

Shifting stoichiometry: Long-term trends in stream-dissolved organic matter reveal altered C:N ratios due to history of atmospheric acid deposition

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Abstract

Dissolved organic carbon (DOC) and nitrogen (DON) are important energy and nutrient sources for aquatic ecosystems. In many northern temperate, freshwater systems DOC has increased in the past 50 years. Less is known about how changes in DOC may vary across latitudes, and whether changes in DON track those of DOC. Here, we present long-term DOC and DON data from 74 streams distributed across seven sites in biomes ranging from the tropics to northern boreal forests with varying histories of atmospheric acid deposition. For each stream, we examined the temporal trends of DOC and DON concentrations and DOC:DON molar ratios. While some sites displayed consistent positive or

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negative trends in stream DOC and DON concentrations, changes in direction or magnitude were inconsistent at regional or local scales. DON trends did not always track those of DOC, though DOC:DON ratios increased over time for ~30% of streams. Our results indicate that the dissolved organic matter (DOM) pool is experiencing fundamental changes due to the recovery from atmospheric acid deposition. Changes in DOC:DON stoichiometry point to a shifting energy-nutrient balance in many aquatic ecosystems. Sustained changes in the character of DOM can have major implications for stream metabolism, biogeochemical processes, food webs, and drinking water quality (including disinfection by-products). Understanding regional and global variation in DOC and DON concentrations is important for developing realistic models and watershed management protocols to effectively target mitigation efforts aimed at bringing DOM flux and nutrient enrichment under control.

KEYWORDS

atmospheric acid deposition, C:N stoichiometry, dissolved organic carbon, dissolved organic matter, dissolved organic nitrogen, long-term trends, streams

1 | INTRODUCTION

Dissolved organic matter (DOM) provides an essential energy and nutrient source to aquatic ecosystems (Webster & Meyer, 1997). DOM varies in availability to biota along the hydrologic continuum (McArthur et al., 1985) and its composition and properties are closely linked to the surrounding landscape (Jaffé et al., 2008; Mattsson et al., 2005; Wymore et al., 2021c; Yates et al., 2019). The DOM pool is a complex mixture of organic compounds mostly composed of dissolved organic carbon (DOC) and dissolved organic nitrogen (DON) to a minor extent (Pagano et al., 2014). Numerous studies have confirmed that increases in DOC concentrations in north temperate freshwater ecosystems have occurred over time. For example, DOC concentrations have increased between 50% and 91% in streams and lakes of northern and central Europe, the United Kingdom, and eastern North America since the 1980s (Couture et al., 2012; de Wit et al., 2016; Driscoll et al., 2003; Evans et al., 2005; Gavin et al., 2018; Hall et al., 2021; Lawrence et al., 2011; Monteith et al., 2007; Worrall et al., 2004). Increased DOC concentration is often attributed to the recovery from acid deposition after the implementation of the Clean Air Act in the United States and similar legislation in Europe (Driscoll et al., 2003). The leading hypothesized mechanism is that a decrease in ionic strength and protonation in soil water following recovery from acid deposition leads to increases in solubility and the mobilization of DOC to adjacent water bodies (Borken et al., 2011; De Wit et al., 2007; Evans et al., 2005; Hruška et al., 2009; Lawrence & Roy, 2021).

A suite of different hypotheses has been put forward to explain the increasing trends in DOC concentrations, in addition to declines in atmospheric deposition, each associated with global change. Mechanisms include increasing precipitation and runoff (De Wit et al., 2007; Strååt et al., 2018; Worrall et al., 2004), rising CO₂ and

increased primary productivity (Freeman et al., 2004), enhanced microbial organic matter decomposition due to increased temperatures (De Wit et al., 2007; Finlay et al., 2006; Worrall et al., 2004), and permafrost thaw in northern high-latitude ecosystems (Frey & McClelland, 2009; Frey et al., 2007; Frey & Smith, 2005; Larouche et al., 2015).

DOC concentrations are not increasing everywhere, however. Declines in DOC concentration over time have been associated with decreasing soil organic matter solubility (Clair et al., 2008), declines in carbon inputs from upstream acidified lakes (Schindler et al., 1997), increases in soil aluminum pools (Löfgren et al., 2010), and greater adsorption of DOM to the mineral layer and infiltration of DOM deeper into permafrost soils (Kendrick et al., 2018; Striegl et al., 2005). Long-term stability in stream DOC concentrations has also been observed (Chow et al., 2017; Clair et al., 2008; Monteith et al., 2007; Worrall et al., 2004), even in some of the longest existing records of stream chemistry (e.g., since 1975; Räike et al., 2012). Trends in total organic carbon concentrations have even varied in direction within a continuous 35-year record (Erlandsson et al., 2008; Lepistö et al., 2008). Despite the evidence that a wide range of changes in DOC concentration can be expected, a broad multi-biome assessment of global DOC trends is lacking. A spatially distributed analysis would allow for the examination of trends along multiple environmental gradients and for the testing of coherent cross biome patterns (e.g., Dodds et al., 2019).

Concentrations of dissolved organic nitrogen (DON) are rarely measured in long-term studies of DOM. Changes in DON concentration can have critical implications for freshwater ecosystems, especially when DON serves as a primary source of N for biota (Kissman et al., 2017; Mackay et al., 2020). While analytical challenges exist in the assessment of DON, researchers often assume that the concentrations of DON track those of DOC (i.e., concentrations are

positively correlated; Campbell et al., 2000; Goodale et al., 2000; Kortelainen et al., 2006; Lepistö et al., 2008; Mann et al., 2012). Other lines of evidence, however, suggest that concentrations of DOC and DON can respond differently to environmental change such as changes in the concentrations of inorganic nutrients (Lutz et al., 2012; Wymore et al., 2015, 2021c; Yates et al., 2019) and seasonal variability in precipitation and stream runoff (Bernal et al., 2005). Recent evidence has pointed to the stream DOC:DON ratio varying according to the extent of nutrient enrichment in catchments, diverging from the soil DOC:DON ratio as systems become more nutrient-enriched through land-use change and increasing human population density (Yates et al., 2019). Such divergent trends in DOC and DON concentrations will lead to changes in DOM stoichiometry (i.e., DOC:DON ratios). DOC:DON ratios provide a relatively simple quantification of bulk DOM characteristics, which serves as an indicator of bioavailability (del Giorgio & Cole, 1998) and of changing OM sources within catchments (Yates et al., 2019). A broad assessment of how DOM stoichiometry changes concurrently with changes in concentrations of DOC and DON could provide insights into how the energy and nutrient balance of one of the larger pools of organic matter in freshwater ecosystems is changing with potential impacts on other biogeochemical reactions (e.g., Strauss & Lamberti, 2002; Wymore et al., 2019).

The objective of this study was to explore long-term trends in DOC and DON concentrations, and DOM stoichiometry in streams and rivers across biomes of the Northern Hemisphere. Our overarching hypothesis is that changes in concentrations of DON will track those of DOC and consequently the stoichiometry of DOM will remain consistent through time (Brookshire et al., 2007; Wymore et al., 2021c). We also hypothesize that sites historically affected by acid deposition will be associated with increases in concentrations of DOC and DON assuming the same external forces are acting on each of these components of the DOM pool (Deininger et al., 2020). A global assessment of how riverine DOM is responding to global change is essential for robust regional and global scale predictive ecosystem models and for future watershed management protocols.

2 | MATERIALS AND METHODS

2.1 | Data set compilation

We compiled long-term data on DOC and DON concentrations for 74 individual streams from 7 different sites (Table 1; Figure S1) in the Northern Hemisphere spanning 42 degrees of latitude (Tables S1 and S2). For each stream, DOC and DON data were collected at either weekly or monthly intervals, except for streams in a tallgrass prairie ecosystem (Konza Prairie: KNZ), for which we have limited DON data. For consistency across sites, we set minimum detection limits (MDL) for each solute: DOC (0.1 mg C/L), TDN (0.05 mg N/L), DON (0.01 mg N/L), NO_3^- (0.005 mg NO_3^- -N/L), and NH_4^+ (0.004 mg NH_4^+ -N/L). In addition, we only used DON values that were 5% or more of the TDN pool, to account for analytical uncertainty (Lloyd

et al., 2016). For data points that were below the MDL, values were replaced with half the MDL. To estimate DOC and TDN from the Finnish data set, we multiplied TOC and TN by 0.95 (Kortelainen et al., 2006; Mattsson et al., 2005). These calculated TDN values for the Finnish data were then used to determine DON for these sites as: $\text{DON} = \text{TDN} - (\text{NO}_3^- + \text{NH}_4^+)$. Molar DOC:DON ratios were determined from the final DOC and DON concentrations and were used to describe change over time in the DOM pool. For more details on analytical methods see Table S4. Concentrations used in this study were not flow-weighted as discharge data were not available for the same time frame as the chemistry time series nor available for all streams. Past work found that long-term data collection can account for the variety of discharge values that occur at a site, and in at least one of our sites DOC concentrations were not correlated with discharge (Coble et al., 2018; Rügge et al., 2015).

2.2 | Time series and trend analyses

We examined time series from mean monthly DOC, DON, and DOC:DON values for each stream using the longest record possible from each site (Table S3) with the exception of the Arctic site, Caribou-Poker Creeks Research Watersheds (CPC) where data are only available from May to August which coincides with the freshet and summer base flow periods. Time series were used to calculate trends using Sen slope (Hirsch et al., 1982) obtained from the *trend* package (Pohlert, 2018) in R (R Core Team, 2016) for DOC, DON, and DOC:DON ratios in each stream. Sen slope is a robust nonparametric method of regression, with the slope similar to the regression slope but less sensitive to outliers and reports a median change on the given parameter over time. Sen slopes with p -value less than .05 were considered statistically significant indicators of either increasing or decreasing trends, while slopes with p -values greater than .05 were considered insignificant and replaced with zeros for further analysis. The length of the data records across the 74 streams ranged from 8 to 45 years, where the longest starts in 1975 and all end between 2010 and 2015. Changes in analytical methods have been previously evaluated to ensure consistency over time (LUQ: McDowell et al., 2021; LMP: Coble et al., 2018; Wymore et al., 2021a; AND: Johnson et al., 2021; HBF: Campbell et al., 2021). Although we recognize that length of the record can be an important factor in trends over time (Argerich et al., 2013), we found no clear relationships between length of the record and trends in DOC, DON, and DOC:DON across our sites (Figure S2).

We used mutual information (MI) to determine the degree to which DOC and DON covary in each stream over time with the *muti* package (Scheuerell, 2017) in R (R Core Team, 2016). Mutual information is a non-parametric method that characterizes the mutual dependence of two time series (Ardón et al., 2017; Cazelles, 2004). Here, we interpret that MI values closer to 1 indicate strong synchrony between DOC and DON where MI values closer to 0 indicate little dependency between the temporal dynamics of DOC and DON. Given that MI values do not provide information on the direction of the relationship, we

TABLE 1 Sites from which DOC, DON, and DOC:DON time series were obtained with the number of individual streams used from each site

Site	Site abbreviation	Biome	Acid deposition history	Individual streams
Finland	FIN	Boreal forest	Yes	32
Caribou Poker Creek, AK	CPC ^a	Boreal forest	No	9
Hubbard Brook Experimental Forest, NH	HBFA	Temperate deciduous forest	Yes	5
Lamprey River Basin, NH	LMP	Temperate deciduous forest	Yes	9
H.J. Andrews Experimental Forest, OR	AND ^a	Coniferous temperate forest	No	9
Konza Prairie, KS	KNZ ^a	Tallgrass Prairie	Yes	2
Luquillo Experimental Forest, PR	LUQ ^a	Tropical rainforest	No	8

Note: DOC, DON, and DOC:DON molar ratios median concentrations with minimum and maximum values, biome, geology, and soil type for each site.

‘-’ represents no data available.

^aSites that are part of the long-term ecological research network.

paired MI values with eight categorical descriptions of the DOC and DON directional trends based on their respective Sen slopes. These categorical descriptions were: increasing DOC and DON, declining DOC and DON, increasing DOC with no trend in DON, increasing DOC with declining DON, no trend in DOC and increasing DON, no trends in DOC and decline in DON, a decline in DOC and no trend in DON and no trend in DOC and DON.

We obtained data from the National Atmospheric Deposition Program (NADP; NRSP-3) to test the effect of atmospheric deposition on DOM trends for sites that have nearby NADP sampling locations (CPC, HBF, AND, KNZ, and LUQ). We used time series of NO_3^- and SO_4^{2-} fluxes to identify sites historically affected by atmospheric deposition (Figure S3). Sites showing a decline in atmospheric NO_3^- and SO_4^{2-} fluxes over time were classified as affected by atmospheric deposition and those that showed constant atmospheric NO_3^- and SO_4^{2-} fluxes were classified as sites that were not affected by acid deposition (Figure S3). We corroborated this approach with the expert knowledge of authors for their respective research sites. We used NADP data from HBF for LMP as these sites are in the same region. The sites across Finland were classified by their history in atmospheric acid deposition based on longitudinal patterns where southern Finland receives the greatest deposition (Ruoho-Airola et al., 2014, 2015; Vuorenmaa, 2004).

We determined potential predictor variables of DOC, DON, and DOC:DON trends via an elastic net analysis, which is a form of penalized regression that shrinks variables that do not influence the model (Zou & Hastie, 2005). Elastic net produces a parsimonious model with the most influential variables and is minimally influenced by collinearity among predictor variables (Finlay et al., 2015). Lambda and alpha values were determined by cross-validation and choosing the lowest mean squared error (Finlay et al., 2015). Lambda controls the shrinkage of variables while alpha selects the type of penalty where alpha values between 0 and 1 denote elastic net regression (Friedman et al., 2010). DOC, DON, and DOC:DON trends that were significant ($p < .05$) were included as the response

variables into three different models focused on either ambient stream chemistry, watershed characteristics, or acid deposition history to determine if trends in DOM were related to in-stream chemistry, the surrounding landscape or location, or atmospheric deposition history. The predictor variables for the ambient stream chemistry model were mean concentrations of DOC, DON, DOC:DON, NO_3^- , NH_4^+ , Na^+ , and Ca^{+2} for each stream. Predictor variables for the watershed characteristics model were mean annual temperature (MAT, °C), mean annual precipitations (MAP, mm), mean watershed elevation (m), and watershed area (km^2). The predictor variables for the atmospheric deposition model were mean and peak NO_3^- and SO_4^{2-} atmospheric deposition. We also tested categorical variables such as soil type (i.e., clay, silt, loam, till, and moraine), watershed geology (i.e., granitic, andesite, and volcanoclastic), and forest type (i.e., evergreen, conifer, deciduous, and mixed forest) with a Kruskal–Wallis rank sum test due to uneven sample size between groups with the *Stats* package (R-Core-Team, 2016). These categorical variables were not included in the models because they are poorly balanced and lead to the overfitting of models. Watershed characteristics were obtained from the site description on each LTER's website and from Råike et al. (2012), Kortelainen et al. (2006), and Vuorenmaa (2004) for Finnish (FIN) sites. Atmospheric deposition values were obtained from the NADP monitor site at the LTER sites, and models were fit using the *glmnet* package (Friedman et al., 2010) in R (R Core Team, 2016).

