smith et al. 2006, 2004, Zheng et al. 2010, Galik et al. 2009). However, many of the tools used to estimate carbon stored in forests do not provide estimates of HWP carbon (e.g., U.S. Forest Carbon Calculation Tool, Smith et al. 2007) while others are restricted to national level HWP accounting (e.g., WOODCARB II, Skog 2008). Neither model independently serves National Forest managers who need accessible and practical tools for estimating and monitoring carbon stocks and flux in HWP, which were harvested since the inception of their units, at the regional or National Forest levels (Ingerson 2011, Stockmann et al. 2012).

Objectives

There is a clear need to develop the means to monitor the contribution of HWP to carbon pools and greenhouse gas mitigation resulting from National Forest harvests both at the regional and forest levels. Our objectives are to:

- 1) Use an established accounting approach to make estimates of HWP carbon stocks and fluxes for the USFS, Southwestern Region;
- 2) Provide a framework with clear metrics and estimation methods that can be applied to other land management units, including individual National Forests.

We do not develop a system for evaluating the future impacts of specific management actions, nor do we advocate any particular course of action to improve carbon stewardship.

Regional Description

The USFS Southwestern Region (R3) follows the combined extent of Arizona and New Mexico. There are six national forests in Arizona (Apache-Sitgreaves, Coconino, Coronado, Kaibab, Prescott and Tonto) and five national forests in New Mexico (Carson, Cibola, Gila, Lincoln and Santa Fe). R3 represents 20.3 million acres (approximately 10.8%) of all National Forest System (NFS) lands (USDA, 2012).

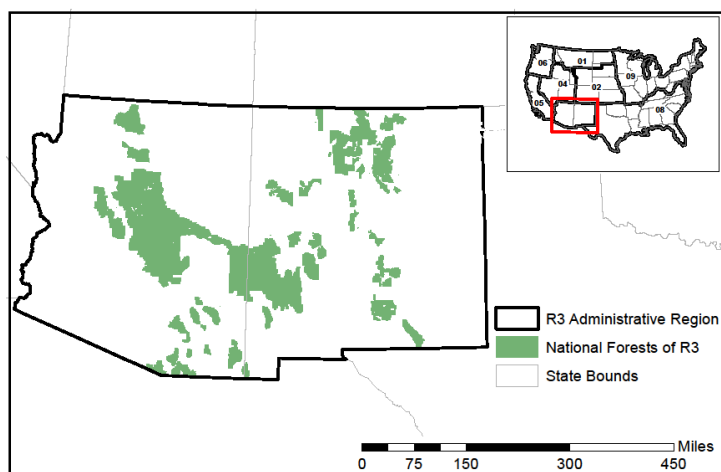


Figure 1. Map of the Southwestern Region (also known as R3).

Historical Southwestern Region land base changes

Forestland included in many Forest Service Regions has changed over time. In cases where administrative boundaries between Regions have changed, we used forest-specific data to standardize Regional harvest totals. Specifically for R3, the boundaries have been fixed at the extents of Arizona and New Mexico since the region's beginnings. Administrative boundary changes among national forests within the Region did occur, but would only be a consideration should further study at forest-level be attempted. Twenty-three national forests have evolved into the current eleven in the Southwestern Region (Davis 1983). However, records indicate that most of these changes occurred in the early years of the region when harvest volumes were low.

Methods

The method used to estimate carbon stored in HWP for the Southwestern Region is discussed here in four parts: accounting approach, computational methods, data sources, and uncertainty analysis. The first part provides a general overview of the framework used for carbon accounting, including defining the scope of analysis, relevant carbon pools, and associated fluxes. The second part provides detailed information about the data we used in our calculations that transform harvest data into carbon accounting metrics. Then we describe the origins of the data used in this analysis, with an emphasis on understanding what inputs are required and how data quality can vary over time. Lastly, the quantitative treatment of uncertainty is discussed in light of limitations of the approach used, computational methods, and data.

Accounting Approach

We use the IPCC production accounting approach, which has been adopted by the US Environmental Protection Agency (EPA; hereafter referred to as the IPCC/EPA approach) to estimate annual changes in HWP pools from the Region (Figure 2). In the IPCC/EPA approach, the annual carbon stock change for the Region's forest sector is a function of carbon flow among the atmosphere, forest ecosystems, and HWP, and is calculated as:

$$\Delta S = (NEE - H) + (\Delta C_R)$$

In this equation ΔS is the annual stock change for the Region's forest sector, NEE is the annual net ecosystem exchange between the atmosphere and the Region's forests from all ecosystem processes including photosynthesis, decay, and natural and anthropogenic fire, H is the annual harvest of wood from the Region's forests for products, and ΔC_R is the annual change in carbon stored in HWP that were made from wood harvested from the Region's national forests (Table 1, Figure2). In the IPCC/EPA approach, the annual change in carbon stored in HWP (ΔC_R) is the sum of the net change in carbon stored in products in use ($\Delta C_{IU\ R}$) and the net change in carbon stored in products at solid waste disposal sites ($\Delta C_{SWDS\ R}$) (Table 1). By estimating stocks and emissions for regional HWP carbon on an annual basis, we can calculate the annual stock change in the HWP carbon pool (ΔC_R), which is the relevant metric for this accounting approach. HWP carbon stock and flux estimates presented here are part of a larger Forest Carbon Management Framework (ForCaMF) intended to address carbon storage in the entire forest system (ΔS).

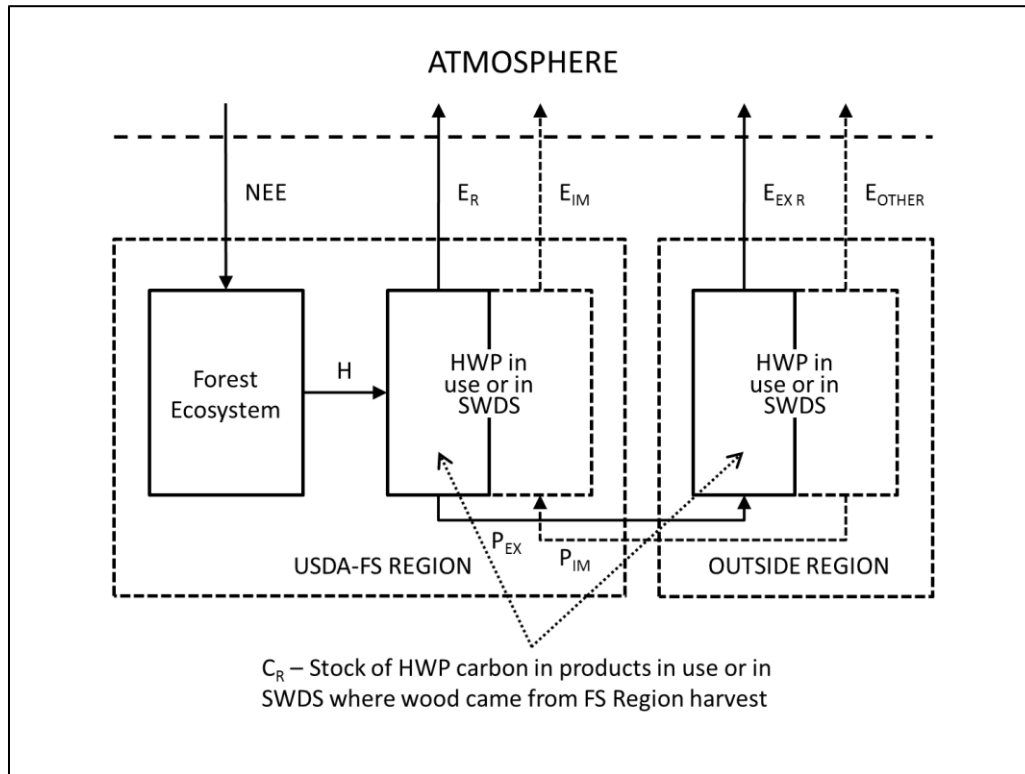


Figure 2. Carbon flows and stocks associated with forest ecosystems and harvested wood products (HWP) to illustrate the IPCC/EPA production accounting approach (adapted from Skog 2008).

Table 1. Variable definitions for the IPCC/EPA production accounting approach shown in Figure 2 (Skog 2008). Units for all variables are MgC yr⁻¹.

