

OPINION

Informing Nature-based Climate Solutions for the United States with the best-available science

Kimberly A. Novick¹  | Stefan Metzger²  | William R. L. Anderegg³  |
 Mallory Barnes¹  | Daniela S. Cala¹  | Kaiyu Guan^{4,5}  | Kyle S. Hemes⁶  |
 David Y. Hollinger⁷  | Jitendra Kumar⁸  | Marcy Litvak⁹  | Danica Lombardozzi¹⁰  |
 Caroline P. Normile¹¹  | Patty Oikawa¹²  | Benjamin R. K. Runkle¹³  |
 Margaret Torn¹⁴  | Susanne Wiesner¹⁵ 

¹O'Neill School of Public and Environmental Affairs, Indiana University—Bloomington, Bloomington, Indiana, USA

²Battelle, National Ecological Observatory Network, Boulder, Colorado, USA

³School of Biological Sciences, University of Utah, Salt Lake City, Utah, USA

⁴College of Agricultural, Consumer and Environmental Sciences, University of Illinois at Urbana-Champaign, Urbana, Illinois, USA

⁵National Center for Supercomputing Applications, University of Illinois at Urbana-Champaign, Urbana, Illinois, USA

⁶Woods Institute for the Environment, Stanford University, Stanford, California, USA

⁷USDA Forest Service, Northern Research Station, Durham, New Hampshire, USA

⁸Environmental Sciences Division, Oak Ridge National Laboratory, Oak Ridge, Tennessee, USA

⁹Department of Biology, University of New Mexico, Albuquerque, New Mexico, USA

¹⁰National Center for Atmospheric Research, Boulder, Colorado, USA

¹¹Bipartisan Policy Center, Washington, D.C., USA

¹²Department of Earth & Environmental Science, California State University—East Bay, Hayward, California, USA

¹³Department of Biological and Agricultural Engineering, University of Arkansas, Fayetteville, Arkansas, USA

¹⁴Lawrence Berkeley National Laboratory, Berkeley, California, USA

¹⁵Department of Biological Systems Engineering, University of Wisconsin-Madison, Madison, Wisconsin, USA

Correspondence

Kimberly A. Novick, O'Neill School of Public and Environmental Affairs, Indiana University—Bloomington, Bloomington, IN, USA.

Email: knovick@indiana.edu

Funding information

National Science Foundation, Grant/Award Number: 1552747 and 1752083; Indiana University Bloomington, O'Neill School of Public and Environmental Affairs; U.S. Department of Energy, Office of Science; Stanford Woods Institute for the Environment

Abstract

Nature-based Climate Solutions (NbCS) are managed alterations to ecosystems designed to increase carbon sequestration or reduce greenhouse gas emissions. While they have growing public and private support, the realizable benefits and unintended consequences of NbCS are not well understood. At regional scales where policy decisions are often made, NbCS benefits are estimated from soil and tree survey data that can miss important carbon sources and sinks within an ecosystem, and do not reveal the biophysical impacts of NbCS for local water and energy cycles. The only direct observations of ecosystem-scale carbon fluxes, for example, by eddy covariance flux towers, have not yet been systematically assessed for what they can tell us about NbCS potentials, and state-of-the-art remote sensing products and land-surface models are not yet being widely used to inform NbCS policymaking or implementation. As a result, there is a critical mismatch between the point- and tree-scale data most often used to assess NbCS benefits and impacts, the ecosystem and landscape scales where NbCS projects are implemented, and the regional to continental scales most

relevant to policymaking. Here, we propose a research agenda to confront these gaps using data and tools that have long been used to understand the mechanisms driving ecosystem carbon and energy cycling, but have not yet been widely applied to NbCS. We outline steps for creating robust NbCS assessments at both local to regional scales that are informed by ecosystem-scale observations, and which consider concurrent biophysical impacts, future climate feedbacks, and the need for equitable and inclusive NbCS implementation strategies. We contend that these research goals can largely be accomplished by shifting the scales at which pre-existing tools are applied and blended together, although we also highlight some opportunities for more radical shifts in approach.

KEYWORDS

climate adaptation, climate mitigation, ecosystem carbon cycling, natural climate solutions, net-zero

1 | OVERVIEW

Terrestrial ecosystems, which sequester about a third of anthropogenic CO₂ emissions (Friedlingstein et al., 2020), have long been studied for their outsized role in mitigating the pace of climate warming (Baldocchi, 2008; Churkina & Running, 1998; Torn & Chapin, 1993). As climate change impacts become more pronounced, and the need to remove CO₂ from the atmosphere becomes more urgent, support is growing for the notion that ecosystems could be actively managed to increase carbon sequestration or reduce greenhouse gas (GHG) emissions (Griscom et al., 2017; Nolan et al., 2021; Seddon et al., 2020). These Nature-based Climate Solutions (NbCS) are not a panacea for climate change mitigation (Anderson et al., 2019), and absolutely cannot be effective without concurrent and dramatic economy-wide decarbonization. Even in the best-case scenarios, NbCS will contribute only a fraction of the emissions reductions necessary to limit warming to <2°C. Nonetheless, removing CO₂ from the atmosphere is part of nearly all net-zero pathways (Rogelj et al., 2018), and NbCS may offer low-cost mitigation along with co-benefits such as improved air and water quality, better soil health, biodiversity maintenance (Fargione et al., 2018), and local climate adaptation (Osaka et al., 2021).

In the United States, intentional implementation of NbCS has been relatively limited, and largely organized around private voluntary carbon markets (Anderegg, 2021; Seddon et al., 2021, but see CNRA, 2021), which offer the promise of revenue streams for landowners and for private entities focused on project development and monitoring. Forest carbon offset projects in California's compliance system perhaps represent a more systematic attempt at coordinated NbCS implementation (Anderegg et al., 2020), though the actual mitigation achieved through these projects is not clear (Badgley et al., 2022). However, looking forward, state- and federal agencies appear poised to authorize large investments in NbCS programs (Fargione et al., 2018; Fleischman et al., 2020; Seddon et al., 2020). For example, the U.S. Senate passed the "Growing Climate Solutions

Act" in 2021, and in early 2022, the USDA released a \$1 billion call for proposals for "Climate-Smart Commodities." Indeed, it is an unusual coalition, including conservation groups, farmers, foresters, bipartisan groups of lawmakers, and private start-ups and industry, that is driving momentum in the NbCS sphere.

Despite this enthusiasm, the realizable benefits of NbCS are not well understood and often difficult to quantify (Seddon et al., 2020). They are usually estimated as a change in carbon stocks determined from biometric soil or tree survey data (Cook-Patton et al., 2020; Griscom et al., 2017). These surveys, however, can miss changes in stocks that are unmeasured or hidden by landscape heterogeneity, and do not provide information about methane and nitrous oxide emissions or concurrent biophysical impacts on temperature and water cycling. Moreover, for many NbCS, existing biometric data are sparse and unrepresentative of naturally occurring gradients in soil and climate. As a result, there is a critical mismatch in scale between the biometric data most often used in NbCS accounting, the ecosystem, and landscape scales where NbCS projects are implemented, and the regional to continental scales at which relevant policies are developed.