Differences in DOC, DON, and DOC:DON trends between acid deposition-affected and unaffected sites were explored with a Kruskal–Wallis rank sum test due to uneven sample size between groups. We also conducted a one-sample *t*-test to determine if the means of DOM trends, in sites affected and not affected from acid deposition, were different from 0. This test is especially important for DOC:DON ratios to indicate whether changes in DOC and DON are proportional. The one-sample *t*-test was performed using the *Stats* package. All statistical analyses were conducted in R (R Core Team, 2016) using RStudio (version 1.2.1335, RStudio, Inc. Team, 2016).

DOC (mg/L)	DON (mg/L)	DOC:DON Molar ratios	Geology	Soil type
11.40 (1.9–78.85)	0.44 (0.02–12.40)	34.74 (20.89–74.47)	Granite	Ground moraine or silty clay
3.63 (0.27–29.44)	0.26 (0.02–1.91)	17.95 (10.98–19.78)	Quartz mica	Silt
2.29 (0.18–24.49)	0.08 (0.01–0.42)	32.59 (31.37–63.53)	Granite	Sandy loam
4.80 (0.04–18.43)	0.19 (0.02–0.89)	30.79 (21.51–36.44)	Granite	Glacial till
0.99 (0.19–4.27)	0.03 (0.01–0.15)	34.61 (26.60–66.50)	Andesite and volcanoclastic	Sandy loam
0.88 (0.10–11.98)	–	–	Limestone	Silty clay
1.05 (0.05–15.16)	0.05 (0.01–1.12)	23.67 (16.90–28.33)	Quartz diorite and volcanoclastic	Silty clay

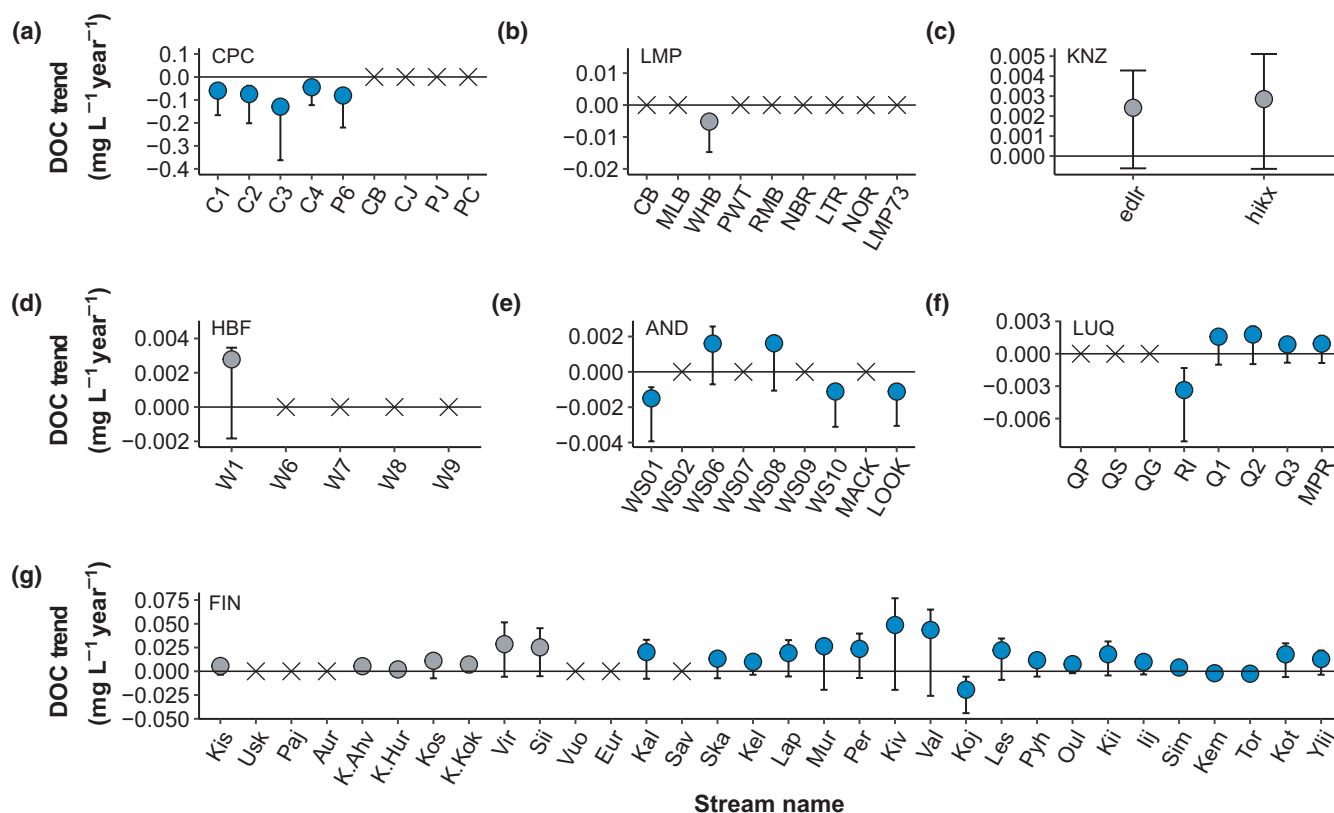


FIGURE 1 Sen slopes for DOC for each individual stream in (a) Caribou-Poker Creeks Research Watershed (CPC), (b) Lamprey River Basin (LMP), (c) Konza Prairie (KNZ), (d) Hubbard Brook Experimental Forest (HBF), (e) H.J. Andrews Experimental Forest (AND), (f) Luquillo Experimental Forest (LUQ), and (g) Finland (FIN). Circles denote streams DOC trends ($p < .05$) along with their respective 95% confidence intervals and the X are streams with no DOC trends ($p > .05$). Blue are sites not affected by acid deposition and grey are sites historically affected by acid deposition. Caribou-Poker Creeks, Lamprey, Hubbard Brook, and Andrews streams are ordered by watershed size from smallest to largest Luquillo streams are ordered by size within their respective major watersheds; Finland streams are ordered by latitude from South to North

3 | RESULTS

3.1 | DOC, DON, and DOC:DON trends

Long-term trends in concentrations of DOC (Figure 1) and DON (Figure 2) showed no consistent patterns across study sites. DOC trends ranged between -0.13 and 0.05 mg C/L per year

(median = 0.003 mg C/L per year) while DON trends ranged from -0.006 to 0.0009 mg N/L per year (median = -0.0002 mg C/L per year). Trends for DOC:DON ratios ranged between -0.05 and 0.96 per year (median = 0.02 per year).

The majority of the streams had no significant temporal trend for either DON (36 of 70; 51%) or DOC:DON ratio (42 of 70; 60%; Figure 4). For DOC, a large portion of streams showed

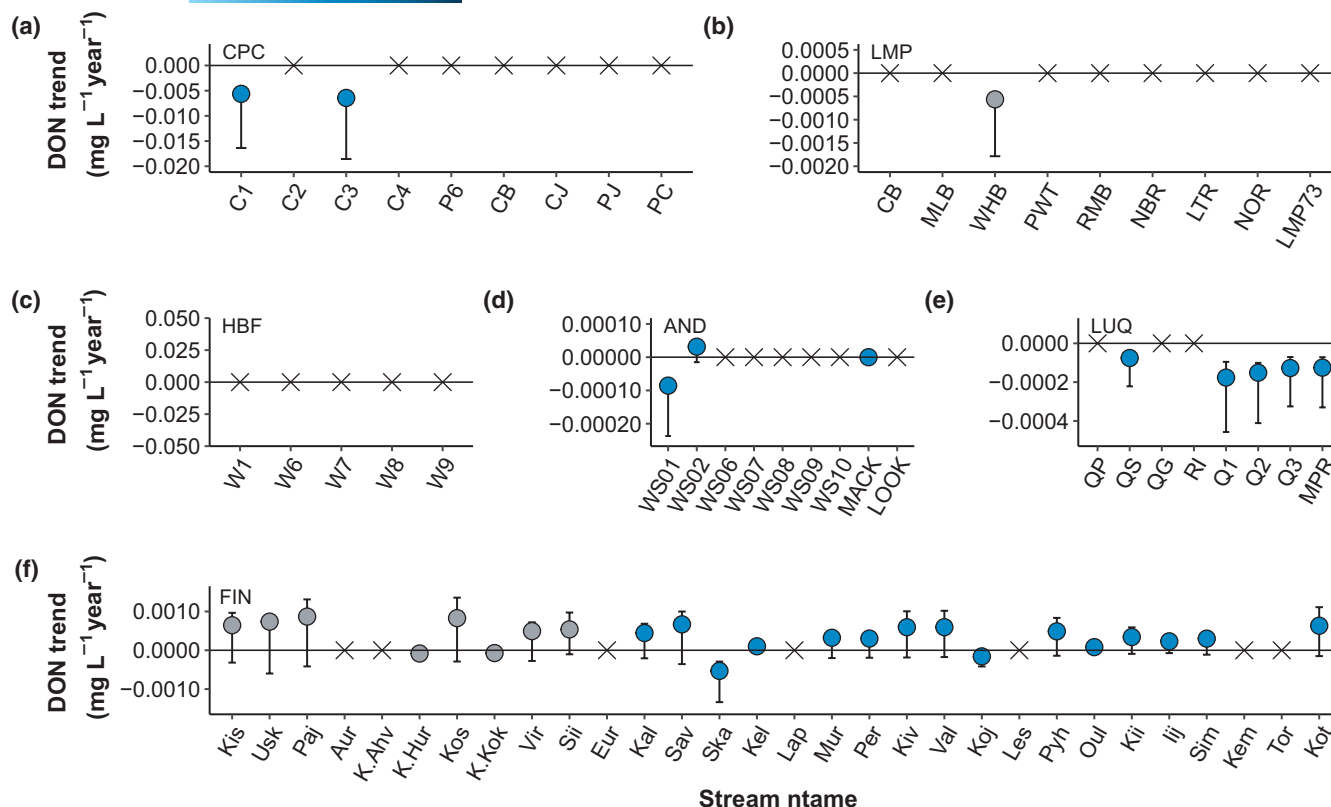


FIGURE 2 Sen slopes for DON for each individual stream in (a) Caribou-Poker Creeks (CPC), (b) Lamprey River Basin (LMP), (c) Hubbard Brook Forest (HBF), (d) H.J. Andrews Forest (AND), (e) Luquillo Experimental Forest (LUQ), and (f) Finland (FIN). Circles denote streams with DON trends ($p < .05$) with their respective 95% confidence intervals and X denote streams with no trends ($p > .05$). Note Konza and 2 Finnish streams (YLIJOKI 1 and Vuoksi Vastuupuomi 061) were excluded due to no DON data available for these streams. Blue represents sites not affected by acid deposition and grey are sites historically affected by acid deposition. Caribou-Poker Creeks, Lamprey, Hubbard Brook, and Andrews streams are ordered by watershed size; Luquillo streams are ordered by size within their respective major watersheds; Finland streams are ordered by latitude from South to North

increasing trends (32 of 74; 43%), followed by 29 streams (37%) with no trends, and 13 streams (18%) with decreasing trends (Figure 4a). For DON, 21 (30%) and 13 (19%) of the streams had significant increasing and decreasing trends, respectively (Figure 4b). For DOC:DON ratios, 20 (29%) and 8 (11%) of the streams had significant increasing or decreasing trends, respectively (Figure 4c).

The directionality of significant DOC trends was not consistent within the sites. At CPC, streams showed consistent decreasing DOC (Figure 1a) while KNZ and most of the FIN streams (44%) showed increases in DOC over time (Figure 1c,g). Other sites (LMP, AND, and LUQ; Figure 1b,e,f) exhibited variability in the direction of their significant DOC trends. Significant trends for DON were mostly declining across and within sites (Figure 2), except for FIN that had streams with both significantly increasing and declining DON trends (Figure 2f). DON trends were not significant at HBF (Figure 2c). Significant trends for DOC:DON ratios were generally positive (Figure 3) except for streams in LMP (Figure 3b) and FIN (Figure 3f), which showed declining stream DOC:DON ratios over time. There were no significant trends in DOC:DON ratios at CPC (Figure 3a).

3.2 | Synchronicity between DOC and DON

Contrary to our hypothesis, trends in DON concentrations did not consistently track those of DOC. In only five of 68 streams (7%) did DON and DOC track each other, with MI values greater than 0.5. Of these five streams, in only three did DOC and DON covary in the same direction, both increasing or both declining in BNZ and FIN (Figure 5a,f). There were also various streams (18 out of 68) across sites where DOC and DON changed in a similar direction (either increased or decreased), but concentrations did not strongly covary according to MI values (i.e., $MI < 0.5$; Figure 5a,b,e,f). For 23 streams (33%), DOC and DON trends were asynchronous (i.e., DOC and DON changed in the opposite direction and $MI < 0.5$).