Variable	Definition
ΔS	Annual carbon stock change, which is calculated as $\Delta S = (NEE - H) + (\Delta C_{RI})$ in the production accounting approach.
NEE	Annual net ecosystem carbon exchange, the annual net carbon that moves from the atmosphere to forests.
H	Annual harvest of wood for products, which includes wood and residues removed from harvest sites, but excludes residues left at harvest sites.
HWP	Harvested wood products in use or at solid waste disposal sites.
E_R	Annual emission of carbon to the atmosphere in the Region from products made from wood harvested in the Region.
E_{IM}	Annual emission of carbon to the atmosphere in the Region from products made from wood harvested outside of the Region and imported into the Region.
P_{EX}	Annual exports of wood and paper products out of the Region, including roundwood, chips, residue, pulp and recovered (recycled) products.
P_{IM}	Annual imports of wood and paper products into the Region, including roundwood, chips, residue, pulp and recovered (recycled) products.
$E_{EX R}$	Annual emission of carbon to the atmosphere in areas outside of the Region from products made from wood harvested in the Region.
E_{OTHER}	Annual emission of carbon to the atmosphere in areas outside of the Region from products made from wood harvested outside the Region.
C_R	Stock of harvested wood products carbon in use or at solid waste disposal sites where products used wood from the Region.
$\Delta C_{IU R}$	Annual change in carbon stored in harvested wood products in use where products used wood from the Region.
$\Delta C_{SWDS R}$	Annual change in carbon stored in harvested wood products at solid waste disposal sites where products used wood from the Region.
ΔC_R	Annual change in carbon stored in harvested wood products in use and at solid waste disposal sites where products used wood from the Region.

System boundaries

Most people are familiar with imports and exports in the context of international trade, but the concept can be applied to understand the treatment of carbon imports and exports in the IPCC/EPA approach. In this case the terms export and import refer to the border of the Southwestern Region. For example, HWP manufactured in a USFS Region may be used locally by consumers inside the Region or exported from the local area for use elsewhere. Similarly, HWP produced outside the Region may be imported for use within the Region. Figure 2 shows that carbon emissions attributed to HWP from the Region (indicated with solid boxes) include both emissions to the atmosphere from wood products harvested and used within the Region (E_R) and emissions to the atmosphere from wood products harvested in the Region that were exported outside the Region ($E_{EX R}$). Emissions (E_R and $E_{EX R}$) are further categorized as emitted with energy capture (e.g. fuelwood) and emitted without energy capture (e.g. decomposition and burning for waste disposal). Exports (P_{EX}) include wood and paper products, as well as roundwood, chips, residue, pulp and recovered (recycled) products from wood harvested in the Region. Under the IPCC/EPA approach, imports from elsewhere (indicated with dotted lines around the right side of both HWP boxes) are not included in regional accounting because the emphasis is on the location of harvest (H).

Additionally, this approach does not account for all emissions associated with HWP. For example, carbon emissions from fossil fuels used in harvest, transportation and manufacture of HWP are not deducted from the HWP pool. Similarly, although HWP emissions with energy capture are quantified in the IPCC/EPA approach, they are not assumed to substitute for an equivalent amount of fossil fuel carbon, potentially reducing fossil fuel emissions in some scenarios (Jones et al. 2010). Furthermore, this approach does not incorporate carbon fluxes associated with product substitution, such as the substitution of HWP for metal or concrete (or vice versa) in building applications, and the associated land use changes that may ensue.

Though these types of emissions tradeoffs are outside the scope and purpose of the approach applied in this report, there are well-developed methods of life cycle assessment (LCA) that account for all carbon emissions associated with manufactured products and that facilitate the comparison between wood products and alternative products (Rebitzer et al. 2004). The IPCC/EPA approach provides information that can be used in an LCA, but in general an LCA is used to address different questions.

If management decisions require information about harvesting, transportation and processing emissions, product substitutions, or other trade components not included in the approach used here, a consequential LCA is appropriate. However, for sub-national carbon accounting, the IPCC/EPA approach has several benefits over LCA. It is relatively easy to apply and congruent with US national carbon accounting standards, which is particularly important in developing tools that can be used by USFS managers to meet carbon monitoring goals.

Computational Methods

Figure 3 provides a flow chart of the computational methods used to calculate annual stock changes and emissions from HWP for the IPCC/EPA production accounting approach. This approach does not apply simple storage ratios to the harvest; rather it tracks carbon through the product life cycle from harvest to timber products to primary wood products to end use to disposal, applying best estimates for product ratios and half-lives at each stage.

When possible, harvest records are used to distribute annual cut volumes among specific timber product classes (e.g., softwood ties, softwood sawlogs, softwood pulpwood, softwood poles, softwood fuel wood, softwood non-saw, etc.). For periods of time when timber product classes were not recorded, ratios available from a more recent time period were used. Timber products are further distributed to specific primary wood products (e.g. softwood lumber, softwood plywood, softwood mill residue used for non-structural panels, etc.) using default average primary product ratios from national level accounting that describe primary products output according to regional forest industry structure (Smith et al. 2006, Appendix A). Mill residues are included as primary wood products with some entering solid waste disposal immediately and some getting converted into products that rely on mill residues as raw material, such as particleboard and paper. The timber product to primary wood product ratios vary by region and in most cases the geography of the regions used in national level accounting does not match perfectly the boundaries of Forest Service administrative regions. Therefore, applying default ratios for part or all of the accounting time period requires some judgment in selecting the appropriate ratios, and the ratios for national regions are sometimes modified. Primary wood product outputs are converted from their reporting units to MgC using standard conversion factors for primary wood products (Smith et al. 2006, Table 2). The ratios from Smith et al. (2006) are applied to the entire time period, but are adjusted with consideration of the timing of manufacturing capacity in each region.

The recalcitrance of carbon in HWP is highly dependent on the end use of those products. For example, carbon in lumber used in new single family home construction has a longer duration than carbon in lumber used for shipping containers, which is released into the atmosphere more quickly through combustion and decay. For years 1950 through 2012, annual primary wood product output was distributed to specific end uses according to annual wood product consumption estimates in McKeever (2009, 2011).

Table 2. Conversion factors used in this analysis.

Conversion	Units
2.0	ccf per mbf, timber harvest prior to 2000 ¹
33 to 42	lbs per cubic foot, primary products
2204.6	lbs per Mg
0.95 to 1.0	Mg wood fiber per Mg product
0.5	Mg carbon per dry Mg wood fiber
0.711 to 0.919	MgC per ccf, primary products

For each of the 203 different possible end uses from the Region's HWP (e.g., softwood lumber/new housing/single family, softwood lumber/new housing/multifamily, softwood lumber/new housing/manufactured housing, softwood lumber/manufacturing/furniture, softwood lumber/packaging and shipping, etc.) for each vintage year, the amount of carbon remaining in use at each inventory year is calculated based on the product half-life and the number of years that have passed between the year of harvest and the inventory year. The half-life value expresses the decay rate at which carbon in the products in use category passes into the discarded category, representing the transition between the two pools. The carbon remaining in HWP in use in a given inventory year is calculated for each vintage year end use based on a standard decay formula:

$$N_t = N_0 \exp(-t \ln(2)/t_{1/2})$$

where N_t is the amount of carbon remaining in use in inventory year t , N_0 is the amount of carbon in the end use category in the vintage year of harvest, t is the number of years since harvest, $t_{1/2}$ is the half-life of carbon in that end use, and $\exp$ is notation for the exponential function. In our calculations, the starting amount (N_0 , at $n=0$) is adjusted downward by 8% to reflect a loss when placed in use, which is assumed to enter the discarded carbon category. This loss in use accounts for waste when primary products (e.g. softwood lumber) are put into specific end uses (e.g. new single family residential housing), and this waste is immediately distributed to the discarded products category. Fuelwood products are assumed to have full emissions with energy capture in the year they were produced.