The situation becomes even more poorly constrained looking forward, when climate-driven feedbacks threaten the permanence of carbon stored in many ecosystems. In forests, the large amount of carbon stored in aboveground plant biomass is threatened by increasing drought, wildfires, and insect outbreaks (Anderegg et al., 2020; Coffield et al., 2021); in soils, warming can stimulate decomposition and CO₂ fluxes (Hicks Pries et al., 2017). The Earth System Models (ESMs, Heavens et al., 2013) used to couple interactions between ecosystems and the climate system account for these feedbacks, but the simpler models used for NbCS benefit evaluation do not. Finally, all this uncertainty propagates into operational barriers hindering NbCS project implementation within carbon markets. That process generally relies on statistical models and biometric soil and tree survey data collected over relatively long timescales (typically 5+ years). The resource-intensive nature of biometric inventories,

along with the relatively low price of carbon, practically excludes all but the largest non-tenant producers and landowners from participating. The approach also exposes the system to risks associated with unduly optimistic assessments of project benefits or practice implementation (Badgley et al., 2022).

Our objective is to identify knowledge gaps surrounding NbCS that may be confronted, over the short term, with pre-existing data, infrastructure, and tools that have long been used to measure and predict ecosystem-scale GHG exchanges, but have not yet been harnessed for what they reveal about NbCS effectiveness. We contend that new perspectives on NbCS climate benefits and unintended consequences can be largely enabled by relatively subtle shifts in the scales at which these existing tools are applied (e.g., “modular innovation”) and blended together (e.g., “architectural innovation, *sensu* Henderson and Clark, 1990, see Figure 1). However, for some uncertainties, and especially those surrounding NbCS permanence, more “radical” shifts in approach may be required. Collectively, the perspectives presented here could function as a proposal describing the work needed to inform NbCS assessments with the best-available science.

2 | NBCS-RELEVANT DATA AND ANALYTICAL TOOLS

The dominant role of terrestrial ecosystems in determining the fate of atmospheric CO₂ has been known for decades. Consequently, huge investments of resources have fostered the development of innovative tools for monitoring and quantifying ecosystem carbon cycles (Figure 2). These include tools for quantifying both carbon stocks (or the amount of carbon stored in soil, litter, or plant biomass) and carbon fluxes (which represent the rates at which carbon is transferred into, out of, and within ecosystems). While fluxes can be inferred from the change in stocks over time, this approach only works if all the relevant stocks are measured (which is often infeasible). Moreover, carbon stock data alone are insufficient to reveal the processes responsible for a change in carbon uptake, or to estimate emissions of non-CO₂ greenhouse gases like methane and N₂O. Right now, the vast majority of the existing approaches

for flux measurement and modeling are not being widely applied to NbCS evaluations. Particularly striking is the fact that the only direct observations of land-atmosphere carbon, water, and energy exchanges (from flux towers, Baldocchi, 2008) have not yet been systematically assessed for what they can tell us about NbCS impacts (Hemes et al., 2021). Likewise, many next-generation remote sensing products and state-of-the-art process-based models are also not being widely used to inform NbCS policymaking or implementation.

A general tradeoff exists between the accessibility of these tools to broad communities of stakeholders, and the robustness with which they describe a full set of relevant ecosystem processes (Figure 3). Biometric soil core and tree survey data are simple, low-cost measurements that are broadly accessible; however, their robustness is limited, as they do not account for all carbon stocks, provide little information about biophysical impacts, and have a low temporal resolution that limits their ability to detect changes quickly. In contrast, flux towers have a high degree of “robustness” linked to their ability to continuously measure the net flux of CO₂ (and other GHGs) between the atmosphere and the ecosystem, as well as a full suite of related water and energy cycle variables. But flux towers are expensive, and quality control and post-processing of flux tower data has historically required specific expertise. Satellites and drones provide spatially robust proxies for NbCS-relevant variables at scales that are increasingly well matched to farms and fields. However, the temporal resolution of these products is often limited, and no technology yet exists to measure the net flux of CO₂ or other GHGs directly from space.

The extent to which data are “open” and discoverable to a wide range of researchers and stakeholders is another dimension of accessibility. Flux tower networks such as AmeriFlux, FLUXNET, and NSF's National Ecological Observatory Network (NEON) have long been on the forefront of open data sharing (Baldocchi, 2008; Metzger et al., 2019; Novick et al., 2018). Tree inventory data from the USDA Forest Inventory and Analysis (FIA) program (Bechtold & Patterson, 2005) are also highly standardized and accessible, and soil carbon data are also becoming more aggregated and open (Bond-Lamberty et al., 2020; Malhotra et al., 2019). However, by and large, these networks are not well connected to each other digitally,

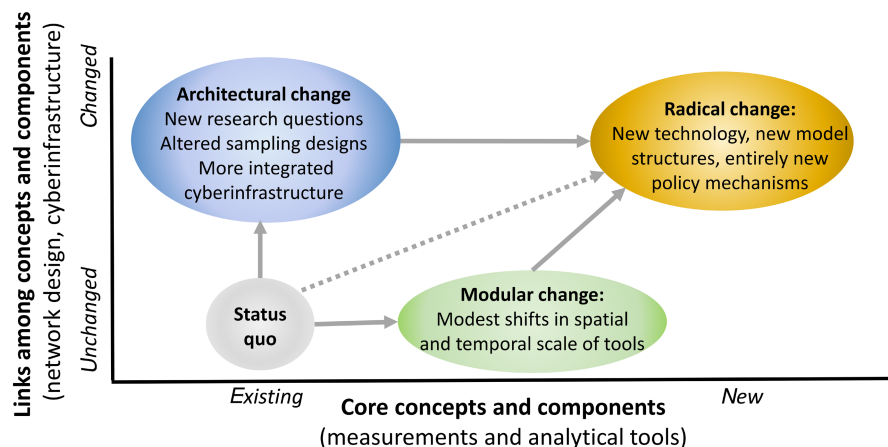


FIGURE 1 Classes of innovation characterized by changes to the tools themselves (x-axis) or changes to the linkages among the tools (y-axis), modified from Henderson and Clark (1990). The bulk of this paper is focused on modular and architectural change, accompanied by some promising directions for “breakthrough” research through radical change [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/doi/10.1111/gcb.16156)]

FIGURE 2 Existing tools and approaches for quantifying carbon pools and GHG emissions from terrestrial ecosystems. See Supplementary Information (S.I.) for detailed description of each tool [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/doi/10.1111/gcb.16156)]