3.3 | Acid deposition history

We did not find strong evidence to confirm our second hypothesis that DOC and DON would both increase in sites historically affected by atmospheric deposition. There was no difference in long-term trends in DOC concentration ($p = .38$; Figure 6a) between sites

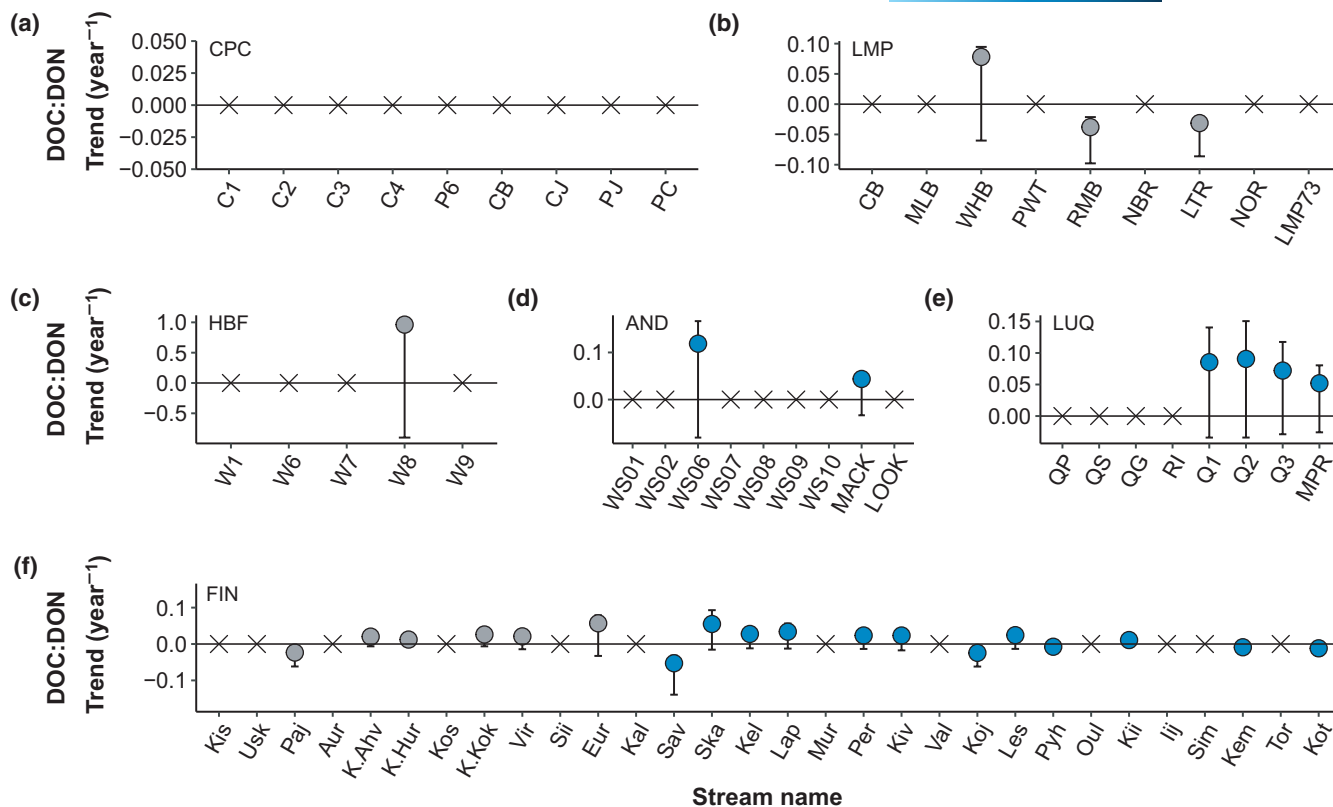


FIGURE 3 Sen slopes for DOC:DON molar ratios for each individual stream in (a) Caribou-Poker Creeks (CPC), (b) Lamprey River Basin (LMP), (c) Hubbard Brook Forest (HBF), (d) H.J. Andrews Forest (AND), (e) Luquillo Experimental Forest (LUQ), and (f) Finland (FIN). Circles denote streams with DOC:DON trends ($p < .05$) with their respective 95% confidence intervals and X denote streams with no trends ($p > .05$). Note Konza and 2 Finnish streams (YLIJOKI 1 and Vuoksi Vastuupuomi 061) were excluded due to no DON data available for these streams. Blue represents sites not affected by acid deposition and grey are sites historically affected by acid deposition. Caribou-Poker Creeks, Lamprey, Hubbard Brook, and Andrews streams are ordered by watershed size; Luquillo streams are ordered by size within their respective major watersheds; Finland streams are ordered by latitude from South to North

historically affected by acid deposition (HBF, LMP, KNZ, and Southern FIN) and those not affected by acid deposition (CPC, AND, LUQ, and Northern FIN). The DOC Sen slope values for sites that were affected by acid deposition were different from zero (t -test $p = .008$), whereas those unaffected by acid deposition were not different from zero (t -test $p = .63$). Trends in DON differed with acid deposition history ($p = .05$, Figure 6b), with trends in acid deposition affected sites being greater than trends in sites not affected by acid deposition. The DON Sen slopes were different from zero for sites affected by atmospheric deposition (t -test $p = .05$), but not for unaffected sites (t -test $p = .35$). Trends in DOC:DON ratios did not differ in their response to acid deposition history ($p = .56$). DOC:DON Sens slopes for sites affected by acid deposition were not different from 0 (t -test $p = .33$), whereas DOC:DON Sens slopes were significantly different from zero in sites unaffected by acid deposition (t -test $p = .008$, Figure 6c).

3.4 | Predictor variables of DOM trends

The elastic net models for chemistry and watershed characteristics identified several predictor variables for DOC, DON, and DOC:DON

trends. For DOC trends, the ambient stream chemistry model accounted for a large percentage of variability, followed by acid deposition, and watershed characteristics: $r^2 = .66$, $r^2 = .34$, and $r^2 = .32$, respectively (Table 2). Variables selected for DOC trends were Ca^{2+} ($\beta = -.005$), Na^+ ($\beta = .002$), and DOC ($\beta = .002$) in the stream chemistry model; MAT ($\beta = .0006$) and elevation ($\beta = -.0001$) for watershed characteristics; mean ($\beta = -.004$) and peak ($\beta = .002$) SO_4^{2-} and peak NO_3^- ($\beta = -.0007$) deposition. DOC trends did not vary across different geology types (Figure S4a), but DOC trends were the greatest in streams draining moraine and clay soil types (Figure S5a) as well as conifer forests (Figure S6a).

For DON, the ambient stream chemistry and the acid deposition models explained 27% and 34% of the variability, respectively, in trends while the watershed characteristics model did not select any variables (Table 2). The highest β coefficients for the chemistry model were mean concentrations of DON ($\beta = .001$), NO_3^- ($\beta = .0002$), Na^+ ($\beta = .0001$), and Ca^{2+} ($\beta = -.0001$). For the acid deposition model, the variables selected were mean and peak SO_4^{2-} deposition ($\beta = -.0004$ and $.0003$, respectively). Significant DON trends were only found in volcanoclastic and granitic watersheds, but trends were more constrained in volcanoclastic areas (Figure S4b).

TABLE 2 Results of elastic net models exploring the influence of ambient stream chemistry (DOC, DON, DOC:DON, NO_3^- , NH_4^+ , Ca^{2+} , Na^+), watershed characteristics (mean annual temperature (MAT °C), mean annual precipitations (MAP, mm), mean watershed elevation, and watershed area (km^2)), and atmospheric acid deposition (mean and peak NO_3^- and SO_4^{2-} deposition (kg/ha)) on DOC, DON, and DOC:DON trends (for streams with significant Sen slopes) that were considered as response variables

DOC						
Model parameters	Stream chemistry		Watershed characteristics		Acid deposition	
r^2	.66		.32		.34	
n	43		41		39	
Lambda	0.003		0.0002		0.0001	
Alpha	.96		.96		.96	
	Variable	β	Variable	β	Variable	β
	Ca ²⁺	-.005	MAT (°C)	.0006	Mean SO ₄ ²⁻	-.004
	Na ⁺	.002	Elev. (m)	-.0001	Peak SO ₄ ²⁻	.002
	DOC	.002			Peak NO ₃ ⁻	-.0007
DON						
Model parameters	Stream chemistry		Watershed characteristics		Acid deposition	
r^2	.27				.34	
n	34				34	
Lambda	0.0004				0.00001	
Alpha	.21				.83	
	Variable	β			Variable	β
	DON	.001			Mean SO ₄ ²⁻	-.0004
	NO ₃ ⁻	.0002			Peak SO ₄ ²⁻	.0003
	Na ⁺	.0001				
	Ca ²⁺	-.0001				
DOC:DON						
Model parameters	Stream chemistry		Watershed characteristics		Acid deposition	
r^2	.10		.18		.01	
n	28		26		28	
Lambda	0.05		0.07		0.06	
Alpha	.58		.19		.94	
	Variable	β	Variable	β	Variable	β
	DON	-.10	Elev. (m)	.0002	Peak NO ₃ ⁻	.0007
	Ca ²⁺	-.004				

Note: Lambda controls the shrinkage of variables while alpha selects the type of penalty where alpha values between 0 and 1 denote elastic net regression, and beta values are the model coefficients for the selected variables. KNZ was excluded for these models due to no DON data. No variables were selected in the DON trends watershed characteristics model.

Similar to DOC trends, DON trends were the greatest in streams draining moraine and clay soil types and mostly negative for loam, silt, and sandy watersheds (Figure S5b) as well as greater in conifer forests (Figure S6b).

Lastly, the goodness of fit of the models for DOC:DON ratios was low with the stream chemistry and watershed characteristics models explaining 10% and 18% of the variability in trends, respectively. The acid deposition model only explained 1% of the variance (Table 2). The predictor variables for DOC:DON trends in the stream chemistry model were DON ($\beta = -.10$), and Ca^{2+} ($\beta = -.004$),

elevation ($\beta = .0002$) in the watershed characteristics, and peak NO_3^- deposition ($\beta = .0007$) for the acid deposition model. See Figures S4–S6 for more details on watershed characteristic relationships. Opposite to DOC and DON trends, DOC:DON trends were mostly positive in volcanoclastic watersheds and showed greater variability in the granitic sites (Figure S4c). There were no statistical differences in DOC:DON trends across soil types (Figure S5c) and differences across forest types were minimal except comparing DOC:DON trends between evergreen and conifer forests (Figure S6c).

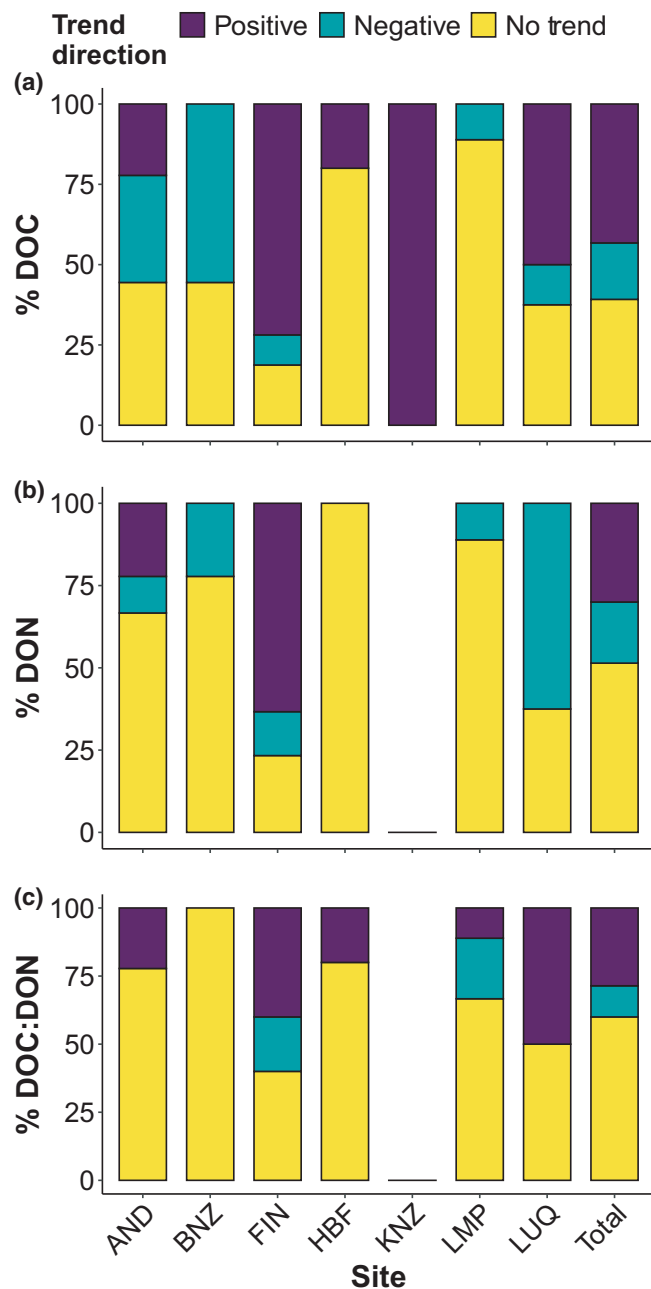


FIGURE 4 Percent of (a) DOC, (b) DON, and (c) DOC:DON ratios trends with positive (purple), negative (blue), or no trend (yellow) per site based on Sen slope $p < .05$. Note KNZ and 2 Finnish streams (YLIJOKI 1 and Vuoksi Vastuupuomi 061) were excluded from (b) and (c) due to no DON data available

4 | DISCUSSION

This is one of the first comprehensive studies in which time-series trends have been simultaneously determined for concentrations of DOC and DON and for DOM stoichiometry of the dissolved organic pool. We captured the inherent heterogeneity of DOM across a wide range of streams (74 individual streams) and six biomes by quantifying changes in three different metrics of the ambient DOM pool.

Contrary to expectations from earlier studies exploring increases in DOC concentrations in freshwater systems, primarily in temperate New England, the United Kingdom, and parts of Europe, concentrations of DOC and DON did not vary over time in any consistent pattern across study sites. Although we found 43% of our streams increasing in DOC (Figure 4), the majority of sites exhibited no significant long-term trends (Figure 4; Arvola et al., 2004; Clair et al., 2008; Coble et al., 2018; Räike et al., 2012; Rodríguez-Murillo et al., 2015), suggesting that increasing DOC is not ubiquitous across the landscape and that local context influences these long-term trends. Increasing DOC concentrations were also not exclusive to sites affected by acid deposition. For example, streams in the tropical rainforest site (LUQ) exhibit mostly positive trends in DOC concentrations that could be related to the high frequency of storm events (Wymore et al., 2017) that can also lead to an increase in the decomposition of organic matter (McDowell et al., 2013) rather than atmospheric deposition history. Another example of the unclear relationship between acid deposition and DOC trends is in the tallgrass prairie sites (KNZ) with positive DOC trends in streams affected by acid deposition. The site is in a karst landscape and well buffered against increases in hydrogen ion concentrations in soils and streams. Other directional changes at KNZ include increased woody vegetation in riparian zones (Veatch et al., 2014) and increased drying in intermittent streams (Dodds et al., 2012) which could lead to changes in instream C concentrations (Rüegg et al., 2015).