For carbon of a particular vintage in a given inventory year, the balance of carbon in HWP that is not in use and not emitted with energy capture is assumed to be in the discarded products category (Figure 3). Carbon in the discarded products category is partitioned into five disposition categories: burned, recovered, composted, landfills and dumps. The proportion of discarded products that ends up in each of these five categories is different for paper and solid wood products, and has changed over time. For example, prior to 1970 wood and paper waste was generally discarded to dumps, where it was subject to higher rates of decay than in modern landfills. Since then, the proportion of discarded wood going to dumps has dropped to below 2%, while the proportion going to landfills has risen to 67%, with the remainder going to the other disposition categories (Skog 2008). Similarly, composting and recovery (i.e. recycling and reuse) have become a more prominent part of waste management systems. In 2004, approximately 50% of paper waste was recovered, compared to 17% in 1960. The disposition of carbon in paper and solid wood products to these categories is based on percentages in Skog (2008).

Carbon from burned and composted discarded products is assumed to be emitted without energy capture. Carbon in the recovered category reenters the products in use category in the year of recovery. Carbon in products discarded to landfills and dumps are subject to decay determined by their respective half-lives. The half-life value for discarded products in dumps and landfills expresses the decay rates at which carbon in these categories is emitted to the atmosphere. However, our calculations consider the fact that only a fraction of the discarded products pool in landfills is considered to be subject to decay; 77% of solid wood carbon and 44% of paper carbon in landfills is identified as fixed carbon, not subject to decay (Skog 2008). For a given vintage year, the carbon remaining in SWDS in a given inventory year is the sum of fixed carbon and the carbon remaining after decay. We do not account for the difference between methane and CO₂ emissions from landfills in terms of CO₂ equivalents, nor do we account for methane remediation that includes combustion and subsequent emissions with energy capture. All landfill and dump emissions are considered emissions without energy capture.

¹ Both mbf and ccf are available in all timber harvest reports after 2000.

These methods were used to calculate annual gross stocks and gross emissions for all inventory years 1909 through 2012. Results for each inventory year were used to calculate net change in stocks of carbon in regional HWP products in use (ΔC_{IU_R}) and SWDS (ΔC_{SWDS_R}), as well as net change in emissions from SWDS and fuelwood (E_R).

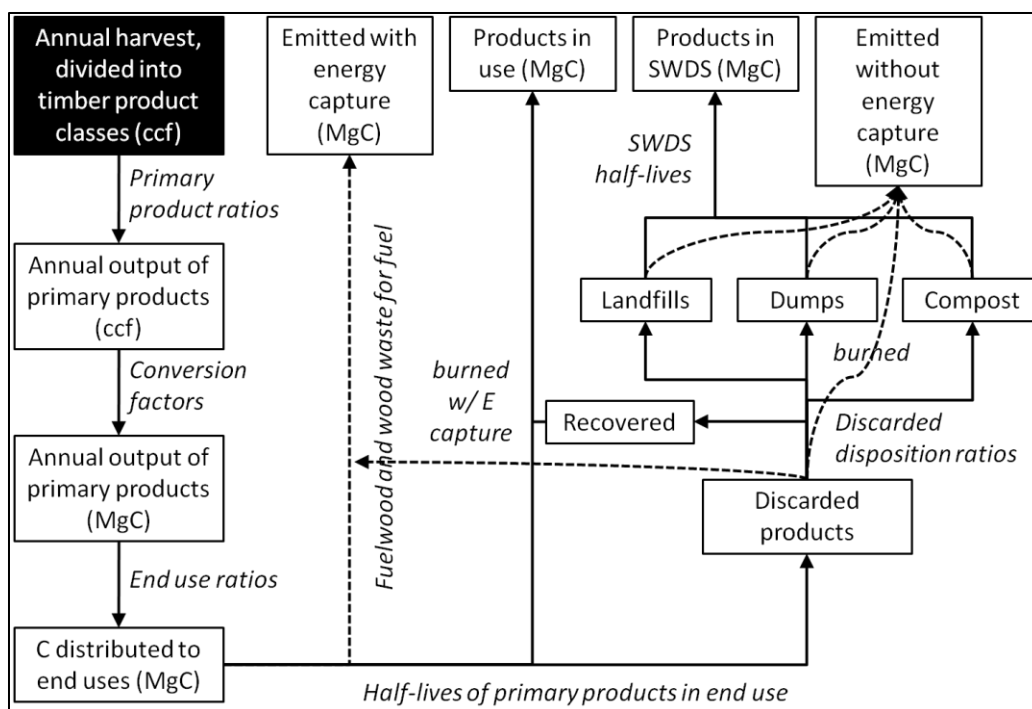


Figure 3. A schematic of calculations to quantify HWP storage and emissions. These calculations quantify HWP products in use, products in SWDS, emissions with energy capture, and emissions without energy capture using the IPCC/EPA approach.

Online Harvested Wood Products Carbon Accounting Tool

Calculations were facilitated by an online HWP carbon accounting tool developed by USFS and cooperators (USURS 2012). The tool requires two inputs: a harvest time series and a time series of timber product ratios that partition the harvest into different timber product classes, which are discussed in the following section. In addition, the user can enter primary product ratios if they are known, or use the default values from Smith et al. (2006). The option to input primary products ratios allows the user to more accurately reflect regional changes in industry structure and associated primary product manufacturing if desired. The user can also provide additional inputs to guide the Monte Carlo simulations that determine statistical confidence intervals, including random variable distributions and number of iterations, or use the default values provided. The latest version of the tool, with supporting documentation, can be found at: <http://maps.gis.usu.edu/HWP>.

Data Sources

Data quality impacts the uncertainty and reliability of our estimates, and the data used in this analysis provide a good illustration of the challenges associated with using historical data in carbon accounting. This section is divided into four parts: first we discuss historical timber harvest data acquisition and limitations, and how those limitations were addressed. Following that we describe how the data were allocated to timber products, how timber products were allocated to primary products and finally how we allocate primary products to end use products for all Regions. By standardizing boundaries and units and partitioning the harvest among different timber and primary product classes, we created a continuous dataset spanning 1909 through 2012 that meets the criteria for estimation established by the IPCC (2006).

Historical timber harvest data

Regional harvests have been reported in detailed cut-and-sold reports and are available online from 1977 to the present². These reports include the value and volume of timber sold and harvested in the region, which are reported by both fiscal and calendar year. In addition, total harvests are partitioned by sale value, timber product class³, tree species, and national forest within the Region. Records for annual harvest prior to 1977 are generally more difficult to obtain; however, annual harvest data from 1909 to present were supplied by the Southwestern Regional Office. Harvest data was recorded in fiscal years from 1909 to 1918 with missing data for the month of June in 1913, a block of 6 months in 1914, and the entire fiscal year of 1915. To fill these voids, Arizona and New Mexico state wide data from fiscal year 1913 to 1915 was recorded from USFS Reports of the Chief. From 1919 to 1976 calendar year data was used resulting in a gap in the harvest data time series from January 1, 1919 to June 30, 1919, which is not captured within this analysis. Fiscal year data is used from 1977 to 2012, resulting in an overlap of 1976 and 1977 data. To resolve this situation calendar year 1976 has had the first quarter of fiscal year 1977 removed to avoid overestimations in our analysis. In instances when parcels within current national forest boundaries were formerly within neighboring national forests, timber volumes in our dataset were likewise apportioned based on available details, but done in this fashion this intra-region apportionment had no bearing on Regional harvest totals (Davis 1983). All results in this report are based upon fiscal year harvest reports.

Because the model developed for this purpose requires cubic foot input metrics for harvested timber, conversion factors for specific timber products were used to convert volumes from thousand board feet (mbf) to hundred cubic feet (ccf) (Table 2). Beginning in 2001, harvested volumes have been reported in both mbf and ccf. Between 1909 and 2000 volumes were reported in mbf only. For this period annual harvest totals for Southwestern Region were reported in mbf and converted to ccf using a conversion factor of 2 ccf per mbf in use by the Region (Table 2). No conversion was necessary for years 2001 and later since ccf is provided in the harvest data.

There is new evidence that ccf per mbf conversion factors have changed in recent decades. For example, Keegan et al. (2010a) have found a 16% decrease in mbf per ccf conversion in California from 1970s to 2000s. This alone would suggest conversions from mbf to ccf in earlier decades overestimate the volume harvested. On the other hand, Keegan et al. (2010b) indicate that utilization represented as cubic feet of green finished lumber per cubic foot of bole wood processed has increased during the same period by roughly the same magnitude (16% in California). This would suggest that estimates of carbon in products in use were underestimated in earlier decades. Assuming that the findings by Keegan et al. essentially cancel each other out, and considering we did not have adequate timber harvest data from all National Forests across the entire period, we chose not to incorporate this information into our calculations. In addition, analyses similar to those found in Keegan et al (2010a, 2010b) are not available for all USFS Regions. To accommodate this type of unknown variability over time, we provide an uncertainty analysis in this report, which is discussed below.