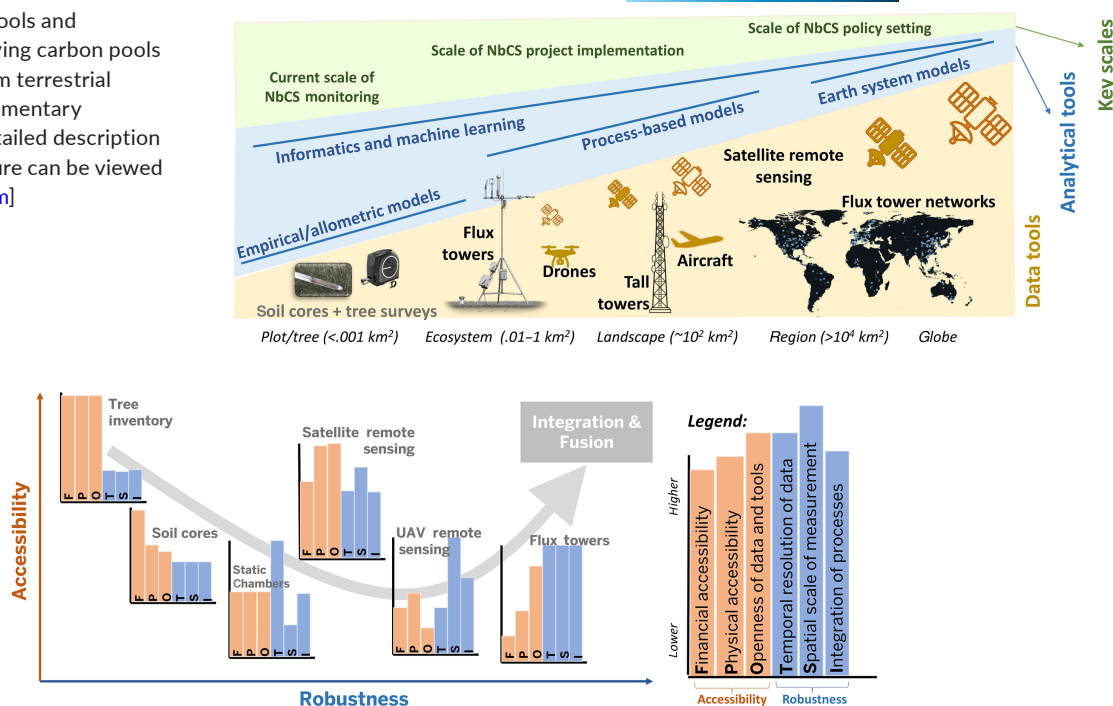


FIGURE 3 The data tools relevant for NbCS differ along many dimensions of accessibility and robustness. Accessibility dimensions include the following: **financial accessibility**, which is inversely related to cost, **physical accessibility** which describes the ease with which data can be physically obtained, and the **openness** of the tool, representing the extent to which data and algorithms are findable and usable. Dimensions of robustness include the following: **temporal resolution** of the measurements, with more frequent observations enabling faster detection of NbCS impacts and better attribution to mechanisms; **spatial scale** of the measurements, and specifically the extent to which the measurement is “ecosystem-scale,” and biophysical **process** robustness in terms of whether the approach integrates information on how NbCS may affect not only carbon pools, but also other GHGs and local biophysics [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/doi/10.1111/gcb.16156)]

and with limited physical overlap between network sites that hinders synthesis (Hinckley et al., 2016).

In summary, no single approach is a perfect tool for assessing the realizable impacts of NbCS. Thus, throughout the rest of this manuscript, we will emphasize the need for standardized collection of multiple data streams, and outline strategies for fusing these data together to maximize their collective accessibility and robustness while minimizing the unique limitations of each tool.

3 | INFORMING NBCSS WITH A FULL SET OF TOOLS AND APPROACHES

3.1 | NbCS assessments at policy-relevant scales

Policymakers and stakeholders seek regional- to global-scale assessments of the expected mitigation potential of NbCS, including information about when and where a given approach is most likely to succeed. Ideally, these assessments fulfill the following criteria: (1) they are *informed* by observations of land-atmosphere GHG fluxes *made directly at the ecosystem scale* ($\sim 1 \text{ km}^2$), thereby integrating over multiple aboveground and belowground GHG sources and sinks; (2) they are *spatially resolved* (e.g., mapped) and describe *where* the benefits of a given NbCS are greatest; and (3) they are *forward looking*, with careful consideration of the durability of benefits into a future

characterized by pervasive climate feedbacks. Right now, a wide gulf separates available information from these idealized criteria. The following subsections highlight ways that flux tower data, survey data, remote sensing data, and models can be used together to narrow the gap. An emergent theme will be the need for “gold-standard datasets” to support a wide array of NbCS assessment and validation goals. We imagine these datasets would represent standardized, open and accessible observations of a full suite of carbon stock and flux measurements, from NbCS “treatments” as well as baseline controls, together with information about historic land use. Sustaining long-term flux tower data records should also be a research priority, since substantial knowledge gaps remain surrounding the extent to which ecosystem carbon uptake may “saturate” in time (Craig et al., 2021; Curtis & Gough, 2018, and see additional text in the S.I.).

3.1.1 | Systematic evaluation of ground-based observations

Forests

In forests, which contain a variety of species at different stages of growth, the ability of flux towers to integrate over all carbon sources and sinks is extremely useful. Multi-year time series from >150 flux towers located in forests are already available from AmeriFlux; of these located in the United States, several dozen represent forests or

tree-dominated savannas co-located with grasslands or croplands (Figure 4), offering an opportunity for a first-order, ecosystem-scale assessment of the realizable mitigation potential of reforestation as an NbCS. However, the relatively high costs of building and maintaining tall forest towers will always limit their spatial representativeness. In contrast, tree survey data are relatively abundant, because of programs like FIA, and may be adequate in some regions to quantify spatial patterns in aboveground biomass (Hemes et al., 2021). However, tree surveys have long (5+ year) sampling intervals, and do not capture patterns of belowground carbon cycling and storage. Comparing tower-based carbon fluxes with estimates derived from biometric survey data from the same site (e.g. Campioli et al., 2016; Wang et al., 2017) can be useful for understanding the biases in carbon uptake potential informed by biometric tree survey data alone. Adding routine biometric sampling (soil cores, tree surveys) to active forest flux tower sites would represent a relatively low-cost initiative that could form the foundation of “gold-standard” datasets for forested NbCS.

Strategic deployment of new forest flux towers may be necessary, especially for NbCS focused on improved forest management,

given that we still lack a clear picture for how carbon uptake varies as a function of forest age (Amiro et al., 2010; Curtis & Gough, 2018; Law et al., 2003; Novick et al., 2015). Moreover, with some exceptions (Gough et al., 2021), flux towers deployed over forests experiencing similar climate, but different management regimes, are largely absent from the networks. This knowledge gap is important to fill to constrain the potential of improved forest management as an NbCS.