A putative hypothesis about DOM properties is that concentrations of DOC and DON are highly correlated (e.g., Campbell et al., 2000; Goodale et al., 2000; Kortelainen et al., 2006; Lepistö et al., 2008; Mann et al., 2012). For those sites showing significant temporal trends, changes in concentrations of DON did not always track those of DOC, with decoupled trends found in over 50% of the analyses. The lack of temporal synchronicity in the changes of DOC and DON concentrations is likely the result of diverse sources of DOM captured among sites, the variable state factors (*sensu lato* Jenny, 1941) represented in this analysis, and variable biogeochemical processing along flow paths and stream networks (McDowell et al., 2004; Yates et al., 2019). Across the array of streams where DOM stoichiometry is significantly changing, DOM is becoming enriched with C and relatively depleted in N suggesting that fundamental changes in the energy and nutrient balance of freshwater ecosystems is occurring over large spatial scales.

4.1 | Cross biome patterns in DOC and DON concentration trends

Many studies examining the response of DOC over time are reported from regions exposed to significant amounts of acidic deposition (Driscoll et al., 2003; Hall et al., 2021; Hruška et al., 2009; Monteith et al., 2007; Worrall et al., 2004). Whereas these studies have informed the notion that DOC concentrations are increasing in northern temperate streams (and we present complementary results), we also show that the directional change in concentrations of DOC and DON is highly variable and site- and stream-specific. Our DOC Sen

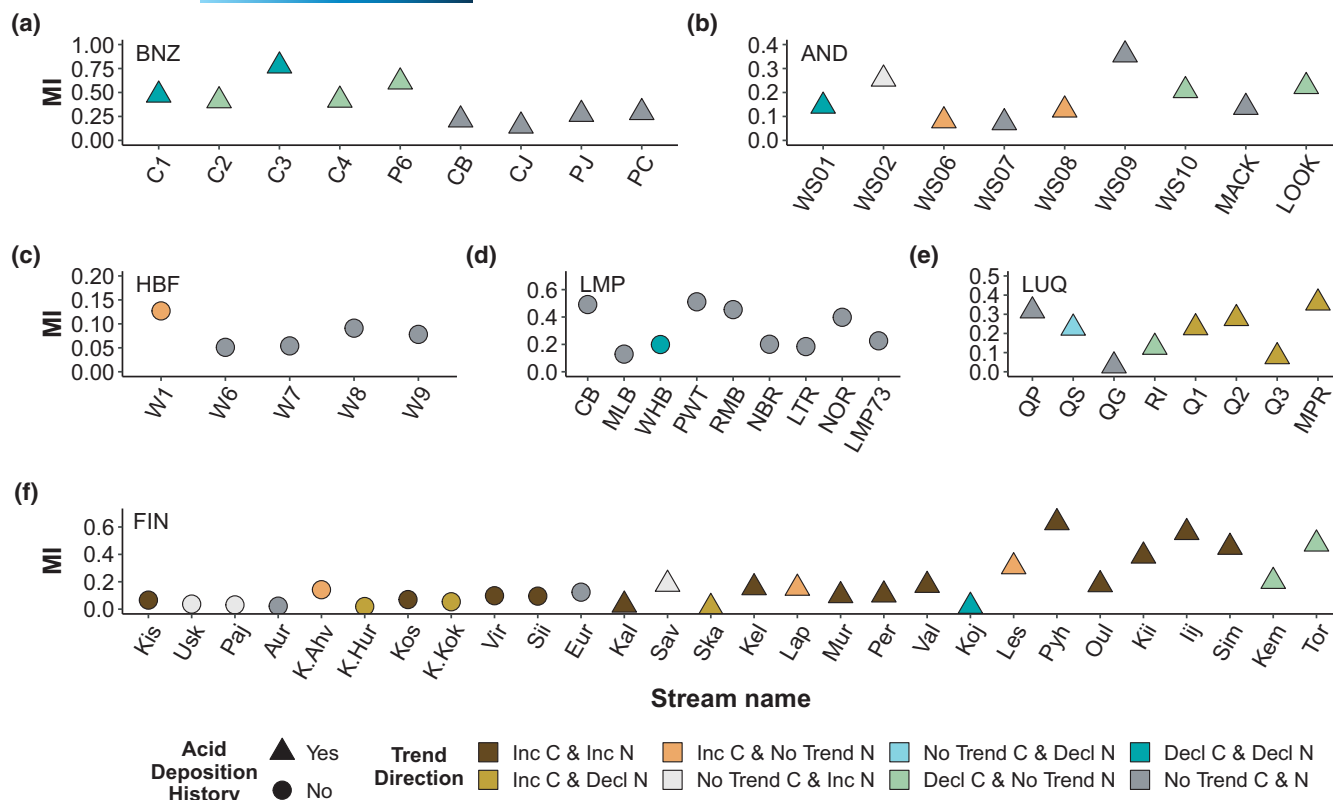


FIGURE 5 Mutual information values (MI) for DOC and DON. Shapes describe the acid deposition history by site, circles represent sites with a history of acid deposition and triangles are for sites that have not been affected by acid deposition. Colors represent the directional trend of DOC and DON based on their Sen slope where no trend are streams with a Sen slope with $p > .05$. Note Konza and 4 Finnish streams (YLIJOKI 1, Vuoksi Vastuupuomi 061, KOTIOJA 1, and Kivipuro 39) were excluded due to limited DON data available for these streams

slopes are within the range of variability reported from other studies that report values ranging between -0.25 and 0.51 mg C/L per year in streams primarily from northern latitudes (Clair et al., 2008; Coble et al., 2018; De Wit et al., 2007; Driscoll et al., 2003; Evans et al., 2005). And while the DON Sen slopes presented here are within the range of those reported earlier (0.0027 – 0.003 mg N/L per year in northern latitudes [Clair et al., 2008; Lepistö et al., 2008]), we also present negative DON trends. Studies addressing the long-term trends in DON are rarer than those of DOC, necessitating a broader assessment of DON trends. Our results suggest that changes in DOM composition may have the greatest impact in ecosystems with the lowest DOM concentrations such as tall grass prairies (KNZ) and tropical rainforest (LUQ). In these ecosystems with low DOM concentration, small changes in DOC and DON can create a large proportional change with potentially meaningful ramifications for stream metabolic regimes (Bernhardt et al., 2018) and biogeochemical reaction rates that are often limited by the availability of energy (Brailsford, Glanville, Golyshin, Johnes, et al., 2019; Brailsford, Glanville, Golyshin, Marshall, et al., 2019; Rodríguez-Cardona et al., 2021).

Contrary to our hypothesis, synchronous changes in concentrations of DOC and DON were only found in a small number of streams (Figure 5). Asynchronous changes in concentrations of DOC

and DON suggest different controls on the C-rich and N-rich fractions of the DOM pool and/or different drivers of DOC and DON flux to streams. We found numerous examples where concentrations of DOC and DON changed in opposite directions, demonstrating that the DOM pool as a whole is highly dynamic and that the different constituents of DOM do not always have the same ecological and biogeochemical sources and roles (e.g., Bernal et al., 2005; Brookshire et al., 2007; Lutz et al., 2011; McDowell et al., 2004; Wymore et al., 2015, 2018; Yates et al., 2019). For example, we found streams increasing in DOC but decreasing in DON (Figure 5) as well as sites that changed in either DOC or DON, but not in the other constituent. These scenarios suggest a biogeochemical decoupling of the C-rich and N-rich fractions of the DOM pool where DON cycling has little effect on the overall DOC pool. Changes in concentrations of DON with no significant trend in concentrations of DOC may be the result of DON being more mobile and reactive along flow paths relative to DOC due to its hydrophilic nature (Aiken et al., 1992; Hood et al., 2003; Inamdar et al., 2012;). Scenarios in which no significant trend in DOC concentrations occurs but DON concentrations decline could also occur in the nutrient limited systems where both terrestrial and aquatic biota mine the N contained within DOM (Brailsford, Glanville, Golyshin, Marshall, et al., 2019; Jones et al., 2005; Kissman et al., 2017; Mackay et al., 2020; Neff

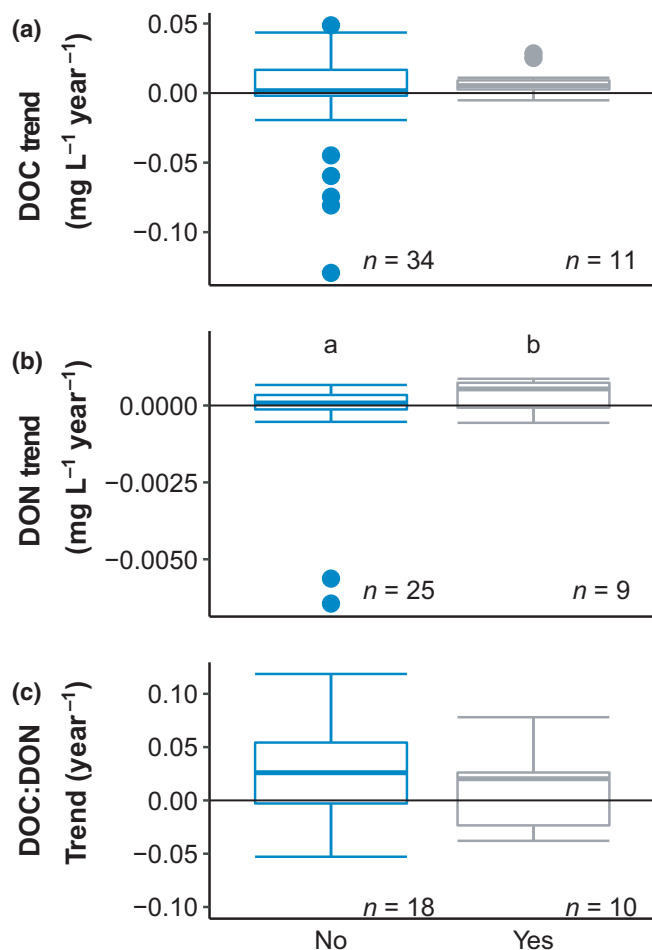


FIGURE 6 Significant Sen slopes ($p < .05$) in (a) DOC, (b) DON, and (c) DOC:DON ratios grouped by history of acid deposition where sites affected by acid deposition are in grey (FIN, HBF, LMP, KNZ) and sites not affected by acid deposition (CPC, AND, LUQ) are in blue. There is an outlier point in the DOC:DON trends for sites not affected by acid deposition that was excluded from the figure and statistics, the value is 0.96 year^{-1} from HBF. Letters denote statistically significant differences determined by Kruskal–Wallis rank-sum test (DOC $p = .38$; DON $p = .05$, and DOC:DON $p = .56$). p -values for one-sample t -test, to determine if means are different from 0 for DOC in acid deposition affected sites $p = .008$ and no acid deposition $p = .35$; DON in acid deposition affected sites $p = .05$ and no acid deposition $p = .41$; DOC:DON in acid deposition affected sites $p = .29$ and no acid deposition $p = .008$

et al., 2003; Wymore et al., 2015). We did not detect any instances where DOC concentration is declining but DON concentration is increasing, suggesting that autochthonous contributions to the DON pool are small compared to heterogeneous terrestrial inputs from the watershed, at least for the streams included in this study.

4.2 | A changing stoichiometry of the DOM pool

DOC and DON concentrations represent two ways to measure the composition of the DOM pool, yet few studies have used both bulk

elemental analyses as a way to describe the heterogeneous DOM pool (McDowell et al., 2019). Although a stoichiometric approach to understanding nutrient and elemental cycling has a rich history (Elser et al., 2000; Redfield, 1958), the principles have seldom been applied to understanding changes in bulk DOM composition over time. For those sites where a significant change in DOC:DON stoichiometry was detected, the predominant direction of change reflected the C-enrichment or N-depletion of DOM. The exception to this general pattern was in streams at the LMP site, located in the temperate deciduous forests of New England (Wymore et al., 2021a), where DOC:DON ratios are decreasing, indicating the relative N-enrichment of DOM. These sites have a high percentage of wetlands (Flint & McDowell, 2015), which are likely contributing to these changing stoichiometric ratios (Coble et al., 2019). N-enriched DOM may provide additional nutrients to microbial communities making more NH_4^+ available through mineralization. In turn, competition for dissolved inorganic N may decline with higher rates of nitrification and increased NO_3^- production and export (Wymore et al., 2019), while increasing DON concentrations instream may provide an alternative nutrient resource for uptake by the primary producers (Mackay et al., 2020). In contrast, streams with increasing DOC:DON ratios may reflect increasing watershed N demand from greater retention in soils and increasing vegetative growth, possibly from CO_2 enrichment (Craine et al., 2018; Groffman et al., 2018; Huang et al., 2015). Just as instream primary producers can take up DON compounds directly as a nutrient resource, trees can bypass microbial symbionts taking up labile forms of DON directly from soils (Neff et al., 2003), which in turn would decrease DON exports to streams leading to increases in DOC:DON ratios. Changes in DOC:DON ratios can alter rates of N transformations including nitrification (Strauss & Lamberti, 2002), and NO_3^- concentrations (Bernhardt & McDowell, 2008). While the ecosystem and biogeochemical consequences of changes in DOC:DON is a relatively understudied topic, stoichiometric shifts in this particular compartment of organic matter will likely influence other biogeochemical cycles (Wymore et al., 2019; Yates et al., 2019), driving changes in the aquatic ecosystem and downstream, creating nutrient export regimes that can affect trophic assemblages in receiving bodies of water (Schade et al., 2005).

4.3 | Atmospheric deposition

In sites historically affected by acid deposition such as HBF, LMP, and Southern Finland, Sen slopes are mostly increasing for DOC and DON concentrations, consistent with previous studies in Northeast regions of the United States, Europe, and United Kingdom (Couture et al., 2012; Driscoll et al., 2003; Evans et al., 2005; Gavin et al., 2018; Monteith et al., 2007; Worrall et al., 2004). Similar trends for DOC and DON in these sites suggest analogous watershed sources, flow paths, and release mechanisms for both the C-rich and N-rich fractions of DOM. In these acid-impacted sites, the rates of change for DOC and DON concentrations remain proportional, leading to

a stoichiometrically stable DOM pool, despite the large changes in DOC concentrations (Wymore et al., 2021c). Sites affected by atmospheric acid deposition generally show increases in DON evidenced by the significant one-sample *t* tests (Figure 6).