Historical timber product data

Southwestern Region harvest records from 1977 to 2012 are partitioned into multiple timber product classes (i.e. sawlogs, pulplogs); however, harvest records 1906 through 1960 do not partition the harvest among different timber product classes, but report only total annual harvest in mbf. Harvested softwood pulpwood is reported from 1961 to 1976 along with total harvest volume. To estimate the proportion of total Southwestern Region harvest that went into each timber product class we first applied the average annual proportion of the harvest represented by each timber product class from 1977 through 2012 to the annual harvest for each year 1961 through 1976. Timber product ratios for years 1961 through 1976 were adjusted to accommodate for the true softwood pulpwood ratio by taking the difference of the average softwood pulpwood ratio calculated from harvest years 1977 through 2012, and the actual softwood pulpwood ratio from harvest years 1961 through 1976, and applying the difference to the other timber product ratios based on each timber product ratio's proportion of the total harvest. Lastly, we developed average timber product ratios for years 1961 to 2012 based upon the steps just described, and applied these ratios to harvest years 1906 through 1960. Timber product classes of "convertible biomass" have been eliminated from 1909

² USFS 2013 (<http://www.fs.fed.us/forestmanagement/products/sold-harvest/cut-sold.shtml>)

³ Many times the timber product classes predicted and recorded in cut sold reports are not actually the products classes that are used after harvest. This reality, in addition to the lack data for these ratios for the entire data period, explain why we include timber and primary product ratios in our uncertainty analysis.

to 1950 because these timber product classes have only been recorded since 1998 and account for less than 0.7% of the total yearly average harvest from 1951 to 2012 (Included in “Other Products” Table 3).

Table 3. The average annual proportion of 1961 to 2012 Southwestern Region harvests distributed to timber product classes between 1909 and 1960 (n=51 years).

Product class	Mean	Std. Error
Sawtimber, softwood	0.54	0.025
Fuel., softwood	0.33	0.022
Pulp, softwood	0.09	0.007
Poles, softwood	0.02	0.003
Fuel, hardwood	0.01	0.0001
Misc. softwood	0.004	0.0008
Other products	0.008	0.0007

Historical primary product data

The carbon in HWP from timber products to primary products is based upon intricate disposition connections from harvested timber products to primary products to end-uses found in Smith et al. (2006). The Smith et al. study defined a single region that combines the southern and central Rocky Mountains of which the NFS Southwestern Region is a subset. The Smith et al., Rocky Mountain South region is defined on the state borders of New Mexico, Nevada, Utah, Wyoming, Colorado and Arizona. Given no finer division to the study regions of Smith et al, we used the primary product data for the region as a whole.

Historical end use data

The historical end use data used for the Southwestern Region comes from McKeever (2009 and 2011). This national data set is used for all NFS regions for the distribution of primary products to end uses for all regions, with no regional variation. Estimates for 1950 were used for 1909 through 1949 and estimates for 2009 were used for 1950 through 2012. We acknowledge that this is not ideal, but no other data are available for these periods. The annual end use wood product estimates are periodically updated, which could allow better HWP storage and flux estimates in the future.

Uncertainty analysis

Interpretation of the results should be made in light of some constraints. Though we attempted to normalize annual harvests to the modern boundary of the Region using forest-specific harvest data, in actuality the annual harvest is from a land base that is somewhat variable over time. The USFS has commonly engaged in land exchanges, divestments and acquisitions in the Regions since their origin, which means that the geographic boundary for Regions has not been consistent. In addition, conversion factors (which depend on average log size, mill technology and efficiency, etc.), distribution of timber products to primary products, and the distribution of primary products to end uses have changed over time. Though we have used annual data whenever possible, there is some uncertainty associated with applying averages to the early years of the harvest series.

Uncertainty is quantified using the methods described in Skog (2008). We identified the most critical sources of uncertainty in our analysis (Table 4), developed probability distributions (using expected ranges) for each of four major sources of uncertainty (conversion factors, reported harvest, product distribution variables, and product decay parameters), and carried out Monte Carlo simulations to determine the collective effect of uncertainty in these variables on estimates of HWP stocks. We did not explore the contribution of each variable in a sensitivity analysis, but instead address collective uncertainty. Further investigation into the level of uncertainty of each random variable and its effect on confidence intervals could help managers determine where to focus improvements in reporting to reduce uncertainty in carbon storage and flux estimates. Across all variables, sensitivity analyses could be used to identify variables that have the greatest impact on carbon storage and flux, and compare alternative levels of those variables associated with different scenarios of forest management and HWP production, use and disposition.

Table 4. Sources of uncertainty and range of the triangular distribution for each random variable used in the Monte Carlo simulation.

Source of Uncertainty	Range of distribution	Years
Reported harvest in ccf	±30%	start to 1945
	±20%	1946 to 1979
	±15%	1980 to end
Timber product ratios	±30%	start to 1945
	±20%	1946 to 1979
	±15%	1980 to end
Primary product ratios	±30%	start to 1945
	±20%	1946 to 1979
	±15%	1980 to end
Conversion factors, ccf to MgC	±5%	all years
End use product ratios	±15%	all years
Product half lives	±15%	all years
Discarded disposition ratios (paper)	±15%	all years
Discarded disposition ratios (wood)	±15%	all years
Landfill decay limits (paper)	±15%	all years
Landfill decay limits (wood)	±15%	all years
Landfill half-lives (paper)	±15%	all years
Landfill half-lives (wood)	±15%	all years
Dump half-lives (paper)	±15%	all years
Dump half-lives (wood)	±15%	all years
Recovered half-lives (paper)	±15%	all years
Recovered half-lives (wood)	±15%	all years
Burned with energy capture ratio	±15%	all years

Because we apply different distributions to different time periods for some variables, the 23 distributions cover 17 different variables. Multiple time-delineated distributions are used for reported harvest, primary products ratios, and end use ratios, with time periods separated at benchmark years related to data quality. The probability distributions of these random variables were developed based on estimates in Skog (2008) and on professional judgment, and are assumed to be triangular and symmetric. A triangular error distribution was selected because without additional empirical information, we reasonably assume the error distribution to be symmetric with greater likelihood of values being centered in between the limits of the distribution than at one or both of the limits of the distribution. In addition, we can reasonably assign values to the limits. The distributions are assumed to be independent of one another.

The effect of uncertainty in these variables on HWP carbon storage was evaluated using Monte Carlo simulation. For each simulation, a mean value and 90% confidence intervals are the results of 3,000 iterations performed to reach a stable standard deviation in the mean (Stockmann et al. 2012). In each iteration, HWP carbon stocks are calculated using values for variables drawn at random from the established distributions. Using thousands of draws, we produce a simulation mean and a distribution of values that can be used to establish the confidence intervals

shown in the tables. These confidence intervals show the range of values in which 90% of all values are expected to fall.

Results for the Southwestern Region

From 1909 to 1929 the annual timber harvest in Southwestern Region remained below 0.3 million MgC (Table 5, Figure 4). This was followed by a decrease during the Great Depression, when in 1934 harvest levels had decreased 74% from the 1929 peak harvest. From 1934 onward harvests grew notably. World War II invigorated harvests somewhat, but harvests still had not exceeded the prior pre-Depression highs by the end of the war. Overall though, a greater erratic trend up had begun that culminated with a harvest in 1972 of 0.59 million MgC. This was followed by a decade of declines in harvests, falling to 0.28 million MgC by 1982. A sharp rise thereafter led to the peak harvest of 0.76 million MgC in 1989. Following this peak harvest declines of the early 1990s bottomed at 0.07 million MgC in 1996, a level marginally above the Region's first recorded harvest in 1909.

Table 5. Annual timber product output in the Southwestern Region for selected years using the IPCC/EPA production accounting approach. This table shows carbon removed from the ecosystem by harvesting.