Croplands

Because they are already intensely managed, croplands represent relatively “low hanging fruit” for NbCS implementation. Climate mitigation benefits of croplands are largely constrained to soil carbon pools, as the majority of aboveground biomass is removed by harvest. While soil carbon is theoretically easy to measure, creating sampling strategies that adequately capture horizontal and vertical heterogeneity in soil carbon, and its change over time, remains difficult and expensive (Smith et al., 2020). Critically, soil carbon data alone are insufficient to identify responsible mechanisms, such as greater carbon uptake from photosynthesis versus reduced carbon

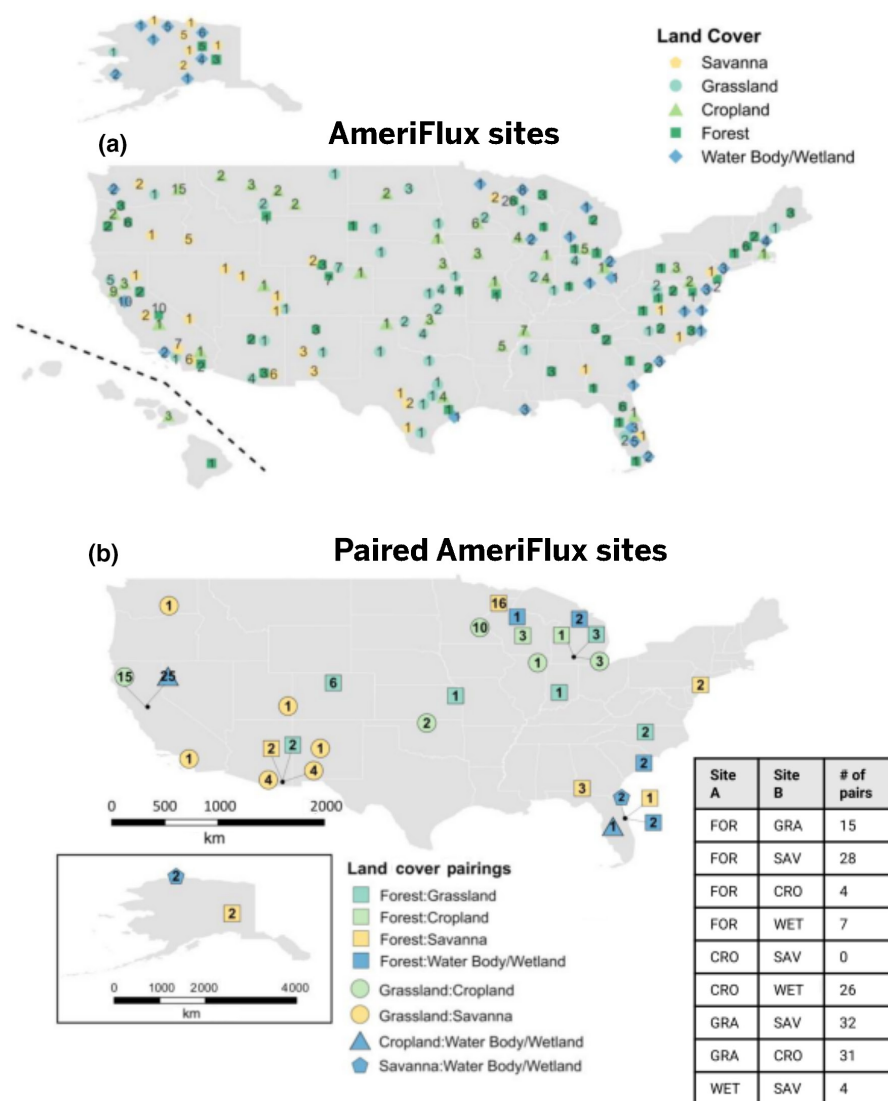


FIGURE 4 The spatial distribution of existing AmeriFlux towers in the United States (those that have registered with the network) classified by biome, with numbers indicating the number of towers of each biome type located within ~30 km (panel a). Panel (b) highlights the location of “paired sites” representing flux towers in different biomes that are within ~30 km. The table describes the number of specific categories of site pairs. Abbreviations: FOR, forest; GRA, grassland; SAV, savanna; CRO, cropland; WET, wetland or water body [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/terms-and-conditions)]

loss in runoff. Concerningly, while soil carbon data tend to report a soil sequestration benefit from cover cropping (Poeplau & Don, 2015), at least one study leveraging flux tower data reports that cover crops do not favorably impact net carbon uptake (Baker & Griffis, 2005). Moreover, because empirical studies reporting on soil carbon changes are limited for many categories of NbCS, spatially explicit maps of cropland NbCS mitigation potentials do not yet exist. Consequently, we do not know where climate conditions favor or disfavor these strategies.

New pilot flux tower studies that pair an NbCS treatment with a conventionally managed field could bring many insights (Hemes et al., 2021). In theory, flux towers are easier and cheaper to operate in ecosystems with short (<3 m) vegetation, although running them alongside active farm operations and on fast-growing crops can be operationally challenging. Because flux towers cannot detect lateral fluxes out of the measurement footprint, the outflow of dissolved and particulate carbon in runoff should also be monitored, which is relatively easy in the tile-drained systems that characterize much of the Corn Belt. Changes to the leaching of carbon through outflow (Nakhavali et al., 2021) may be an important factor that can cause an increase in soil C that does not necessarily reflect a climate benefit, but rather a tradeoff between GHG emissions in the field and distal emissions downstream. Amending the sampling design around existing cropland flux towers with soil carbon monitoring, outflow monitoring, and static chambers is a relatively straightforward path for creating “gold-standard” datasets for cropland NbCS.

Wetlands and coastal systems

Inundated and/or saline conditions provoke a decline in carbon mineralization and offer an opportunity for enhanced soil storage of carbon. Tidal wetland restoration is an especially promising wetland solution (Kroeger et al., 2017; Fargione et al., 2018), and together with seagrass restoration (or avoided loss), represents significant potential for coastal landscapes (~25 Tg CO₂e year⁻¹). Away from the coasts, riparian zone and peatland restoration (Günther et al., 2020; Vermaat et al., 2021), and methane emissions reduction in rice (Runkle et al., 2019) represent additional opportunities for managed wetlands to contribute to climate solutions. A particular challenge for NbCS is optimizing carbon uptake and sequestration of existing soil carbon against possible production of CH₄ and N₂O (Hemes et al., 2018; Rosentreter et al., 2021; Valach et al., 2021) and biophysical effects (Lee et al., 2011).

Flux towers are well-positioned to assess these impacts, as they enable measurement of complementary gases (CH₄, N₂O) at a high temporal resolution that enables detection and interpretation of spikes (often called “hot moments”) of gas release associated with sudden changes in water levels or biological conditions (Turner et al., 2021). They must be placed alongside estimates of lateral carbon flows (Arias-Ortiz et al., 2021; Bogard et al., 2020) and then analyzed in concert with tidal or water flow data. Like many agricultural sites, the shorter vegetation in these landscapes may reduce some costs, and site management may be conducive to paired or clustered site

experimentation. Indeed, more sites are needed to capture the impact of different hydroperiods, vegetation, and biogeochemistry (Matthes et al., 2014); blue carbon flux sites are also only one-third as prevalent as forest, agriculture, or grassland sites (Hemes et al., 2021).