Conversely, sites not affected by acid deposition generally show declines in DON, falling below the DON zero line (Figure 6). These results highlight the importance of DON as a component of the DOM pool, and its sensitivity to changes in N loading from the watershed. Sites unaffected by acid deposition exhibit the greatest change in DOC:DON ratios, very likely driven by larger declines in DON concentration relative to DOC, as demonstrated by the decoupling of DOC and DON (Figure 5). Recovery from atmospheric deposition plays an important role in the delivery of DOM to aquatic ecosystems but appears to have little effect on the stoichiometry of DOM.

4.4 | Predictors of DOM trends

The ambient stream chemistry models for all DOM trend models selected major dissolved ions such as Ca^{2+} . This result supports the idea that these streams are recovering from acid deposition and as soil Ca^{2+} recovers, DOM declines due to decreased DOM solubility (Miller et al., 2016). For both DOC and DON models, ambient mean DOC and DON concentrations, respectively, were selected suggesting that streams with greater DOC or DON concentrations will experience the greatest changes over time. Ambient NO_3^- concentrations had the second-highest beta coefficients for DON trends model, demonstrating the connection between the organic and inorganic N pool in streams and how DIN can influence concentrations of DON (Wymore et al., 2015). Either peak NO_3^- or SO_4^- deposition were selected for all DOM trends in the atmospheric acid deposition models demonstrating the lasting effects and recovery of acid deposition on aquatic ecosystems, but for DOC:DON trends this model explained very little of the variance. This result suggests that although atmospheric acid deposition can influence DOC and DON concentrations, the proportional changes might not be large enough to be detected in stoichiometry, at least for the subset of streams selected in this model.

In the watershed characteristics models, mean annual temperature (MAT) and watershed elevation were selected for DOC and DOC:DON trends demonstrating that the geographical location of the streams (MAT as a surrogate for latitudinal changes), can have a strong influence on DOM over time. DOC and DON trends were greatest in granitic watersheds as this bedrock has a poor buffering capacity (Robinson, 1997) allowing a greater release of DOM to adjacent aquatic systems. This translates to an opposite pattern for DOC:DON ratios where they were lower in granitic watersheds and greater in the volcanoclastic watersheds (Figure S4). The type of soil also played a role in DOM trends being streams in silty and sandy loam landscapes the ones showing the lowest DOC and DON trends (Figure S5). This finding suggests that the adsorption to silt particles can influence DOM availability by controlling the long-term storage and export of DOC and DON (Dosskey & Bertsch, 1997; Kaiser &

Guggenberger, 2000). The higher trends of DOC and DON in moraine sites (Figure S5) could be due to greater OM availability and associated microbial decomposition activity (Bruhn et al., 2021). Collectively, the results of these models support the hypothesis that regional state factors such as geology and soil type are important controls of stream long-term DOM trends.

5 | CONCLUSION

This study adds to the growing body of literature on long-term trends of stream water DOM by expanding the scope of past studies through the inclusion of underrepresented biomes and ecosystems including tropical rainforests, arctic taiga, and tallgrass prairie. Our study also provides one of the first large-scale assessments of long-term trends in concentrations of DON and DOM stoichiometry in streams. We demonstrated increases in DOC concentrations in sites recovering from acid deposition, similar to previous findings, and we have shown that those are accompanied by proportional increases in DON. In addition, there can be increasing trends in DOC in sites with no atmospheric acid deposition history where in these cases, trends are associated with local state factors such as soil and geology. Although we have shown that positive trends in DOC can occur regardless of the acid deposition history, we also demonstrate that increases in DOC are not ubiquitous across broad latitudinal gradients. Declining trends in DON suggest on the differential cycling of DOC and DON in soils and within streams. The biomes in which streams are embedded are expected to influence the biogeochemistry of those systems (Dodds et al., 2019), but exactly how is poorly known. Changes in DOC and DON concentrations will have implications for in-stream biogeochemical processes as well as the bulk composition of DOM exports to receiving bodies of water, especially those systems where changes in DOC and DON are decoupled. Continued monitoring of these long-term trends in DOM concentration and stoichiometry in response to climatic and landscape attributes is important to better understand the ultimate fate of DOM and nutrients in freshwater ecosystems in the face of global change.

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AUTHOR CONTRIBUTIONS

Bianca M. Rodríguez-Cardona and Adam S. Wymore conceived the project. Bianca M. Rodríguez-Cardona, Adam S. Wymore, Ashley A. Coble, and Carla López-Lloreda, organized and synthesized the data, and Jeremy B. Jones, Sherri L. Johnson, Pirkko Kortelainen, Walter K. Dodds, and William H. McDowell provided data. Bianca M. Rodríguez-Cardona, Adam S. Wymore, Alba Argerich analyzed the data. All authors helped to interpret that data. Bianca M. Rodríguez-Cardona wrote the initial draft of the manuscript with significant input from Adam S. Wymore, Robert G. M. Spencer, Ashley M. Helton, Susana Bernal, and William H. McDowell. All co-authors discussed hypotheses, data and results, and commented and revised the manuscript.

DATA AVAILABILITY STATEMENT

The data used in these analyses represent a synthesis of multiple data sets. The individual data sets and their associated repositories and references can be found in Tables S4 and S5. The archived data set (Wymore et al., 2021b) is openly shared at the Environmental Data Initiative (EDI) according to FAIR principles of data sharing (<https://environmentaldatainitiative.org>). Data can be accessed here: <https://doi.org/10.6073/pasta/50965f9e091ffa833da3c73bce2467fa>.

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SUPPORTING INFORMATION

Additional supporting information may be found in the online version of the article at the publisher's website.

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Supporting Information

Shifting Stoichiometry: Long-term trends in stream dissolved organic matter reveal altered C:N ratios due to history of atmospheric acid deposition

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Figure S1 Location of all sites across the Northern Hemisphere, see Table 1 for complete site names.

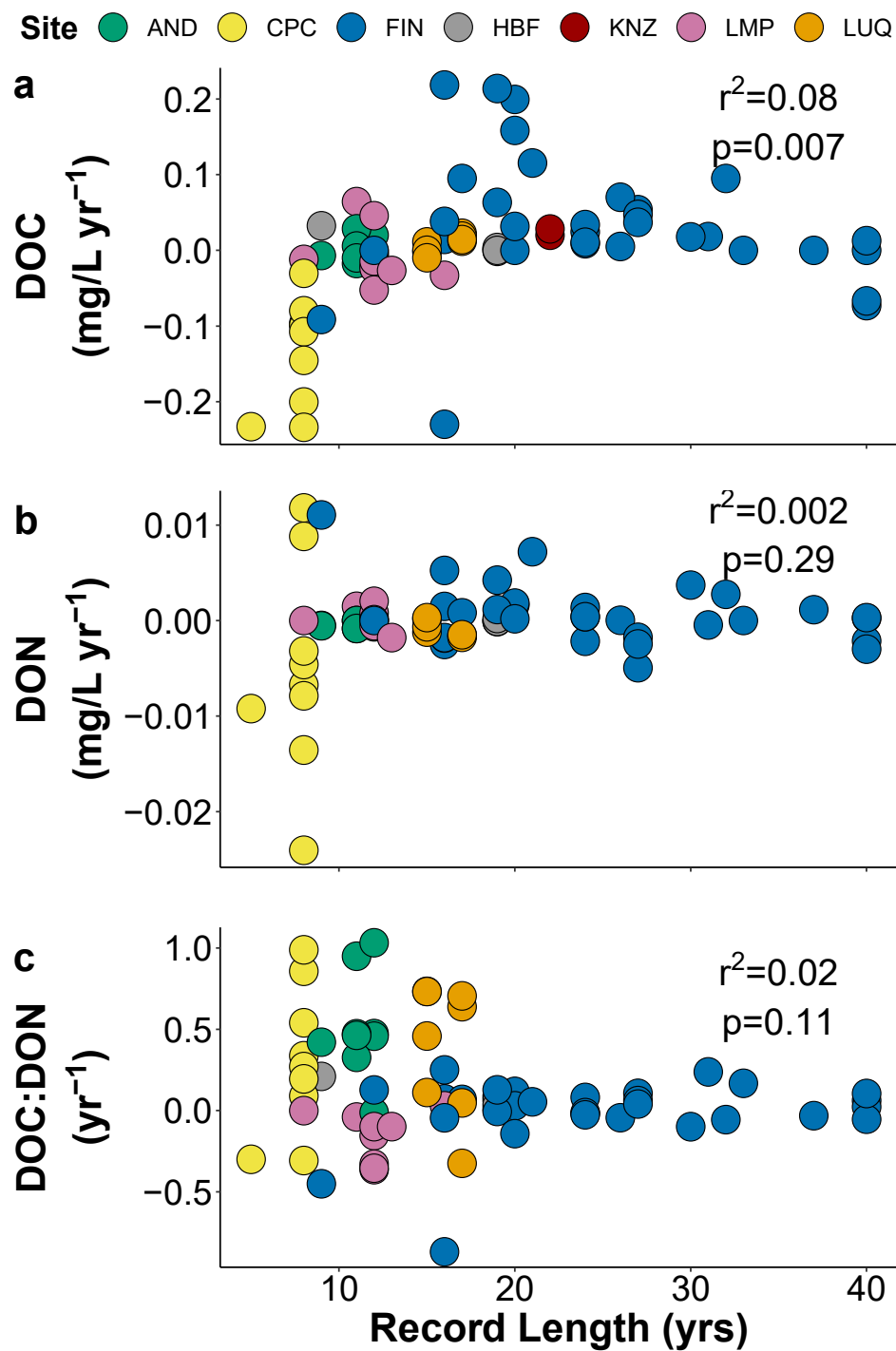


Figure S2. Relationship between trends of (a) DOC, (b), DON, (c) DOC:DON and the total record length of years used to calculate DOM trends. Each point is an individual stream for each site. R^2 and p -values were determined from simple linear regressions.

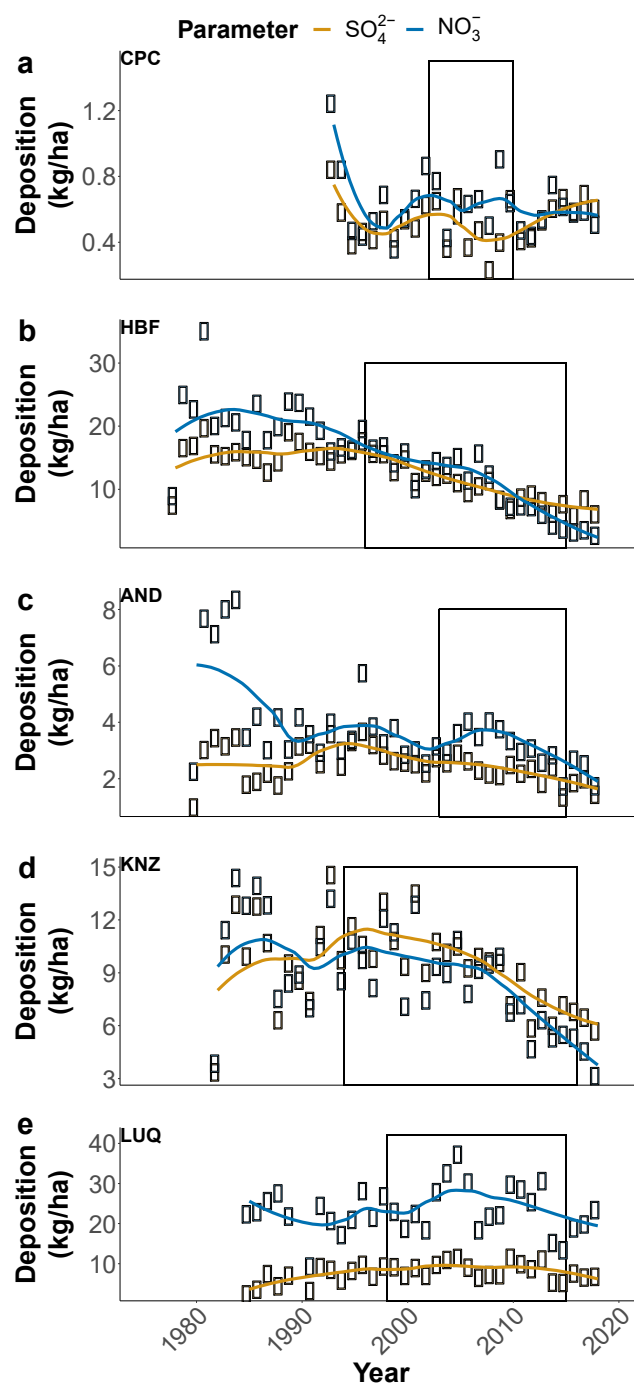


Figure S3. Time series of monthly SO_4 (Yellow) and NO_3 (blue) atmospheric deposition from National Atmospheric Deposition Program collection sites at (a) Caribou-Poker Creeks Research Watershed, (b) Hubbard Brook Experimental Forest, (c) H.J. Andrews Experimental Forest, (d) Konza Prairie, (e) and Luquillo Experimental Forest. The grey box corresponds to the length of records for stream water DOC and DON concentrations and DOC:DON molar ratios analyzed for trends in this study. Note different y-axes.