Harvest year	Harvest (ccf)	Timber product output (MgC)
1910	107,310	79,473
1920	181,882	134,700
1930	303,992	225,134
1940	273,972	202,901
1950	580,636	430,014
1960	507,761	376,050
1970	790,832	585,735
1980	730,788	538,370
1990	866,456	640,557
1995	199,176	148,573
2000	130,558	96,459
2005	152,129	112,997
2006	169,369	125,814
2007	172,191	128,016
2008	180,959	134,590
2009	155,629	115,891
2010	157,556	117,344
2011	188,985	140,041
2012	222,414	164,626

The cumulative carbon stored in the Southwestern Region HWP peaked in 1994 at approximately 9.7 million MgC (Appendix A). For reference, this is equivalent to 35.6 million MgCO₂ or the CO₂ equivalent annual emissions from 6.8 million passenger vehicles, 82.7 million barrels of oil, or the CO₂ equivalent emissions from 185,727 railcars of coal. Since 1994, carbon stocks in the HWP pool for the Southwest Region have been in a fairly slow decline and the 2012 HWP pool is estimated to be around 8.8 million MgC (Figure 5, Table 6).

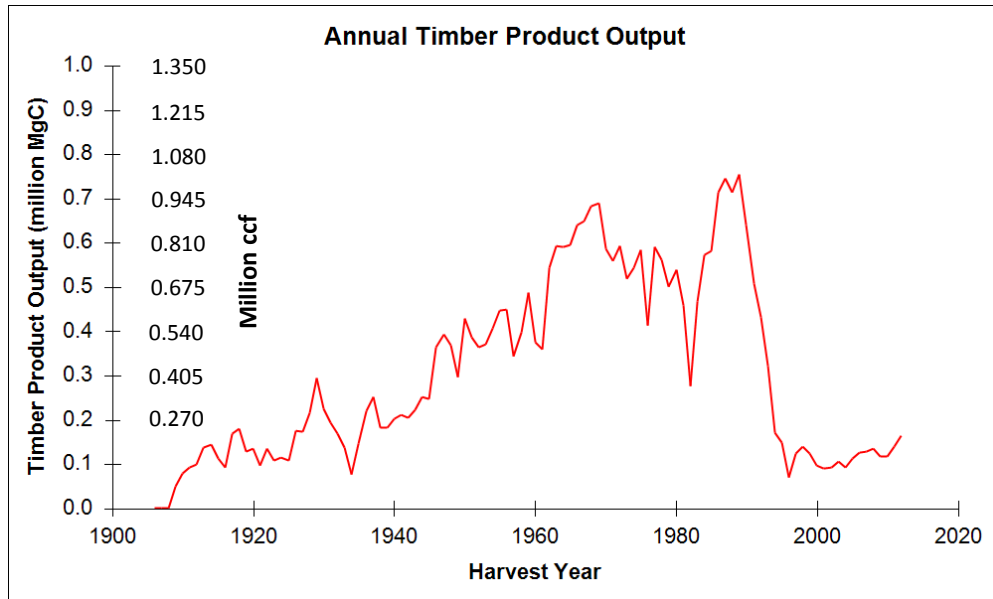


Figure 4. Annual timber product output in the Southwestern Region, 1909 to 2012. Harvest estimates are based on data collected from USDA Forest Service Archives and Cut/Sold reports.

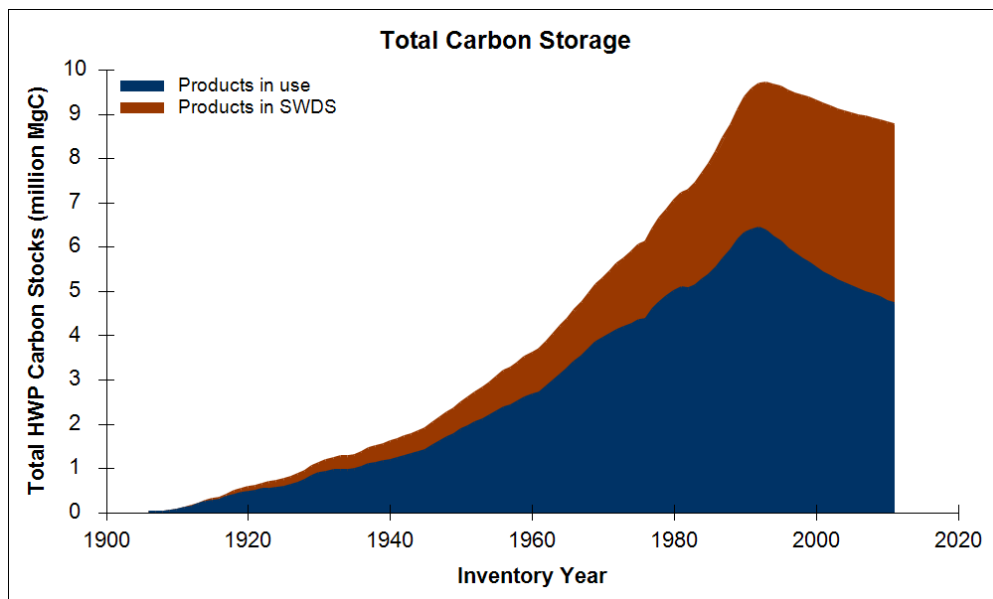


Figure 5. Cumulative total carbon stored in HWP manufactured from Southwestern Region timber using the IPCC/EPA approach. Carbon in HWP includes both products that are still in use and carbon stored at solid waste disposal sites (SWDS), including landfills and dumps.

All else being equal, higher harvest levels result in more carbon removed from the ecosystem pool and added to the HWP pool (Figure 2). Figure 5 shows the cumulative carbon in both products in use and SWDS components of the HWP pool for the Region. Based on the years that match the most recent EPA report (US EPA 2012), Table 6 shows how the disposition of HWP carbon is broken into the four IPCC/EPA categories: emitted with energy capture, emitted without energy capture, products in use and products in SWDS. For each inventory year shown in the first column, the second column shows aggregate carbon emitted with energy capture (i.e. fuelwood), the third column shows aggregate carbon emitted through decay or combustion from SWDS, and the fourth and fifth columns show carbon stored in products in use and products in SWDS, respectively. The final column, the “Total in HWP pool,” is the sum of products in use and carbon in SWDS. Note that the estimate for each inventory year includes the portion of HWP carbon still in use and in SWDS for all previous harvest years back to 1909 in addition to carbon harvested

in the inventory year. Some of the cumulative emissions from the burned and decayed HWP (Table 6, second and third columns) are theoretically taken out of the atmosphere by regrowth on harvested sites, but this effect is accounted for in the ecosystem carbon component (NEE) of the change in carbon stock equation, not in the HWP component (H and ΔC_R).

Table 6. Cumulative disposition of Southwestern Region HWP carbon for selected years using the IPCC/EPA production accounting approach. This table shows the fate of all carbon removed from the ecosystem by harvesting.

(1) Inventory year	(2) Emitted with energy capture	(3) Emitted without energy capture	(4) Products in use	(5) SWDS	(6) Total in HWP Pool ^a
(MgC)					
1910	26,142	534	21,171	-	21,171
1920	681,818	70,866	426,603	82,597	509,201
1930	1,509,076	261,732	813,966	219,622	1,033,588
1940	2,455,422	587,941	1,140,362	394,033	1,534,395
1950	3,923,151	1,056,298	1,747,500	589,981	2,337,481
1960	6,095,508	1,766,799	2,592,077	921,925	3,514,002
1970	9,133,246	2,790,972	3,831,093	1,306,857	5,137,950
1980	11,703,910	3,944,087	4,864,827	1,962,682	6,827,508
1990	14,126,826	5,008,744	6,149,702	2,953,613	9,103,315
1995	15,017,372	5,686,840	6,228,105	3,433,190	9,661,294
2000	15,383,435	6,276,378	5,635,460	3,723,468	9,358,928
2005	15,699,055	6,781,949	5,160,842	3,881,248	9,042,090
2006	15,768,690	6,873,252	5,090,609	3,905,972	8,996,582
2007	15,843,181	6,961,804	5,030,706	3,930,357	8,961,064
2008	15,921,099	7,047,735	4,972,184	3,954,838	8,927,022
2009	16,005,806	7,131,285	4,915,745	3,979,156	8,894,901
2010	16,093,577	7,212,245	4,841,432	4,003,473	8,844,905
2011	16,184,203	7,290,698	4,769,197	4,026,755	8,795,952
2012	16,281,608	7,366,985	4,714,472	4,049,256	8,763,728
2013	16,384,447	7,441,547	4,679,120	4,071,897	8,751,017

^a Sum of Products in use and SWDS.