3.1.2 | Blending flux tower data with state-of-the-art remote-sensing observations

Remote sensing data are indispensable for extending ground-based observations to scales relevant for policymaking. Already, remotely sensed proxies for aboveground biomass are being used to map forest carbon stocks (Rodríguez-Veiga et al., 2017), and next-generation laser and radar missions (Ustin & Middleton, 2021) will produce high-resolution (25–200 m) and three-dimensional biomass estimates of the world's forests. In croplands, remote sensing is already proving useful for detecting the presence/absence of NbCS-relevant management practices like cover crops and no-till management regimes (Azzari et al., 2019; Barnes, Yoder, & Khodaei, 2021), as well as crop yield (Guan et al., 2017)—a major carbon cycle “flux.” However, while maps of carbon stocks and practice adoption rates provide useful information for NbCS policy evaluation, they are not the same as maps of the potential of a management practice to avoid emissions and enhance sequestration. The change in remotely sensed biomass over time can be blended with allometric equations to infer the flux of CO₂ from the atmosphere to aboveground vegetation (Quegan et al., 2019; Rodríguez-Veiga et al., 2017). But these approaches are generally only possible for forest ecosystems, and they suffer from the same biases affecting biometric tree surveys.

Progress toward spatially explicit maps of NbCS mitigation potential could be enabled by a growing suite of spaceborne instruments, the so-called “Flux Towers in the Sky” (Schimel, D., Schneider, F.D. and JPL Carbon and Ecosystem Participants, 2019), which can sense key aspects of plant function. These next-generation platforms include (a) solar-induced fluorescence (SIF), which is physiologically related to the rate of photosynthesis (Magney et al., 2020), (b) column-averaged atmospheric CO₂ which can be used for “inverse” estimates of land carbon fluxes (Wang et al., 2019), and (c) instruments for sensing ecosystem water stress (e.g., ECOSTRESS, Fisher et al., 2020, and microwave data on canopy water content, Konings et al., 2021). While the spatial resolution of these satellite products can be coarse, some are now available at scales that match those of individual farms (e.g., ECOSTRESS, Fisher et al., 2020), and the need for finer-scale versions of other products has been clearly articulated (Konings et al., 2021). In many cases, substantial increases in resolution are possible with drone-mounted instruments (e.g., SIF, Mohammed et al., 2019). The information provided by these platforms is sensitive to limitations and biases, which are well-reviewed elsewhere (Fisher et al., 2020; Konings et al., 2021; Magney et al., 2020), and will continue to benefit from validation with flux tower data (e.g. Fisher et al., 2020; Sun et al., 2017), including novel strategies for fusing data across different scales of observation (see Section 3.3).

Next-generation remote sensing products can also enable machine learning (ML) extrapolation, or upscaling, of flux tower data into gridded maps (Figure 5), with the FLUXCOM project representing the most notable example (Jung et al., 2019). FLUXCOM is informed by data from flux towers spanning many continents and biomes, and provides an important constraint on global land carbon uptake generally. However, regional discrepancies exist, which can be partially explained by the representativeness of the flux tower data ingested into the ML algorithms (Jung et al., 2019). Several opportunities exist to refine ML methods so they can be used to map NbCS potentials directly for specific regions or biomes. Tower and remote-sensing data for a specific ecosystem type from a specific region (e.g., Eastern US temperate forests, or conventionally managed croplands in the Corn Belt) could be ingested into ML algorithms to produce regional (as opposed to global) “baseline” maps, which could be compared against each other or against data from pilot studies of novel NbCS “treatments.” These targeted, regional-scale mapping exercises could also be further guided by ecosystem-scale understanding of the locally important environmental drivers (e.g., Barnes, Farella, et al., 2021). Cyberinfrastructure that links “cut-outs” of remote-sensing products with flux tower data at a given site could be an important feature of “gold-standard” NbCS verification datasets.

3.1.3 | Models

Spatially explicit and forward-facing NbCS assessments are not possible without the use of predictive models. Indeed, models are already used to interpolate ground observations into regional-scale potential maps (e.g., for forest biomass, see Section 3.1.2) and to prescribe the value of project-scale market credits. But the models used for these objectives are highly empirical (e.g., regression based),

relying on observed relationships between driver and response variables that cannot be extrapolated into a future characterized by climate conditions profoundly different than those experienced historically.

Earth system models (ESMs), which predict future climate states for a range of anthropogenic emission scenarios, are currently the only tool for mechanistic prediction of climate–ecosystem feedbacks, and the only way to estimate the net effect of the combined physical and biogeochemical impacts into the future. The land component of ESMs—“Terrestrial biosphere models” (TBMs)—come in many flavors (Fisher et al., 2018), but they are generally constrained by fundamental conservation laws and rely on biogeochemical and biophysical theory to predict flows of carbon, water and other elements through the natural world (see Supplementary Information, hereafter S.I., for an extended discussion).

TBMs have not yet been widely applied to assess NbCS impacts, which may relate to the fact that TBMs were initially developed to transfer fluxes of energy, moisture, and momentum to the atmosphere, with prognostic carbon cycling largely developed in the 2000s (e.g. Cox et al., 2000; Fung et al., 2005). The inclusion of management-relevant processes, including land use change, agriculture, and nutrients, came even more recently (Fisher & Koven, 2020). The models are still limited by their capacity to represent management and disturbance processes, and in their skill at quantifying avoided emissions of non-CO₂ GHGs in agriculture. Addressing these limitations is an active research field. For example, mechanistic representation of species demographics and climate-sensitive disturbances like fire are rapidly being implemented in TBMs (Fisher et al., 2018). With respect to agricultural systems, several TBMs now include a basic, but coarse, representation of agricultural and pasture management (Lombardozzi et al., 2020; Pongratz et al., 2018), and TBMs have been used to explore coarse-scale tradeoffs and unintended consequences associated with managed land cover change

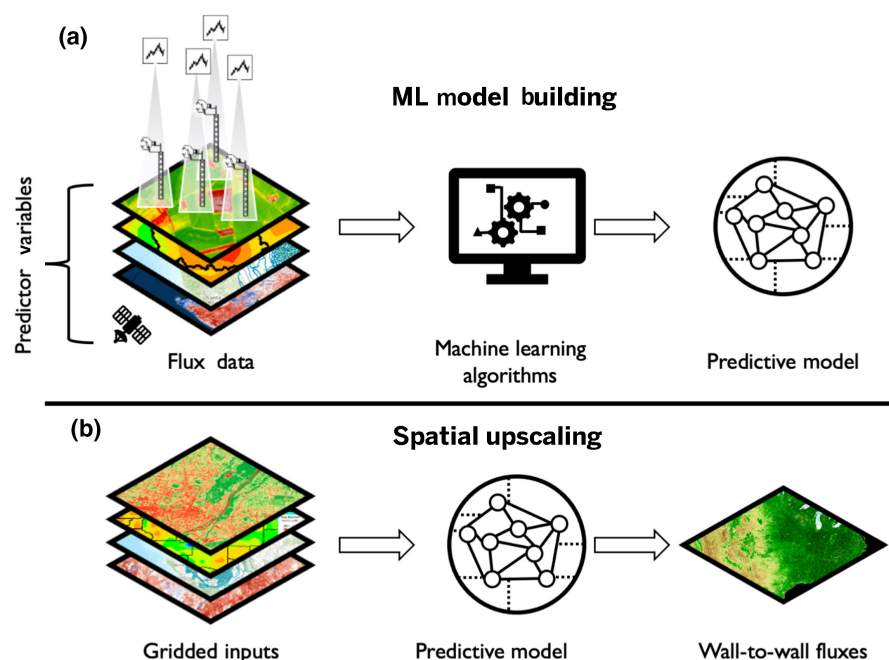


FIGURE 5 Machine-learning (ML) upscaling of eddy covariance fluxes. (a) To upscale fluxes, a ML algorithm first “learns” relationships between tower fluxes and input variables (i.e., remotely sensed metrics and meteorological data). (b) Then, wall-to-wall flux estimates are generated by applying the predictive ML model to each pixel of spatially continuous input variables [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/terms-and-conditions)]

(DuVeiller et al., 2020; Harper et al., 2018). Thus, despite their limitations, TBMs are very useful for general assessments of when and where NbCS are likely to be most effective (see, e.g., Graham et al., 2021; Harper et al., 2018).