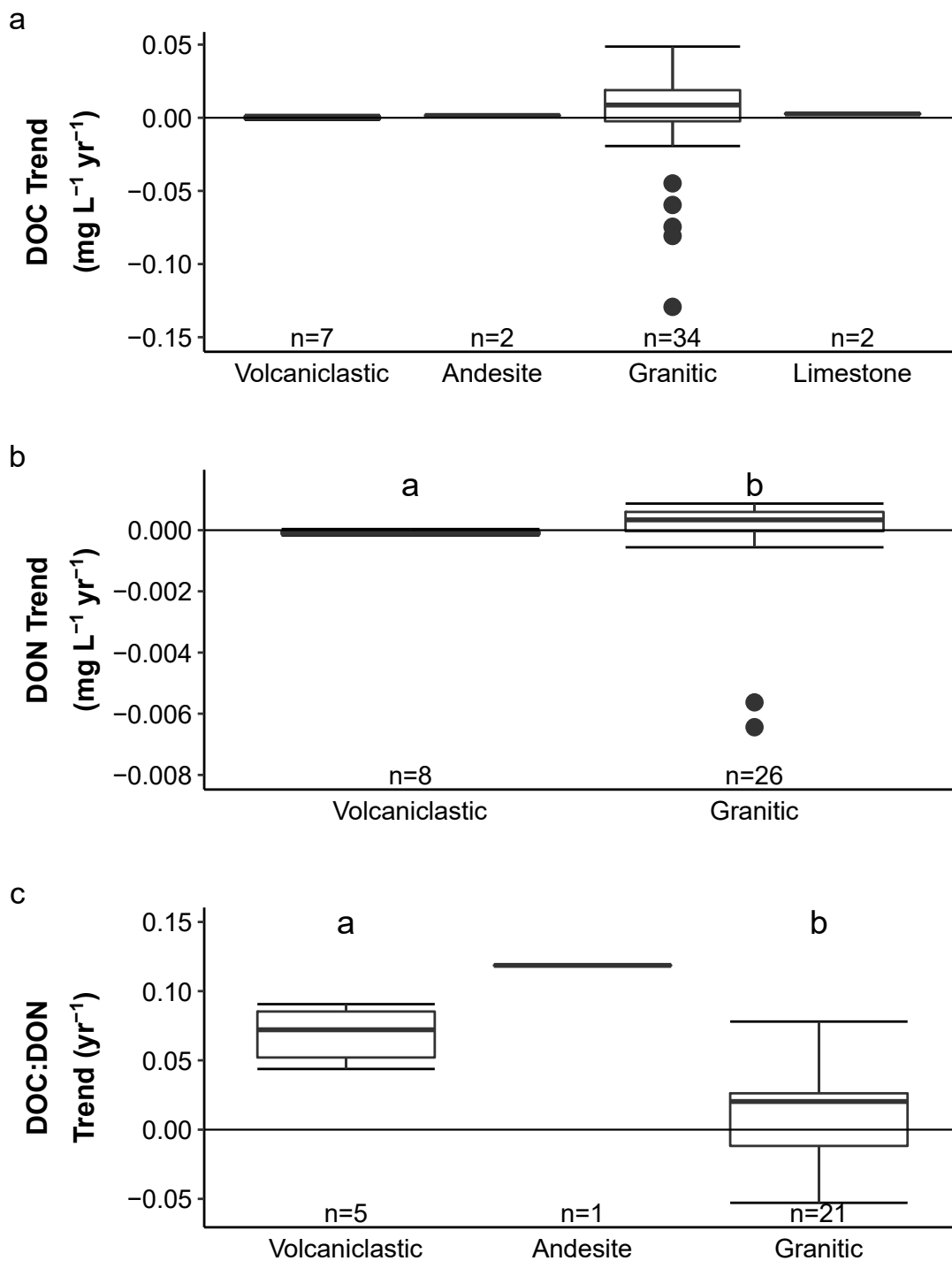


Fig S4. Boxplots of DOC ($p=0.22$), DON (0.01), DOC:DON ($p=0.005$) trends (significant Sen slopes, $p<0.05$) by watershed geology. Geology types with $n=1$ were excluded from Kruskal-Wallis test. An outlier for DOC:DON trends near 1 yr^{-1} was removed, but differences were still significant when it was included ($p=0.01$).

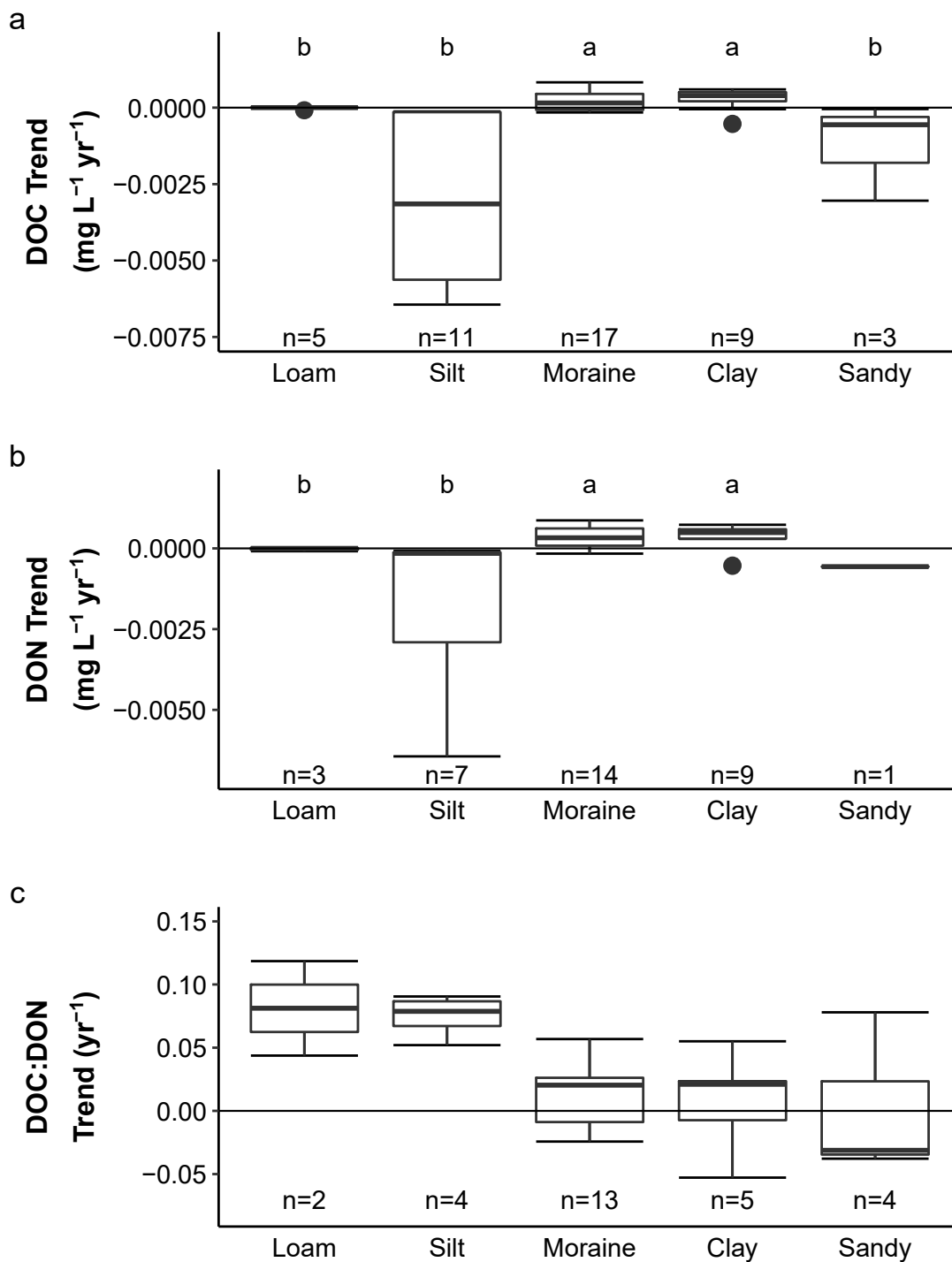


Fig S5. Boxplots of DOC ($p=0.0001$), DON ($p=0.002$), DOC:DON ($p=0.06$) trends (significant Sen slopes, $p<0.05$) by watershed soil type. Soil types with $n=1$ were excluded from Kruskal-Wallis test.

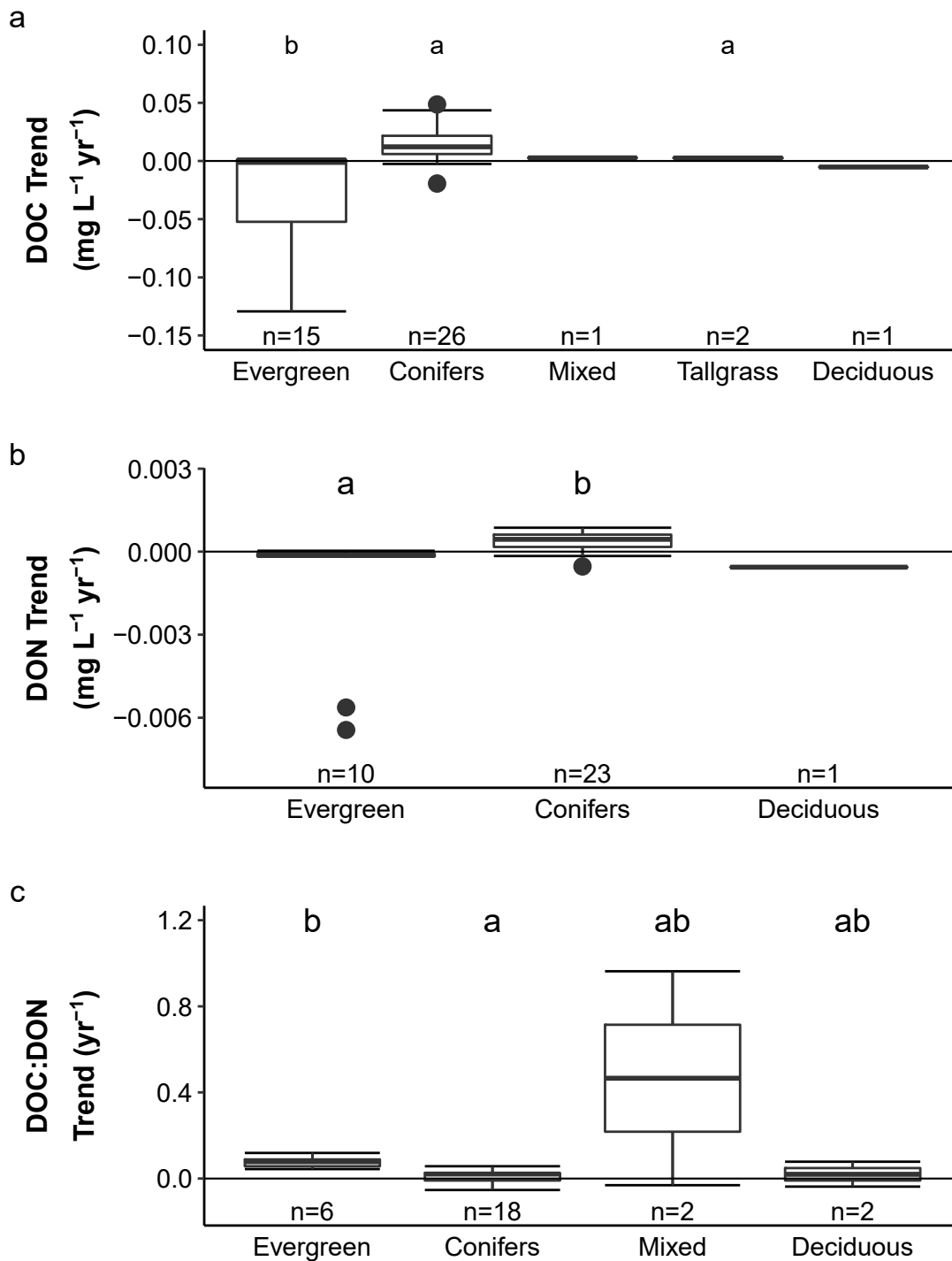


Fig S6. Boxplots of (a) DOC ($p=0.00001$), DON ($p=0.0001$), DOC:DON ($p=0.03$) trends (significant Sen slopes, $p<0.05$) by watershed forest type. Forest types with $n=1$ were excluded from Kruskal-Wallis test.

Table S1. Individual stream (with abbreviated stream name) characteristics such as latitude (Lat) and longitude (Long), mean annual temperature (MAT), mean annual precipitation (MAP), watershed elevation (Elev) and area along with mean stream pH, temperature (Temp), and Conductivity (Cond). Reference describe sites that are specifically labeled reference watersheds, or no management or manipulations have been documented for these watersheds. Forested in Finland sites are less impacted than those described as suburban. – denote data is not available for that stream.

Site	River Name	Stream Name	Abbrev. Stream Name	Lat.	Long.	MAT (°C)	MAP (mm)	Elev. (m)	Area (km ²)	pH	Temp (°C)	Cond (µs/cm)	Land-use History	Anthropogenic Activity
FIN	Aurajoki	Aura 54 ohikulku va6401	Aur	60.47	22.36	5.1	660	10	874	-	-	-	Suburban	Disturbed
FIN	Eurajoki	Eura 42 Pori-Rma va6900	Eur	61.20	21.73	4.3	622	9	1336	-	-	-	Suburban	Disturbed
FIN	Iijoki	Iijoki Raasakan voimal	Iij	65.33	25.42	1.6	474	17	14191	-	-	-	Forested	Forested
FIN	Kalajoki	Kalajoki 11000	Kal	61.95	27.22	2.3	524	110	13	-	-	-	Agriculture	Disturbed
FIN	-	Kelopuro 28	Kel	63.16	30.69	3.2	554	166	0.76	-	-	-	Forested	Forested
FIN	Kemijoki	KEMIJOKI ISOHAARA 14000	Kem	65.79	24.55	1.3	516	13	51127	-	-	-	Forested	Forested
FIN	Kiiminginjoki	Kiiminkij 13010 4-tien s	Kii	65.18	25.36	1.7	468	0	3814	-	-	-	Forested	Forested

FIN	Kiskonjoki	Kisko 14 Vanhak va6111	Kis	60.13	23.16	5.1	612	12	1047	-	-	-	Agriculture	Disturbed
FIN	-	Kivipuro 39	Kiv	63.87	28.65	3.2	554	186	0.54	-	-	-	Forested	Forested
FIN	Kokemäenj oki	Kojo 35 Pori- Tre	Koj	63.88	28.67	4.2	615	210	1.20	-	-	-	Agriculture	Forested
FIN	Koskenkylä njoki	Koskenkylänj oki 3,0 6030	Kos	60.50	25.95	3.2	554	9	895.00	-	-	-	Agriculture	Disturbed
FIN	-	Kotioja 1	Kot	66.14	26.15	3.2	554	162	18.00	-	-	-	Forested	Forested
FIN	Kymijoki	Kymij Huruksela 033 5600	K.Hur	60.49	26.45	4.7	608	20	37159	-	-	-	Forested	Forested
FIN	Kymijoki	Kymijoki Ahvenkoski 001	K.Ahv	60.49	26.45	3.2	554	0	-	-	-	-	Forested	Forested
FIN	Kymijoki	Kymijoki Kokonkoski 014	K.Kok	60.53	26.91	3.2	554	12	-	-	-	-	Suburban	Disturbed
FIN	Lapuanjoki	Lapuanjoki 9900	Lap	63.53	22.53	2.9	536	9	4122	-	-	-	Agriculture	Disturbed
FIN	Lestijoki	Lestijoki 10800 8-tien s	Les	64.06	23.66	2.5	528	5	1373	-	-	-	Suburban	Disturbed
FIN	-	Murtopuro 42	Mur	63.76	28.49	3.2	554	179	4.90	-	-	-	Forested	Forested