Figure 6 and Table 7 present the trend in terms of net annual change in HWP carbon stocks. Negative net annual change in HWP carbon stocks values means that total carbon stored in the HWP pool in the inventory year is lower than in the previous year. In other words, a decline in the HWP pool results in a transition from a positive net annual change in carbon stocks to a negative net annual change in carbon stocks. While there seems to be only low and erratic growth in carbon stock additions between the forest's first recorded harvest and the mid-1930s, carbon stock net change began to rise in 1935 with the addition of 6,181 MgC until reaching peak additions in 1990 with a positive flux in carbon stocks of just over 335,000 MgC. In the mid-1990's, the flux changed from positive to negative and since then the Southwestern HWP pool has become a net source of atmospheric carbon with the largest negative net annual change in 2004 when stocks decreased by over 57,000 MgC. However, in most years since 2004 additions to the HWP C through new harvest have grown faster than emissions from the HWP pool. Recall that these estimates relate only to HWP and do not quantify carbon fluxes in the ecosystem pool.

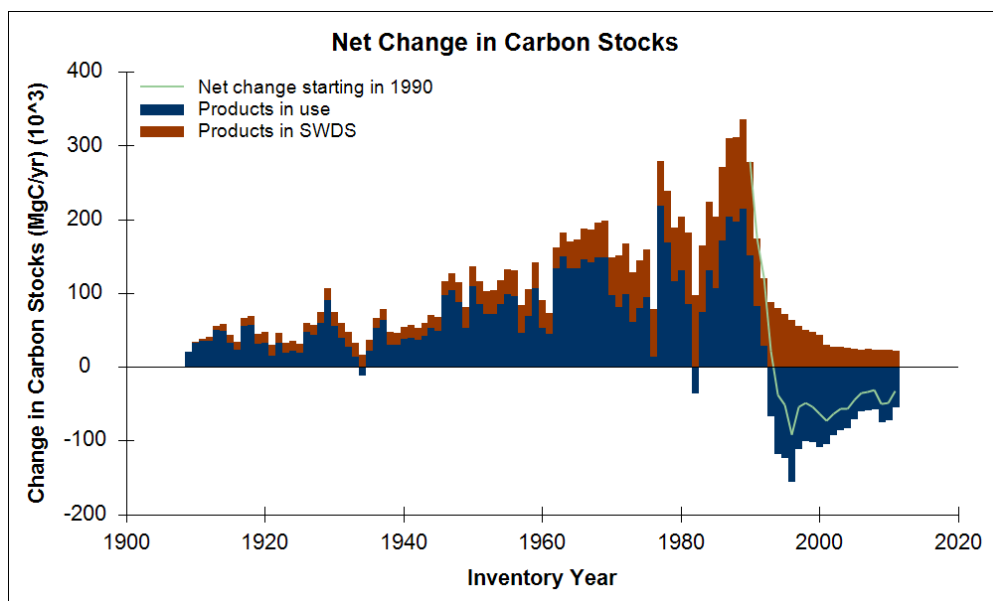


Figure 6. The net change in carbon stocks in HWP from the previous year using the IPCC/EPA production accounting approach. The net stock change is the sum of net change for SWDS and products in use. The total net change trend line shows a transition from net additions to carbon stocks in HWP to a period of net loss in HWP.

Table 7. Annual net change in HWP carbon stocks for selected years for harvests.

Inventory Year	Stock change ^a (MgC yr ⁻¹)
1910	21,171
1920	45,749
1930	107,475
1940	46,818
1950	81,170
1960	142,230
1970	198,898
1980	189,535
1990	335,716
1995	-37,535
2000	-54,459
2005	-57,120
2006	-45,508
2007	-35,518
2008	-34,041
2009	-32,121
2010	-49,996
2011	-48,953
2012	-32,224
2013	-12,711

^aNet annual change in C in products in use and SWDS.

To quantify uncertainty, confidence intervals were estimated for HWP stock estimates using Monte Carlo simulation, representing 18 random variable distributions, with distributions determined from publications and expert opinion. Table 8 shows the resulting confidence intervals for the IPCC/EPA estimates for selected years. For 1995, the year of peak carbon stocks in Table 8, the 90% confidence interval ranges from 9,651,653 MgC to 9,662,518 MgC, with a mean value of 9,657,086 MgC. This is equivalent to a $\pm 0.06\%$ difference from the mean.

Table 8. Confidence intervals for cumulative carbon in HWP for selected years for harvests beginning in 1910 using the IPCC/EPA production accounting approach. Means and confidence intervals were calculated using Monte Carlo simulation (3,000 iterations).

Inventory year	Simulation Mean (MgC)	90% Confidence interval	
		Lower limit (MgC)	Upper limit (MgC)
1910	21,187	21,096	21,278
1920	509,097	508,388	509,806
1930	1,033,653	1,032,510	1,034,795
1940	1,532,666	1,531,190	1,534,141
1950	2,335,473	2,333,542	2,337,405
1960	3,510,234	3,507,682	3,512,786
1970	5,134,394	5,130,898	5,137,889
1980	6,822,111	6,817,785	6,826,437
1990	9,098,267	9,093,057	9,103,477
1995	9,657,086	9,651,653	9,662,518
2000	9,354,729	9,349,408	9,360,050
2005	9,038,014	9,032,785	9,043,243
2006	8,992,556	8,987,340	8,997,773
2007	8,957,097	8,951,887	8,962,307
2008	8,923,121	8,917,915	8,928,327
2009	8,890,889	8,885,687	8,896,090
2010	8,840,917	8,835,719	8,846,114
2011	8,791,957	8,786,763	8,797,151
2012	8,759,693	8,754,495	8,764,890
2013	8,746,871	8,741,666	8,752,076

Discussion of Regional-level Estimates

National context

Although these results rely on numerous calculations, the time series of annual harvest volume (Figure 4) is at the root of the trends in carbon stocks and flux for the regional HWP pool. Several recent publications help put these HWP carbon estimates in the context of the total forest carbon, including both ecosystem carbon and HWP carbon (Heath et al. 2011, US EPA 2012). By dividing the 2001 HWP stock estimate of 9.3 teragrams of carbon (TgC) presented in Appendix B by the sum of this stock estimate and Heath et al.'s (2011) estimated 2001⁴ Southwestern Region ecosystem carbon stock of 641 TgC, we estimate that the Southwestern Region HWP carbon stocks represent roughly 1.4% of total forest carbon storage associated with National Forests in the Southwestern Region as of year 2000. At the national level, based on the EPA's total US HWP year 2001 stock estimate of 2,244 TgC (US EPA 2012), the Southwestern Region HWP carbon stocks represented 0.4% of total US HWP carbon stocks.

⁴ Mean measurement year reported as 2000.8.

Estimates of forest ecosystem flux in the western US exist (Healey et al. 2009, Heath et al. 2011, Van Deusen and Heath 2007) and others in development. However, long-term data collection requirements will delay reporting until the USFS Forest Inventory and Analysis Program completes its second cycle of plot measurements. However, our calculations of HWP carbon flux will allow the Southwestern Region to reasonably account for carbon that was harvested from National Forests over the study period. Ideally, when changes in forest ecosystem carbon are quantified in subsequent research they can be linked with the HWP estimates described here.

Applications of this approach by forest managers

The methods presented here for estimating the HWP carbon pool will allow resource managers and the public to develop a more complete understanding of the dynamics of HWP as a component of total forest carbon pool, and may allow the evaluation of the effect of alternative harvesting intensities on carbon stocks and fluxes. Furthermore, a benefit may be realized by evaluating the feasibility, utility, uncertainty, and limitations of the metrics and estimation methods that could be used to meet carbon monitoring objectives.