However, substantial gaps must be addressed before process-based models can be fully applied to the many pressing sources of NbCS uncertainty, and in particular uncertainties linked to the spatial resolution of the models and their ability to predict the permanence of NbCS benefits. Right now, TBM spatial resolution is typically too coarse to resolve the field and farm scales where carbon credits are assessed and monitored. Moreover, future projections of land carbon uptake are very uncertain in ESMs, particularly into the latter half of the 21st century (Arora et al., 2020). Put simply, the models do not agree on the magnitude, and in some cases the direction, of future land-carbon uptake at the global scale (Friedlingstein et al., 2014). This fundamentally large and potentially irreducible uncertainty (Bonan & Doney, 2018) poses major challenges for predicting NbCS permanence. While radical changes to model structure and parameterization may help, several very pertinent questions remain relatively unexplored: First, will the uncertainty problem be reduced when models are tasked with predicting the *change* in land carbon uptake driven by a specific NbCS approach, as opposed to the absolute magnitude thereof? Can this uncertainty be priced into the market systems? And to what extent is model agreement improved when assessed at landscape and regional (as opposed to global) scales? Progress on the latter question may be facilitated by model-data assimilation approaches for near-term “ecological forecasting” (Dietze, 2017) and landscape scale model-data fusion (see Section 3.3).

3.2 | Toward generalizable frameworks for assessing biophysical co-benefits and/or unintended consequences

Ecosystem carbon uptake is closely coupled with ecosystem water use such that a managed alteration to land cover designed to affect C cycling will also affect the local hydrology. In general, greater C uptake will likely be associated with greater evapotranspiration (or ET); whether or not this is a favorable biophysical impact depends on climate regime, time of year, and management intent. For example, an increase in ET in spring may be welcomed by farmers throughout much of the Corn Belt, when the primary soil water problem is usually one of overabundance (e.g., flooding, Yin et al., 2020). On the other hand, when and where soil moisture deficits are common, NbCS-driven increases in ET that further deplete soil moisture and runoff may be undesirable. With some exceptions (e.g. Jackson et al., 2005; Windisch et al., 2021), systematic assessments of tradeoffs between NbCS carbon benefits and water cycle consequences are rare, and generally not interpreted in the context of predicted future changes in precipitation and soil moisture balance.

Land cover and management shifts also affect local energy budgets, not only by impacting ET, but also by modifying albedo and

sensible heat fluxes. The interplay between these mechanisms can cause NbCS strategies in some regions to cool the surface (e.g., tropical and temperate zone reforestation, Ge et al., 2019; Windisch et al., 2021; Zhang et al., 2020; wetland restoration, Hemes et al., 2018, and conversion to frequently flooded agriculture lands, Liu et al., 2019). In other cases (e.g., semi-arid and boreal forests), the radiative impacts of NbCS may lead to additional warming (Duman et al., 2021; Lee et al., 2011). Since temperature is rising everywhere, surface cooling relative to the baseline will usually represent a favorable biophysical impact, and some NbCS may represent a tool for local climate adaptation in addition to global climate mitigation. However, several gaps in our understanding of NbCS impacts on local temperature remain, including on the relationship between surface and air temperature (Schwingshackl et al., 2017) and the dynamics of both during climate extremes like heat waves (Teuling et al., 2010).

Substantial opportunity exists to leverage pre-existing data in networks like AmeriFlux for synthetic assessments of carbon and biophysical impacts of NbCS, since they measure most terms of the water and energy cycle. Moreover, unlike carbon uptake, direct quantification of ET and land surface temperature is possible from remotely sensed data (Fisher et al., 2020) such that more precise mapping of present-day NbCS biophysical impacts should be relatively straightforward, especially when flux tower network data are leveraged for ground-truthing. High-frequency flux tower data could be more carefully analyzed for what they reveal about biophysical impacts at sub-seasonal scales, including hot summer days when cooling benefits are needed most. Finally, emerging approaches that leverage flux tower data to understand land cover change impacts on air temperature (e.g., Helbig et al., 2021; Novick & Katul, 2020) can be more widely deployed, noting that near-surface air temperature is arguably the more important target from a climate adaptation perspective.

3.3 | Accessible and robust market-relevant quantification of project-scale impacts

Balancing accessibility and robustness is a pivotal challenge facing quantification strategies for NbCS projects, typically implemented at scales <100 km². Assessments that forgo direct measurement may enhance accessibility to landowners but run the risk of over- or under-quantifying the true climate benefits, eroding trust in NbCS claims or missing an opportunity to finance important activities (Günther et al., 2018). The most robust quantification—one that would require frequent physical sampling of each carbon pool over much of the project area—is a Sisyphean task, and will make quantification operationally and economically inaccessible to the vast majority of landowners. The appropriate balance between accessibility and robustness will vary among ecosystem and NbCS project types, scales, policy requirements, and the acceptable level of uncertainty.

In practice, the typical approach to quantifying NbCS benefits relies on periodically inventorying small changes to large carbon

stocks, and differencing these from the carbon stocks that would have been present in a baseline case. The latter is usually estimated with empirical models and without consideration of climate feedbacks. For simplicity, most methodologies conservatively omit consideration of less prominent carbon pools when accounting would lead to greater avoided emissions or removals. This status quo approach has the advantage of relying on established tools, but usually omits others (e.g., flux towers) that offer a more robust perspective on the full scope of NbCS impacts (Hemes et al., 2021). The rest of this section discusses approaches for NbCS project evaluation that meet the following criteria: (1) they leverage ecosystem-scale observations for robust yet financially feasible assessments, (2) they rely on transparent and reproducible protocols and algorithms, and objective validation, (3) they account for biophysical impacts and the future permanence of NbCS benefits, and (4) they aim to enhance equity and justice for demographic groups who have historically been, or stand to be, disproportionately impacted by NbCS projects (Fleischman et al., 2020).

3.3.1 | Leveraging ecosystem-scale data for monitoring and verification of NbCS projects

Flux towers are attractive tools for monitoring and verifying individual NbCS projects. They provide continuous data on the integrated carbon sources and sinks of an ecosystem, and their high level of precision ($\sim 50 \text{ tC km}^{-1}$, Hemes et al., 2021) can justify the crediting of a larger fraction of the projected carbon uptake compared to other quantification schemes. Moreover, the rapid response of ecosystem fluxes to land cover and management changes (e.g., Aguilos et al., 2020) can reveal the impacts of an NbCS intervention faster than inventorying slowly evolving biomass and soil carbon pools.