FIN	Oulujoki	Oulujoki 13000	Oul	65.02	25.47	1.9	460	0	22845	-	-	-	Forested	Forested
FIN	Paimionjoki	Pajo 44 Isosilta va6301	Paj	60.46	22.68	4.9	640	10	1088	-	-	-	Agriculture	Forested
FIN	Perhonjoki	Perhonjoki 10600	Per	63.85	23.22	2.6	534	5	2524	-	-	-	Suburban	Disturbed
FIN	Pyhäjoki	Pyhäjoki Hourunk 11400	Pyh	64.46	24.27	3.2	554	10	3712	-	-	-	Suburban	Disturbed
FIN	-	Savi 12 mittapato	Sav	62.51	26.03	3.2	554	94	5.39	-	-	-	Agriculture	Disturbed
FIN	Siikajoki	Siikajoki 8- tien s 11600	Sii	60.60	22.67	2.1	486	51	15.40	-	-	-	Forested	Forested
FIN	Simojoki	Simojoki as. 13500	Sim	65.66	25.08	1.5	490	9	4218	-	-	-	Forested	Forested
FIN	Kyrönjoki	Skatila vp 9600	Ska	63.09	21.89	3.4	525	3	3160	-	-	-	Forested	Forested
FIN	Tornionjoki	Tornionj kukkola 14310	Tor	65.96	24.05	1.1	540	21	4923	-	-	-	Forested	Forested
FIN	Uskelanjoki	Uske 16 Salon yp va6101	Usk	60.39	23.13	4.8	618	3	566	-	-	-	Agriculture	Disturbed
FIN	-	Valipuro 38	Val	63.87	28.66	3.2	554	186	0.86	-	-	-	Forested	Forested

FIN	Virojoki	Virojoki 006 3020	Vir	60.58	27.71	4.6	624	9	357	-	-	-	Agriculture	Disturbed
FIN	-	Vuoksi Vastuupuomi 061	Vuo	61.20	28.78	3.8	671	70	61466	-	-	-	Forested	Forested
FIN	-	Ylijoki 1	Ylij	66.15	26.16	3.2	554	164	56	-	-	-	Forested	Forested
CPC	-	C1	C1	65.15	-147.65	-4.3	312	537	6.70	7.42	4.00	53.3	Reference	Forested
CPC	-	C2	C2	65.16	-147.60	-4.3	312	375	5.20	7.66	4.90	78.7	Reference	Forested
CPC	-	C3	C3	65.14	-147.57	-4.3	312	357	5.70	7.54	3.80	72.7	Reference	Forested
CPC	-	C4	C4	65.16	-147.50	-4.7	309	267	10.00	7.82	5.09	103	Burned 1999	Forested
CPC	-	CB	CB	65.10	-147.40	-5.4	296	250	16.70	7.55	5.00	67.4	Reference	Forested
CPC	-	CJ	CJ	65.15	-147.49	-4.7	309	224	41.70	7.60	6.50	82	Reference	Forested
CPC	-	P6	P6	65.18	-147.39	-5.6	294	465	7	7.82	4.18	123	Burned 1999	Forested
CPC	-	PC	PC	65.15	-147.48	-4.7	309	217	101.50	7.81	8.00	102.5	Burned 1999	Forested
CPC	-	PJ	PJ	65.15	-147.48	-4.7	309	222	59.80	7.85	8.49	111	Reference	Forested

HBF	-	Watershed 1	W1	43.95	-71.73	4.8	1160	482	0.12	5.53	8.00	13	Wallastonite Additions	Disturbed
HBF	-	Watershed 6	W6	43.95	-71.74	4.8	1160	546	0.13	5.21	5.50	12.5	Reference	Forested
HBF	-	Watershed 7	W7	43.93	-71.77	4.8	1160	614	0.77	5.99	4.80	12.6	Reference	Forested
HBF	-	Watershed 8	W8	43.93	-71.76	4.8	1160	602	0.59	5.66	4.60	12.9	Reference	Forested
HBF	-	Watershed 9	W9	43.93	-71.75	4.8	1160	696	0.68	4.64	3.40	17.9	Reference	Forested
LMP	-	College Brook	CB	43.14	-70.94	8.9	1169	17	2.27	7.16	10.74	524	Urban	Disturbed
LMP	-	Lamprey River	LMP73	43.10	-70.95	8.9	1169	14	479.20	6.70	11.93	87	Suburban	Forested
LMP	-	Little River	LTR	43.11	-71.01	8.9	1169	26	51.70	6.69	9.55	62.9	Suburban	Forested
LMP	-	Moonlight Brook	MLB	43.08	-70.94	8.9	1169	4	0.90	6.58	9.48	455	Urban	Disturbed
LMP	-	North Branch	NBR	43.06	-71.24	8.9	1169	64	41.50	6.46	10.86	74	Suburban	Forested
LMP	-	North River	NOR	43.08	-71.04	8.9	1169	30	128.90	6.62	11.21	61	Suburban	Forested
LMP	-	Pawtuckaway	PWT	43.10	-71.20	8.9	1169	136	2.60	5.88	12.16	25	Forest	Forested

LMP	-	Rum Brook	RMB	43.05	-71.03	8.9	1169	30	4.90	6.68	9.84	113	Suburban	Forested
LMP	-	Wednesday Hill Brook	WHB	43.12	-71.00	8.9	1169	32	1.00	7.09	9.46	221	Suburban	Disturbed
AND	-	Lookout Creek	LOOK	44.28	-122.10	7.5	1950	422	62.42	7.40	-	31.3	Partially Harvested	Forested
AND	-	Mack Creek	MACK	44.22	-122.17	7.5	1950	755	5.81	7.30	-	26.75	Partially Harvested	Forested
AND	-	Watershed 1	WS01	44.21	-122.26	7.5	1950	439	0.96	7.50	-	43.1	Clear-cut	Forested
AND	-	Watershed 2	WS02	44.21	-122.23	7.5	1950	545	0.60	7.50	-	36.45	Reference	Forested
AND	-	Watershed 6	WS06	44.27	-122.18	7.5	1950	878	0.13	7.50	-	38.6	Clear-cut	Forested
AND	-	Watershed 7	WS07	44.27	-122.17	7.5	1950	918	0.15	7.50	-	42.9	Partial Overstory Harvest	Forested
AND	-	Watershed 8	WS08	44.27	-122.17	7.5	1950	962	0.21	7.50	-	38.3	Reference	Forested
AND	-	Watershed 9	WS09	44.20	-122.25	7.5	1950	426	0.09	7.50	-	48.45	Reference	Forested
AND	-	Watershed 10	WS10	44.22	-122.25	7.5	1950	461	0.10	7.50	-	43.6	Clear-cut	Forested
KNZ	-	Elder Spring	edlr	39.10	-96.61	12.1	866	337	-	-	-	-	Reference	Forested

KNZ	-	Hiking Trail	hikx	39.11	-96.61	12.1	866	324	-	-	-	-	Agriculture	Forested
LUQ	-	Bisley 1	Q1	18.32	-65.75	22.8	1768	377	0.07	7.26	23.00	97.4	Reference	Forested
LUQ	-	Bisley 2	Q2	18.31	-65.75	22.8	1768	218	0.06	7.29	22.90	92.6	Reference	Forested
LUQ	-	Bisley 3	Q3	18.31	-65.75	22.8	1768	198	0.35	7.32	22.63	85.6	Reference	Forested
LUQ	-	Qda. Guaba	QG	18.28	-65.79	21.4	3229	642	0.13	6.75	20.80	47	Reference	Forested
LUQ	-	Qda. Prieta	QP	18.32	-65.82	22	4587	431	0.31	7.02	22.00	79.6	Reference	Forested
LUQ	-	Qda. Sonadora	QS	18.32	-65.82	22	3495	740	2.60	6.90	21.60	49.8	Reference	Forested
LUQ	-	Río Icacos	RI	18.28	-65.79	21.4	2254	686	3.30	6.76	21.00	55.5	Reference	Forested
LUQ	-	Mameyes Puente Roto	MPR	18.33	-65.75	23	2843	498	17.7	7.44	23.3	94.5	Reference	Forested

Table S2. Mean ambient solute concentrations for every stream. – denote data is not available for that stream.

Site	Stream	DOC (mg/L)	DON (mg/L)	TDN (mg/L)	NO ₃ ⁻ (mg/L)	NH ₄ ⁺ (mg/L)	PO ₄ ³⁻ (mg/L)	Si (mg/L)	TDP (mg/L)	Na ⁺ (mg/L)	K ⁺ (mg/L)	Ca ²⁺ (mg/L)	Mg ²⁺ (mg/L)	SO ₄ ²⁻ (mg/L)	Cl ⁻ (mg/L)
FIN	Aur	13.30	0.64	2.09	1.40	58	81	-	0.04	8.00	3.20	10.80	6.90	-	-
FIN	Eur	9.31	0.51	1.52	0.84	210	17	-	0.01	7.40	3.20	13.80	5.60	-	-
FIN	Iij	9.69	0.32	0.37	0.04	8	8	-	0.01	1.70	0.60	2.90	1.00	-	-
FIN	Kal	19.95	0.69	1.33	0.60	85	40	-	0.04	4.60	2.30	6.70	4.00	-	-
FIN	Kel	9.50	0.22	0.25	0.01	9	2	-	0.00	1.00	0.30	1.10	0.20	-	-
FIN	Kem	7.60	0.28	0.33	0.04	12	6	-	0.01	1.70	0.60	4.40	1.30	-	-
FIN	Kii	14.06	0.42	0.51	0.06	8.50	11	-	0.02	2.20	0.70	3.30	1.30	-	-
FIN	Kis	9.50	0.53	0.95	0.44	25	21	-	0.02	3.80	1.60	6.70	2.80	-	-
FIN	Kiv	27.08	0.47	0.48	0.01	6	5	-	0.01	1.50	0.60	1.30	0.80	-	-
FIN	Koj	9.50	0.44	1.05	0.53	64	16	-	0.01	6.80	2.00	7.70	2.70	-	-

FIN	Kos	9.41	0.50	1.43	0.96	43	42	-	0.02	6.30	3.20	9.70	5.10	-	-
FIN	Kot	15.20	0.44	0.52	0.07	10	10	-	-	1.20	0.50	5.40	2.00	-	-
FIN	K.Hur	7.41	0.31	0.54	0.22	22	3	-	0.01	6.23	1.50	5.30	1.50	-	-
FIN	K.Ahv	7.32	0.33	0.57	0.24	22	4	-	0.01	6.20	1.60	5.50	1.60	-	-
FIN	K.Kok	7.13	0.31	0.55	0.23	21	3	-	0.01	6.40	1.60	5.40	1.50	-	-
FIN	Lap	19.00	0.63	1.53	0.70	230	34	-	0.02	6.50	3.10	8.50	4.00	-	-
FIN	Les	19.00	0.58	0.88	0.27	54	30	-	0.03	2.70	1.40	3.90	1.70	-	-
FIN	Mur	23.75	0.45	0.51	0.01	7.50	11	-	0.01	1.50	0.60	1.70	0.80	-	-
FIN	Oul	8.93	0.29	0.36	0.05	14	6	-	0.01	1.70	0.70	2.90	1.00	-	-
FIN	Paj	11.40	0.63	2.14	1.40	56	98	-	0.04	7.90	4.30	9.90	7.10	-	-
FIN	Per	19.00	0.58	1.05	0.36	110	30	-	0.02	2.70	1.50	4.60	1.70	-	-
FIN	Pyh	17.10	0.56	0.89	0.30	44	23	-	0.03	4.60	1.90	9.90	2.70	-	-
FIN	Sav	10.45	0.55	1.73	1.10	80	65	-	0.03	6.60	2.80	9.22	6.70	-	-

FIN	Sii	18.05	0.54	0.82	0.22	55	36	-	0.03	3.10	1.40	4.45	2.05	-	-
FIN	Sim	11.40	0.37	0.45	0.04	13	5	-	0.01	1.50	0.60	3.30	1.40	-	-
FIN	Ska	19.00	0.64	1.90	0.98	190	43	-	0.02	7.10	3.10	9.30	4.84	-	-
FIN	Tor	5.42	0.21	0.27	0.01	7	5	-	0.01	1.40	0.70	3.80	1.00	-	-
FIN	Usk	10.45	0.58	1.90	1.20	79.50	73	-	0.03	10.00	3.80	12.00	8.00	-	-
FIN	Val	30.35	0.48	0.52	0.01	7	3.25	-	0.01	1.10	0.30	0.80	0.50	-	-
FIN	Vir	14.25	0.57	0.92	0.34	30	16	-	0.02	3.30	1.60	5.80	1.70	-	-
FIN	Vuo	6.75	0.22	0.42	0.15	8	2	-	0.00	5.20	1.30	5.10	1.30	-	-
FIN	Ylij	13.11	0.44	0.57	0.06	19	12	-	-	1.30	0.60	5.25	2.00	-	-
CPC	C1	3.66	0.24	0.53	0.27	27.17	-	15.11	-	0.97	0.33	7.65	1.57	3.18	0.28
CPC	C2	2.63	0.25	0.82	0.56	26.61	-	21.76	-	1.04	0.37	10.81	3.18	6.41	0.33
CPC	C3	5.01	0.30	0.71	0.43	26.61	-	16.16	-	1.18	0.38	11.07	1.82	7.17	0.27
CPC	C4	2.28	0.24	0.87	0.65	25.21	-	21.31	-	1.30	0.57	15.71	3.01	7.08	0.32