The IPCC/EPA approach requires harvest information for many prior years to make an estimate of net change to carbon stocks each inventory year over time. We recommend that all applications of the IPCC/EPA approach consider the quality of the data and adjust their uncertainty analysis accordingly, particularly with regards to the distributions of random variables (e.g., Table 4). However, though carbon of older vintages may be associated with higher uncertainty, it is also likely to have a smaller impact on current stocks and fluxes than more recent harvests. For example, the importance of the early harvests for the Northern Region – which spans northern Idaho, Montana, North Dakota, South Dakota, and eastern Washington – was estimated by Stockmann et al. (2012) by quantifying the portion of the current HWP pool that is attributable to carbon harvested prior to 1950. In 1950 the Northern Region HWP carbon pool was 4.5 million MgC. By inventory year 2010, only 1.7 million MgC of the carbon harvested before 1950 remained in products in use and SWDS, which accounted for 6.6% of the total stocks of 25.8 million MgC in 2010. Although we do not provide a similar estimate for the Southwestern Region, we believe the same trend is likely to hold for most regions. This small contribution to current stocks is a result of two factors. First, there was greater harvesting activity for the period after than before 1950. Second, following the passage of the Resource Conservation and Recovery Act of 1976 (RCRA, 42 USC 6901) and after a short lag, a much larger portion of discarded HWP goes into modern landfills where it is subject to lower rates of decay than in aerobic dumps or disposal by open burning, which were the dominant disposal methods prior to RCRA.

Obtaining historical information may present a challenge for some National Forests. It may be particularly difficult to reconstruct harvest data prior to the mid-1940s, though regression of trends after the period might be appropriate for extrapolation to earlier periods. Alternatively, regions could base their carbon accounting on national level parameters, making the assumption that national-level numbers are adequate for regional and sub-regional analysis. If national level values represent the best available data, the IPCC/EPA method requires only harvest volume information from the user. Many regional and forest type-specific default dynamics and decay functions are supplied by national level efforts (Skog 2008, Smith et al. 2006). The simplicity associated with using national data in calculations may make the system functional and effective in meeting monitoring needs for forest managers both within and outside the USFS, regardless of data quality. If superior information exists for smaller scale units, it may be possible to substitute these ratios and conversion factors into the modeling effort. However, one needs to be mindful that the results of tailored analyses might not match up with results across the country and NFS. This could be a source of interesting future research.

We successfully applied the methods described by Skog (2008) to estimate the uncertainty associated with our HWP carbon stock estimates (Table 8). However, it is unclear how the magnitude of this uncertainty would change, if at all, if the analysis were done on smaller management units (e.g. the individual National Forest level). The change in uncertainty would, in large part, depend on assumptions made about the distributions of random variables used in the analysis. In some cases, a regional analysis may be sufficient to inform forest-level land management planning, forest management practices, and planning of long-term (programmatic) timber harvest levels and associated effects on carbon flux. A detailed sub-regional analysis may be needed where there are significant within-region differences in ecosystems and disturbance processes and harvest levels.

Conclusions

HWP is an important carbon pool that should be considered in decision making associated with carbon monitoring and climate change adaptation and mitigation. However, as $\Delta S = (NEE - H) + (\Delta C_R)$ shows, total forest carbon is a function of both HWP and ecosystem carbon, which may have increased over the study period. This report fits into a larger effort to address this entire system, the Forest Carbon Management Framework, which is currently under development. Together with accounting and modeling methods that quantify ecosystem forest carbon, the approaches used in this study provide a powerful tool to monitor carbon stocks, stock change, as well as the ability to assess the possible outcomes of management actions intended to reduce the vulnerability of forest resources to climate change.

Though our analysis is at the Regional level, we provide a framework by which the IPCC/EPA method can be applied broadly at other administrative units and forests to estimate harvest (H) and the resulting change in HWP carbon stocks for the region (ΔC_R). We estimated ΔC_R each year by summing our estimates for the change of carbon stored in products in use from wood harvested in the region ($\Delta C_{IU\ R}$) and the change of carbon stored in solid waste disposal systems from wood harvested in the region ($\Delta C_{SWDS\ R}$). Although we did not have access to detailed recent information about wood harvest in agency cut-and-sold reports, we were fortunate to have archived historic harvest volume records. As expected, records for the partitioning of the harvest to timber and primary product classes improved markedly as our records approached the present time. Although we applied timber product distributions, primary product distributions, and end use product distributions from the more recent years to earlier years of harvest and we made adjustments to primary product distributions to reflect the manufacturing onset for several primary product classes based on historical information, in general we had a strong set of historical data to use in our calculations.

The Southwestern Region HWP pool is now in a period of negative net annual stock because the decay of products harvested between 1909 and 2012 exceeds additions of carbon to the HWP pool through harvest (Tables 6 and 7). The IPCC/EPA production accounting approach is data intensive because it includes past harvest and product disposition data for each inventory year, but it provides estimates of total stocks and stock change making it congruent with national accounting and reporting protocols.

The IPCC/EPA approach could be used to predict changes to the HWP component of the forest carbon pool resulting from planned or potential change in the amount of wood harvested. Quantifying uncertainty is an important component regardless of the analytical approach used because it quantifies the confidence we have in estimates of carbon stocks. We believe further research is necessary to help policy makers and managers better understand the implications of alternative forest management strategies on forest carbon stocks and stock change. An integrated approach might include consequential LCA that evaluates changes in harvest activity on carbon emissions including all sources of emissions and product substitutions.

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Appendix A

Distribution of timber products to primary wood products for regions of the US (Smith et al. 2006).

Table D6.—Fraction of each classification of industrial roundwood according to category as allocated to primary wood products (based on data from 2002)^a

Region	Category ^b		Softwood lumber	Hardwood lumber	Softwood plywood	Hardwood plywood ^c	Oriented strandboard	Non- structural panels	Other industrial products	Wood pulp	Fuel and other emissions
	SW/HW	SL/PW									
Northeast	SW	SL	0.391	0	0.004	0	0	0.020	0.083	0.072	0.431
		PW	0	0	0	0	0.010	0.016	0	0.487	0.487
	HW	SL	0	0.492	0	0.005	0	0.022	0.038	0.058	0.386
		PW	0	0	0	0	0.293	0.007	0	0.350	0.350
North Central	SW	SL	0.378	0	0	0	0	0.049	0.120	0.084	0.370
		PW	0	0	0	0	0.020	0.009	0	0.486	0.486
	HW	SL	0	0.458	0	0.006	0	0.013	0.044	0.064	0.415
		PW	0	0	0	0	0.361	0.009	0	0.315	0.315
Pacific Northwest, East	SW	All	0.422	0	0.069	0	0	0.001	0.001	0.144	0.363
Pacific Northwest, West	SW	SL	0.455	0	0.089	0	0	0.009	0.073	0.114	0.260
		PW	0	0	0	0	0	0	0	0.500	0.500
Pacific Southwest	HW	All	0	0.160	0	0.140	0	0.002	0	0.229	0.469
		All	0.454	0	0	0	0	0.040	0.036	0.145	0.325
Rocky Mountain	SW	All	0.402	0	0.054	0	0	0.033	0.062	0.153	0.296
		SL	0.350	0	0.076	0	0	0.027	0.054	0.129	0.364
Southeast	SW	PW	0	0	0	0	0.103	0.004	0	0.447	0.447
		SL	0	0.455	0	0.006	0	0.049	0.012	0.087	0.391
	HW	PW	0	0	0	0	0.180	0.002	0	0.409	0.409
		SL	0.324	0	0.130	0	0	0.019	0.023	0.133	0.371
South Central	SW	PW	0	0	0	0	0.135	0.006	0	0.430	0.430
		SL	0	0.434	0	0.023	0	0.025	0.003	0.102	0.413
	HW	PW	0	0	0	0	0.160	0.001	0	0.419	0.419
West ^d	HW	All	0	0.039	0	0.301	0	0.015	0.066	0.147	0.432

^aData based on Adams and others (2006).

^bSW/HW=Softwood/Hardwood, SL/PW=Saw log/Pulpwood. Saw log includes veneer logs.

^cHardwood plywood fractions are pooled with nonstructural panels when allocating roundwood to the primary products listed in Tables 8 and 9.

^dWest includes hardwoods in Pacific Northwest, East; Pacific Southwest; Rocky Mountain, North; and Rocky Mountain, South.

Appendix B

Disposition of Southwestern Region HWP carbon for all years. This table shows the fate of all carbon removed from the ecosystem by harvesting.