However, flux towers are expensive to install and operate, and it is not yet clear when they represent a cost-effective tool for project accounting. To address this question, we conducted a sensitivity analysis exploring how the benefit:cost ratio (BCR) of flux tower monitoring varies as a function of the project market value (representing the combined influence of sequestration potential and price of carbon) and the effective tower measurement height (which determines monitoring cost as well as the size of the measurement footprint, Chu et al., 2021). The analysis adopts the conservative constraint that flux towers should continuously monitor a project area of 1 km^2 (even if that requires the use of multiple towers via an economy of scale) for a project lifetime of 30 years. The project market value was initially set to $\sim \$10,000 \text{ km}^{-2} \text{ year}^{-1}$, based on an “additional” sequestration of $200 \text{ tCO}_2\text{e km}^{-2} \text{ year}^{-1}$ and a price of carbon at $\$50$ per tCO_2e , and reference annualized costs representing status quo monitoring approaches were set to $\$7000 \text{ km}^{-2} \text{ year}^{-1}$ (see S.I. for details). On this basis, the reference BCR was estimated to be $\sim 140\%$, that is, the revenue created by the project exceeds its cost by 40% .

While the BCR increased as a function of effective measurement height, even a tall tower ($\sim 100 \text{ m}$) did not reach cost neutrality

(BCR = 93% , Figure 6). BCR also increased as a function of the project market value, but not enough to motivate the use of standalone towers as a monitoring tool in most cases. The cost-effectiveness of flux towers as a monitoring tool may improve if they are deployed for shorter periods of time, or if the monitoring area is reduced to $<1 \text{ km}^2$. More work to understand the minimum requirements for tower time series length and footprint size, with full consideration of uncertainty related to neglected time periods and land surface heterogeneity, should be a research priority.

Substantial opportunity also exists to fuse flux tower data with complementary biometric observations (e.g. Harris et al., 2021; Smith et al., 2020) which can improve the BCR. Specifically, the development of comprehensive physical “Carbon Observing and Data Analysis Systems” (CODAS, see Figure 7) can enable scaling from measurement plots to ecosystems to landscapes with rigor, and also provide information on the processes that flux towers cannot see (e.g., lateral runoff, or non- CO_2 GHG emissions that are below instrument detection limits, Detto et al., 2011). In CODAS, data integration and scaling can be achieved with the use of the so-called “environmental response functions” (ERFs, Metzger, 2018; Metzger et al., 2013). The underlying principle of ERF is to use high-frequency (minute to minute) tower footprint variation to extract the relationships between tower-measured fluxes, meteorological forcing variables, and surface ecological and soil properties. Then, these extracted relationships

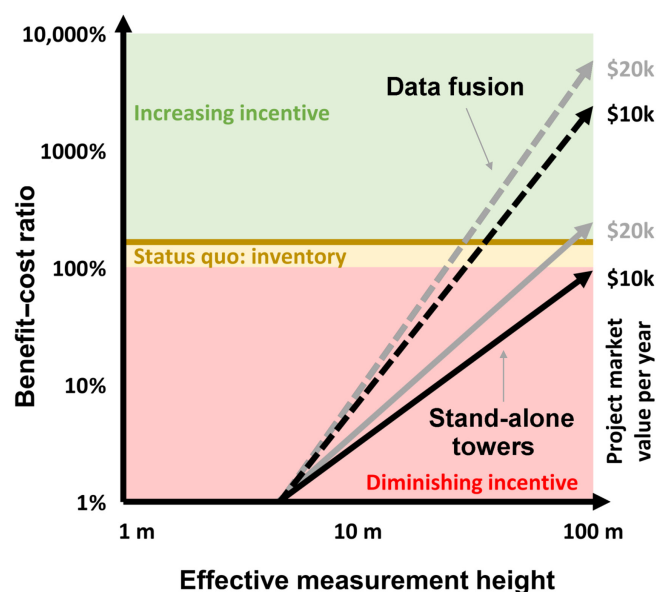


FIGURE 6 Benefit-cost ratio (BCR) of flux towers, and flux tower data fusion, for NbCS project monitoring. For the conservative constraint of continuously monitoring a 1 km^2 project area, and a 30-year project lifetime, cost neutrality is only approached when using very tall towers. However, substantial gains in BCR are achievable with data fusion for a virtual extension of the flux tower footprint (see details below). In each case, results are shown for two different estimates of the annual project market value ($\$10\text{K}$ and $\$20\text{K}$ per year). The thick yellow line shows the reference BCR of 140% [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/terms-and-conditions)]

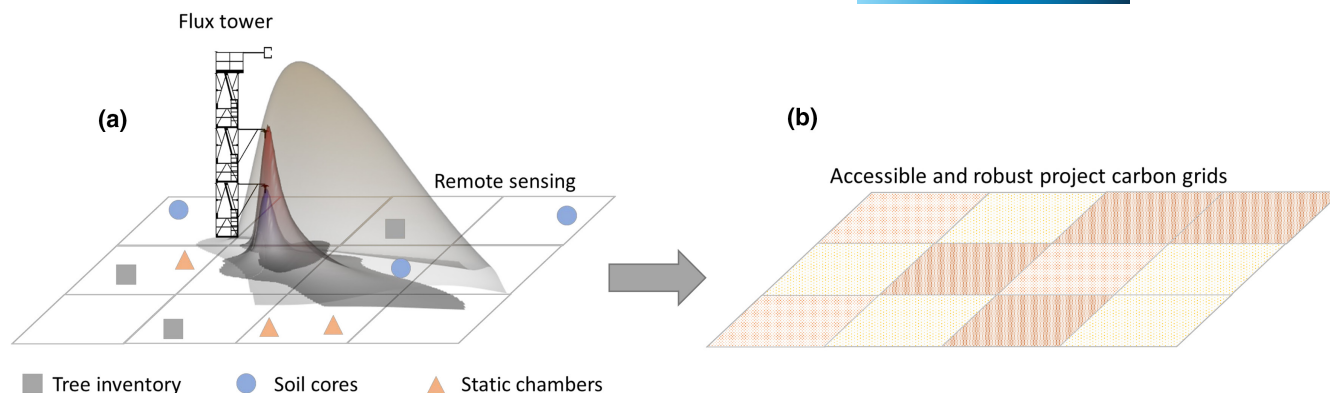


FIGURE 7 A Carbon Observing and Data Analysis System (CODAS) integrates multi-scale observations to accessible and robust project carbon grids. Panel (a): Flux tower CODAS reconciles the differing contexts in space and time among ground-based, airborne and spaceborne carbon observations: data fusion aims to harnesses the benefits and offset the limitations among the individual observation methods, thus fully utilizing their joint information to quantify NbCS project performance. Panel (b): The results are project carbon grids at half-hourly and decameter resolution with reduced cost per unit area and improved robustness compared to any individual observation method alone. This near-real time spatialization enables continuous assessment and optimization of localized management practices, and timely intervention for underperforming plots within the NbCS project area [Colour figure can be viewed at [wileyonlinelibrary.com](https://onlinelibrary.wiley.com/doi/10.1111/gcb.16156)]

can be combined with remote sensing data to create half-hourly, decameter-resolution carbon flux grids (e.g. Metzger et al., 2013; Xu et al., 2017; Figure 7), improving accuracy and precision for geographically representative and integrated impact assessments. Importantly, this virtual extension of the tower footprint substantially improves the BCR such that the status quo BCR can be exceeded for measurement heights on the order of 30 m or greater (Figure 6) even when carbon prices are relatively low (see S.I. for more details).