CPC	CB	3.28	0.23	0.54	0.33	16.81	-	20.80	-	1.14	0.41	10.18	2.32	4.89	0.34
CPC	CJ	5.06	0.31	0.62	0.32	16.11	-	22.68	-	1.27	0.52	13.71	2.51	6.36	0.28
CPC	P6	4.36	0.27	0.72	0.46	24.16	-	20.93	-	1.32	0.51	18.80	3.17	15.26	0.38
CPC	PC	4.74	0.28	0.54	0.27	15.97	-	23.50	-	1.27	0.58	15.92	3.16	8.40	0.39
CPC	PJ	5.10	0.33	0.60	0.26	21.01	-	22.47	-	1.38	0.66	17.18	3.44	9.27	0.41
HBF	W1	2.28	0.08	0.28	0.17	4	0.3	5.29	-	0.69	0.11	1.18	0.16	0.98	0.43
HBF	W6	1.94	0.07	0.12	0.01	4	0.3	4.26	-	0.77	0.15	0.63	0.21	1.21	0.41
HBF	W7	1.68	0.06	0.10	0.02	4	0.3	5.20	-	0.77	0.15	0.91	0.30	1.19	0.37
HBF	W8	2.63	0.07	0.13	0.04	4	0.7	5.80	-	0.83	0.16	0.76	0.30	1.18	0.38
HBF	W9	7.52	0.14	0.21	0.02	4	0.3	6.10	-	0.71	0.16	0.62	0.21	1.15	0.39
LMP	CB	4.16	0.23	0.83	0.60	38.87	16.51	10.31	-	80.79	5.01	23.84	6.56	4.55	140.46
LMP	LMP73	5.25	0.21	0.37	0.12	0.02	0.00	5.18	-	14.54	1.15	6.37	1.32	1.87	24.25
LMP	LTR	5.45	0.17	0.30	0.10	0.01	0.00	7.17	-	11.93	0.79	5.04	1.16	1.38	20.16

LMP	MLB	1.64	0.09	0.79	0.70	0.04	0.00	10.85	-	70.66	3.86	37.53	6.56	4.47	163.60
LMP	NBR	5.03	0.18	0.25	0.05	0.01	0.00	6.39	-	14.36	0.73	4.98	1.05	1.62	24.55
LMP	NOR	5.59	0.20	0.28	0.05	0.01	0.00	6.42	-	10.53	0.95	5.44	1.03	1.82	16.68
LMP	PWT	5.48	0.19	0.24	0.01	0.01	0.00	7.87	-	2.06	0.76	3.98	1.01	1.00	2.49
LMP	RMB	5.56	0.21	0.38	0.11	0.02	0.01	10.67	-	20.59	2.14	10.80	1.67	2.16	36.84
LMP	WHB	3.07	0.14	0.81	0.65	0.01	0.00	9.05	-	32.23	1.39	16.18	6.70	3.59	50.46
AND	LOOK	0.90	0.03	0.04	0.00	6.00	11.00	7.78	0.01	2.07	0.40	3.00	0.88	0.08	0.67
AND	MACK	0.84	0.03	0.09	0.05	8.00	7.00	6.80	0.02	1.49	0.34	2.46	0.87	0.10	0.65
AND	WS01	1.14	0.05	0.08	0.01	13	21.00	9.10	0.03	2.64	0.20	4.64	1.02	0.18	0.74
AND	WS02	1.47	0.03	0.03	0.00	6	22.00	9.00	0.04	2.54	0.38	3.47	0.82	0.12	0.90
AND	WS06	0.52	0.02	0.03	0.00	5	13.00	8.20	0.03	2.11	0.38	3.94	0.99	0.07	0.61
AND	WS07	0.53	0.02	0.02	0.00	4	22.00	9.32	0.03	2.32	0.58	4.07	1.12	0.05	0.60
AND	WS08	1.14	0.02	0.03	0.00	5	23.00	8.04	0.04	3.05	0.47	3.46	0.67	0.11	0.71

AND	WS09	1.77	0.04	0.05	0.00	6	17.00	8.90	0.03	2.67	0.17	4.69	1.35	0.19	1.31
AND	WS10	0.89	0.03	0.04	0.00	6	32.00	8.88	0.05	2.49	0.23	4.30	1.09	0.15	1.26
KNZ	edlr	0.80	-	-	0.06	16	-	-	-	-	-	-	-	-	-
KNZ	hikx	0.94	-	-	0.29	26	-	-	-	-	-	-	-	-	-
LUQ	Q1	0.72	0.05	0.20	0.14	5.34	1.00	34.41	-	8.58	0.91	5.52	3.11	1.28	8.35
LUQ	Q2	0.85	0.05	0.19	0.13	2.50	1.00	29.48	-	7.43	0.93	5.98	2.85	0.94	7.39
LUQ	Q3	0.80	0.05	0.15	0.11	2.50	1.00	28.42	-	7.56	0.72	4.68	3.07	0.64	8.36
LUQ	QG	0.90	0.04	0.11	0.07	2.50	1.00	15.25	-	5.27	0.45	2.76	1.07	0.41	6.30
LUQ	QP	1.30	0.06	0.12	0.06	2.50	1.00	19.11	-	6.24	0.33	4.35	3.27	0.48	8.18
LUQ	QS	1.58	0.07	0.14	0.08	2.50	1.00	11.10	-	4.72	0.28	2.49	1.47	0.57	7.17
LUQ	RI	0.99	0.04	0.15	0.09	7.29	1.00	19.21	-	5.40	0.55	3.57	1.24	0.43	6.17
LUQ	MPR	1.12	0.06	0.13	0.06	2.50	1.00	20.84	-	6.57	0.68	7.70	1.99	1.17	7.78

Table S3. Start and end date of time series for each stream along with the length of the data record. All records started in January and ended in December for every year analyzed for DOC and DON except CPC which started in May and ended in August for every year analyzed. – correspond to no data available.

Site	Stream	Start Year		End Year		Total record length	
		DOC	DON	DOC	DON	DOC	DON
LUQ	RI	2000		2015		15	
LUQ	MPR	1998		2015		17	
LUQ	QP	2000		2015		15	
LUQ	Q1	1998		2015		17	
LUQ	Q2	1998		2015		17	
LUQ	Q3	1998		2015		17	
LUQ	QG	2000		2015		15	
LUQ	QS	2000		2015		15	
AND	MACK	2003		2015		12	
AND	WS02	2003		2015		12	
AND	WS08	2004		2015		11	
AND	WS09	2004		2015		11	
AND	WS01	2004		2015		11	
AND	WS06	2003		2015		12	

AND	WS07	2003	2015	12		
AND	WS10	2004	2015	11		
AND	LOOK	2006	2015	9		
LMP	CB	2005	2016	11		
LMP	LMP73	2000	2016	16		
LMP	LTR	2004	2016	12		
LMP	MLB	2008	2016	8		
LMP	NBR	2004	2016	12		
LMP	NOR	2004	2016	12		
LMP	PWT	2004	2016	12		
LMP	RMB	2004	2016	12		
LMP	WHB	2003	2016	13		
FIN	Aur	1988	2015	27		
FIN	Eur	1991	2015	24		
FIN	Iij	1995	2015	20		
FIN	Kal	1975	1984	2012	37	31
FIN	Kel	1989	1975	2015	26	40
FIN	Kem	1991	1988	2015	24	27

FIN	Kii	1995	1975	2015	2012	20	37
FIN	Kis	1996	1989	2015		19	26
FIN	Kiv	1978	1991	1994	2015	16	24
FIN	Koj	1975	1995	2015		40	20
FIN	Kos	1994	1996	2015		21	19
FIN	Kot	1981	1978	2001	1994	20	16
FIN	K.Hur	1975		2015		40	
FIN	K.Ahv	1984	1994	2015		31	21
FIN	K.Kok	1988	1981	2015	2001	27	20
FIN	Lap	1996		2012	2015	16	19
FIN	Les	1998	1995	2015		17	20
FIN	Mur	1978		1994		16	
FIN	Oul	1975		2015		40	
FIN	Paj	1985		2015		30	
FIN	Per	1982		2015		33	
FIN	Pyh	1996		2015		19	
FIN	Sav	2006		2015		9	
FIN	Sii	1983		2015		32	

FIN	Sim	1991	2015	24	
FIN	Ska	1975	2015	40	
FIN	Tor	1991	2015	24	
FIN	Usk	1988	2015	27	
FIN	Val	1978	1994	16	
FIN	Vir	2003	2015	12	
FIN	Vuo	1989	2015	26	-
FIN	Ylij	1981	2001	20	-
HBF	W1	2006	2015	9	
HBF	W6	1996	2015	19	
HBF	W7	1996	2015	19	
HBF	W8	1996	2015	19	
HBF	W9	1996	2015	19	
KNZ	edlr	1994	2016	22	-
KNZ	hikx	1994	2016	22	-
CPC	C1	2002	2010	8	
CPC	C2	2002	2010	8	
CPC	C3	2002	2010	8	

CPC	C4	2002	2010	8
CPC	CB	2002	2010	8
CPC	CJ	2002	2010	8
CPC	P6	2005	2010	5
CPC	PC	2002	2010	8
CPC	PJ	2002	2010	8

Table S4. Sampling rates and analytical methods for analysis of dissolved organic carbon (DOC) or total organic carbon (TOC), total dissolved nitrogen (TDN), total nitrogen (TN) or dissolved organic nitrogen (DON), nitrate (NO_3^-), and ammonium (NH_4^+) concentrations for each site. – corresponds to no analysis performed.

Site	Sampling Frequency	Sampling and storage	TOC or DOC	TDN	TN or DON	NO_3^-	NH_4^+	Reference
AND	3-week composites	Samples composited over a 3-week period, collected weekly and refrigerated; filtered after arriving at the lab and analyzed within 2 days	DOC - combustion	Persulfate digestion and analysis by automated colorimetric, Technicon Auto-analyzer II	DON - mathematically determined by subtracting $\text{NH}_3\text{-N}$ and $\text{NO}_3\text{-N}$ from TDN	Colorimetric, automated cadmium reduction	Colorimetric, Technicon Auto-analyzer II	Martin and Harr 1998 Johnson et al. 2021
CPC	Daily to Bi-weekly	Daily samples were collected with autosamplers as a composite of samples collected at 6 pm and 6 am the following day	DOC - combustion Shimadzu TOC-5000	Shimadzu TOC-5000 plumbed to an Antek 7050 nitric oxide detector to quantify total dissolved nitrogen	DON - mathematically determined by subtracting $\text{NH}_3\text{-N}$ and $\text{NO}_3\text{-N}$ from TDN	Ion chromatography	Ion chromatography	Petrone et al. 2006
FIN	Monthly to Weekly	TOC - deep-frozen and analyzed within 1–3 months TN - analyzed on the day after sampling	TOC - oxidized to carbon dioxide by combustion and determined by infrared spectrometry; DOC - determined as 95% of TOC	–	TN - analyzed calorimetrically after oxidation with peroxodisulfate and reduction with Cd–Cu column; DON - determined as 95% of TN	–	–	Mattsson et al. 2005; Kortelainen et al. 2006

HBF	Weekly	Since June 1, 2013 samples are brought back to the Pierce Lab, and immediately filtered; Prior, samples were not filtered or frozen before analysis.	DOC - high-temperature catalytic oxidation technique	High-temperature catalytic oxidation technique	DON - mathematically determined by subtracting NH ₃ -N and NO ₃ -N from TDN	Ion chromatography	Colorimetric, automated indophenol- blue method	Buso et al. 2000 Campbell et al. 2021
LMP	Weekly	Filtered during collection and frozen until analyzed	DOC - high-temperature catalytic oxidation technique	High-temperature catalytic oxidation technique	DON - mathematically determined by subtracting NH ₄ -N and NO ₃ -N from TDN	Ion chromatography	Colorimetric, SmartChem 200 using the alkaline phenate	Coble et al. 2018 Wymore et al 2021
KNZ	Every other day	Frozen until analyzed	DOC - high temperature combustion	—	—	Colorimetric determination on a flow solution analyzer	Colorimetric determination on a flow solution analyzer	Kemp and Dodds 1998
LUQ	Weekly	Filtered during collection and frozen until analyzed	DOC - high-temperature catalytic oxidation technique	High-temperature catalytic oxidation technique	DON - mathematically determined by subtracting NH ₄ -N and NO ₃ -N from TDN	Ion chromatography	Colorimetric, SmartChem 200 using the alkaline phenate	Merriam et al. 2002 McDowell et al. 2021

Table S5. References and repositories for each sites data set

Program	Dataset	URL or DOI	Reference if applicable	Contact	Email
LTER	AND Oregon, USA	https://doi.org/10.6073/pasta/bb935444378d112d9189556fd22a441d	Johnson, S.L and R.L. Fredriksen. 2019. Stream chemistry concentrations and fluxes using proportional sampling in the Andrews Experimental Forest, 1968 to present ver 23.	Sherri Johnson	sherri.johnson2@usda.gov
LTER	CPC, Alaska, USA	https://doi.org/10.6073/pasta/257da36f3edb8df2976696f98a219b7d	Jones, J., F.S. Chapin, R.W. Ruess, and Bonanza Creek LTER. 2014. Stream water chemistry of CPCRW, 2002-2010 ver 20. Environmental Data Initiative.	Jeremy Jones	jay.jones@alaska.edu
LTER	HBF New Hampshire, USA	Submitted by PI or primary contract	NA	Emily Bernhardt	emily.bernhardt@duke.edu
LTER	LUQ Puerto Rico	https://doi.org/10.6073/pasta/f9df56348f510da0113b1e6012fa2967	-McDowell, W. H. 2017. Chemistry of stream water from the Luquillo Mountains ver 4923051. Environmental Data Initiative. -W. H. McDowell, Leon, C. L., Shattuck, M. D., Potter J. D., Heartsill-Scalley, T., González, G., Shanley, J. B., and Wymore, A. S. 2021. The Luquillo Experimental Forest: catchment science in the montane tropics. Hydrological Processes.	William H. McDowell	bill.mcdowell@unh.edu
LTER	KNZ Kansas, USA	Submitted by PI or primary contract	NA	Walter K. Dodds	wkdodds@ksu.edu
Finnish Environment Institute (SYKE)	FIN Finnish watersheds (multiple locations)	https://www.ymparisto.fi/scripts/oiva.asp	NA	Pirkko Kortelainen	pirkko.kortelainen@syke.fi
NA	LMP Lamprey River Hydrological Observatory, New Hampshire, USA	Submitted by PI or primary contract	Wymore, A.S., M. D. Shattuck, J.D. Potter, L. Snyder, and W. H. McDowell. (2021a). The Lamprey River Hydrological Observatory: suburbanization and changing seasonality. Hydrological Processes.	William H. McDowell	bill.mcdowell@unh.edu