Inventory year	Emitted with energy capture (MgC)	Emitted without energy capture (MgC)	Products in use (MgC)	SWDS (MgC)	TOTAL in HWP Pool (MgC)
1910	26,142	534	21,171	-	21,171
1911	68,408	1,939	53,752	1,243	54,995
1912	117,621	4,358	89,565	4,286	93,851
1913	170,461	7,852	125,867	9,084	134,951
1914	243,262	12,812	176,031	15,412	191,443
1915	319,506	19,292	225,757	23,859	249,616
1916	378,808	26,917	258,796	34,138	292,934
1917	427,989	35,439	282,155	45,002	327,158
1918	517,990	45,655	337,883	55,754	393,638
1919	613,892	57,667	395,152	68,300	463,452
1920	681,818	70,866	426,603	82,597	509,201
1921	753,456	85,276	460,088	96,887	556,976
1922	804,905	100,446	476,065	111,218	587,283
1923	876,992	116,752	508,954	124,589	633,543
1924	934,409	133,868	528,726	138,099	666,825
1925	995,942	151,841	551,634	151,038	702,673
1926	1,053,612	170,564	570,900	163,644	734,545
1927	1,146,388	190,723	618,332	175,778	794,110
1928	1,238,042	212,283	662,344	189,161	851,506
1929	1,352,343	235,672	722,601	203,511	926,112
1930	1,509,076	261,732	813,966	219,622	1,033,588
1931	1,628,811	289,678	869,881	239,116	1,108,997
1932	1,731,517	319,098	909,658	259,513	1,169,171
1933	1,820,726	349,657	937,484	279,608	1,217,092
1934	1,893,268	380,962	951,675	298,704	1,250,380
1935	1,933,486	412,306	940,498	316,131	1,256,629
1936	2,011,675	444,419	962,669	330,622	1,293,291
1937	2,128,325	478,075	1,015,668	344,500	1,360,168
1938	2,261,927	513,611	1,079,634	359,798	1,439,432
1939	2,358,939	550,253	1,110,531	377,046	1,487,577
1940	2,455,422	587,941	1,140,362	394,033	1,534,395
1941	2,563,332	627,438	1,178,852	410,608	1,589,460
1942	2,675,278	668,148	1,219,305	426,875	1,646,180
1943	2,784,162	709,978	1,255,848	443,501	1,699,349
1944	2,902,593	753,096	1,299,035	460,216	1,759,251

Inventory year	Emitted with energy capture (MgC)	Emitted without energy capture (MgC)	Products in use (MgC)	SWDS (MgC)	TOTAL in HWP Pool (MgC)
1945	3,036,224	797,794	1,352,935	477,383	1,830,318
1946	3,167,301	843,987	1,402,389	495,583	1,897,971
1947	3,360,593	892,946	1,500,311	514,443	2,014,755
1948	3,569,148	944,964	1,604,974	536,663	2,141,636
1949	3,765,511	999,714	1,694,004	562,306	2,256,311
1950	3,923,151	1,056,298	1,747,500	589,981	2,337,481
1951	4,151,848	1,116,137	1,857,335	617,277	2,474,612
1952	4,357,717	1,178,545	1,943,031	647,569	2,590,600
1953	4,551,388	1,243,255	2,015,117	679,259	2,694,376
1954	4,748,806	1,310,265	2,087,706	711,493	2,799,199
1955	4,965,310	1,379,913	2,173,198	744,217	2,917,415
1956	5,202,819	1,452,495	2,272,415	778,173	3,050,588
1957	5,441,803	1,527,935	2,368,681	813,884	3,182,565
1958	5,624,636	1,605,068	2,415,384	850,920	3,266,304
1959	5,836,127	1,684,384	2,485,295	886,477	3,371,772
1960	6,095,508	1,766,799	2,592,077	921,925	3,514,002
1961	6,295,501	1,852,226	2,645,970	959,449	3,605,419
1962	6,493,844	1,939,758	2,690,752	988,776	3,679,528
1963	6,778,045	2,031,321	2,825,385	1,016,987	3,842,372
1964	7,087,534	2,127,385	2,975,460	1,049,012	4,024,473
1965	7,403,001	2,227,519	3,109,757	1,085,319	4,195,075
1966	7,717,552	2,331,634	3,243,520	1,124,435	4,367,954
1967	8,056,732	2,440,026	3,389,645	1,165,975	4,555,620
1968	8,401,828	2,552,707	3,532,236	1,210,381	4,742,617
1969	8,767,582	2,669,724	3,681,669	1,257,383	4,939,052
1970	9,133,246	2,790,972	3,831,093	1,306,857	5,137,950
1971	9,444,629	2,897,279	3,928,632	1,358,336	5,286,968
1972	9,742,452	3,005,731	4,010,513	1,427,953	5,438,466
1973	10,056,335	3,116,382	4,109,548	1,496,396	5,605,944
1974	10,335,070	3,228,429	4,170,371	1,564,535	5,734,905
1975	10,620,454	3,342,139	4,250,157	1,629,886	5,880,043
1976	10,927,622	3,457,873	4,345,554	1,694,278	6,039,832
1977	11,148,618	3,574,146	4,359,447	1,759,411	6,118,858
1978	11,329,346	3,694,463	4,578,887	1,819,997	6,398,885
1979	11,521,962	3,817,965	4,748,394	1,889,579	6,637,973
1980	11,703,910	3,944,087	4,864,827	1,962,682	6,827,508
1981	11,905,500	4,048,411	4,996,147	2,035,819	7,031,967
1982	12,075,351	4,153,605	5,081,194	2,133,241	7,214,435

Inventory year	Emitted with energy capture (MgC)	Emitted without energy capture (MgC)	Products in use (MgC)	SWDS (MgC)	TOTAL in HWP Pool (MgC)
1983	12,192,353	4,258,572	5,046,055	2,231,657	7,277,712
1984	12,388,424	4,363,713	5,121,004	2,322,487	7,443,491
1985	12,625,925	4,469,107	5,252,926	2,415,177	7,668,103
1986	12,894,838	4,574,655	5,359,960	2,512,969	7,872,929
1987	13,223,918	4,681,006	5,531,779	2,612,883	8,144,662
1988	13,541,567	4,788,576	5,736,300	2,719,261	8,455,561
1989	13,827,057	4,897,689	5,934,401	2,833,199	8,767,600
1990	14,126,826	5,008,744	6,149,702	2,953,613	9,103,315
1991	14,373,119	5,143,582	6,300,741	3,080,953	9,381,694
1992	14,573,111	5,280,737	6,383,233	3,173,179	9,556,412
1993	14,751,604	5,418,499	6,412,860	3,264,289	9,677,149
1994	14,927,415	5,554,414	6,346,245	3,352,584	9,698,829
1995	15,017,372	5,686,840	6,228,105	3,433,190	9,661,294
1996	15,102,469	5,814,933	6,104,406	3,504,885	9,609,291
1997	15,154,029	5,937,654	5,948,562	3,568,990	9,517,552
1998	15,223,510	6,055,399	5,838,246	3,624,625	9,462,871
1999	15,306,179	6,168,256	5,737,576	3,675,811	9,413,387
2000	15,383,435	6,276,378	5,635,460	3,723,468	9,358,928
2001	15,447,235	6,384,892	5,526,573	3,768,000	9,294,573
2002	15,510,980	6,489,732	5,421,927	3,799,113	9,221,040
2003	15,571,801	6,589,833	5,329,253	3,827,345	9,156,598
2004	15,641,421	6,687,586	5,243,754	3,855,456	9,099,210
2005	15,699,055	6,781,949	5,160,842	3,881,248	9,042,090
2006	15,768,690	6,873,252	5,090,609	3,905,972	8,996,582
2007	15,843,181	6,961,804	5,030,706	3,930,357	8,961,064
2008	15,921,099	7,047,735	4,972,184	3,954,838	8,927,022
2009	16,005,806	7,131,285	4,915,745	3,979,156	8,894,901
2010	16,093,577	7,212,245	4,841,432	4,003,473	8,844,905
2011	16,184,203	7,290,698	4,769,197	4,026,755	8,795,952
2012	16,281,608	7,366,985	4,714,472	4,049,256	8,763,728
2013	16,384,447	7,441,547	4,679,120	4,071,897	8,751,017