Novel quantification approaches like ERF data fusion require validation. Here, multi-scale benchmarking approaches originating from earth system modelling can provide a useful path forward. For example, the International Land Model Benchmarking (ILAMB) package (Collier et al., 2018) was originally designed to evaluate and benchmark land model results through comparison with site-, regional-, and global-scale observations, including from airborne CO₂ concentration measurements (Cui et al., 2021), satellite-based remote sensing (Eldering et al., 2017), machine-learning flux up-scaling products (Jung et al., 2019), and carbon cycle models and data assimilation products (e.g., NOAA CarbonTracker, Peters et al., 2007). When referenced to CODAS ground-truthing data, benchmarking systems like ILAMB could provide an important perspective on the magnitude and uncertainty of realized NbCS project benefits, as estimated from a range of ground-, airborne-, and spaceborne observations.

3.3.2 | Gold-standard datasets for transparent, reproducible, and objective validation of existing valuation schemes

Right now, most carbon market protocols use a combination of physical data and empirical models for carbon credit valuation, and then rely on independent, third-party verification to ensure methodological standards are met. These verifications tend to be costly,

lack standardization and transparency, and do not include truly independent validation based on alternative methods. Multi-scale, integrated, “gold-standard datasets” from representative flux tower sites could be used to validate and improve market verification schemes, especially if they include concurrent observations of carbon stocks, fluxes (from towers and including lateral exports in runoff), and near-surface and satellite remote sensing data. Since carbon accounting seeks to measure CO₂ removals or avoided emissions that are additional to the baseline, gold-standard datasets would be most useful if they (a) capture both pre- and post-intervention periods, (b) rely on paired-sites (pairing an NbCS treatment with a baseline control that experiences the same macroclimate), and/or (c) are developed in ways that leverage the ERF benefit of localizing fluxes across heterogeneous landscapes.

Substantial physical and cyberinfrastructure is already in place to support the creation of gold-standard datasets. At the time of this writing, data from more than 550 flux towers in the Americas have registered with AmeriFlux, with >400 having shared data. These towers include many paired sites (Figure 4) and multiple tall towers that are well suited for ERF scaling. Additional site pairs or tall towers could be created with strategic investment in new physical infrastructure; for example, adding a cropland monitoring site near an existing forest tower, or by physical amendments to extend the measurement height of the tower.

Many tower sites are already collecting some combination of biometric data, including measurements of soil C pool size, tree biomass, and/or chamber-based emissions (e.g. Campioli et al., 2016; Hollinger et al., 2021; Wang et al., 2017) which themselves may be shared to AmeriFlux to other relevant networks (e.g., the International Soil Carbon Network, Malhotra et al. (2019), or the COSORE soil respiration network, Bond-Lamberty et al., 2020). However, few sites are recording and sharing the full set of observations that would be most useful for robust assessments of NbCS. Moreover, with some exceptions (e.g., NEON sites, Metzger et al.,

2019), biometric data are not collected at flux tower sites using standardized protocols. Thus, enhancing at least a subset of existing flux towers with a fuller set of standardized biometric measurements could substantially advance a number of research goals.

The “gold-standard” datasets described here would provide a critical resource for systematic evaluation of existing accounting schemes currently in use in private carbon markets, which vary substantially from one market or entity to the next (see S.I. for details). One way to do this is through model-intercomparison projects (MIPs), which compare predictions from a variety of models driven by the same forcing data. The flux research community has substantial experience performing MIPs to benchmark and cross-compare TBMs (e.g. Friedlingstein et al., 2020; Huntzinger et al., 2013). To our knowledge, no such activity has been attempted for the diverse array of models used to project and quantify NbCS project benefits. A “Carbon Market MIP,” referenced to gold-standard data, could provide an unprecedented view of when and why the carbon market forecasting schemes differ. It would also enable the exploration of which physiological and ecological processes matter most for the application-based questions at hand (e.g., C storage and permanence), and could directly test the effects of NbCS management actions on these long-term carbon market aims. These information-rich datasets would also permit a systematic “measurement intercomparison” project, to understand where and why empirical accounting approaches differ. Prior work comparing flux tower and biometric data has been limited to forests (Campioli et al., 2016; Wang et al., 2017), and not designed with the specific goal of evaluating quantification schemes actually used in carbon market systems. Finally, these open and accessible datasets could also be accessible to private entities (e.g., independent third-party verifiers) working to develop new approaches for market-ready accounting protocols.

3.3.3 | Biophysical impacts and permanence

NbCS projects that modify local water and energy cycles in ways that exacerbate the negative consequences of climate change are counterproductive. On the other hand, NbCS projects that confer adaptative benefits for local hydrology and temperature may be more “valuable” from a climate mitigation and adaptation perspective. However, strategies to incorporate biophysical impacts and other co-benefits in carbon market structures are not at all clear (Anderson et al., 2011), since biophysical impacts tend to be local or regional, whereas enhanced C uptake or reduced GHG emissions are global benefits. It is also counterproductive to offset CO₂ emissions with carbon stored in forests that are likely to be decimated by wildfires, drought, or insect outbreaks within a few decades. Viable paths for factoring permanence into carbon credit valuation are also murky: the simple empirical models used for project accounting do not have a mechanism for considering climate feedbacks, whereas highly mechanistic Earth System Models do not agree on how climate feedbacks will impact global land carbon uptake. Rigorous,

multi-method approaches to estimating permanence risks—even if uncertainty is high—are needed.

For these reasons, incorporating biophysical feedbacks and permanence into market valuation schemes would likely require radical transformation of accounting and verification protocols, data, and model structures. In the case of biophysical impacts on energy balance, it may be relatively straightforward to “put a price” on the local temperature impacts of an NbCS strategy, since changes in both carbon and energy balance fluxes can be expressed in units of “radiative forcing” (Williams et al., 2021) or CO₂-e (Windisch et al., 2021). Moreover, if robust projections of carbon storage permanence and associated uncertainty become possible at the project scale, market structures should be able to accommodate some discounting of credits, since protocols already accommodate contributions to “buffer” insurance pools.

However, these new market structures would certainly take time to implement. In the meantime, policy mechanisms could be developed that specifically favor the implementation of NbCS in places where biophysical impacts are likely to be favorable, and where the threat of impermanence is comparatively low. For example, in the mesic and highly productive Eastern US, the risks of wildfire, drought, and insect-driven tree mortality are relatively small (Anderegg et al., 2021), and enhancing plant cover in the Eastern part of the country tends to have a surface cooling effect (Kaye & Quemada, 2017; Zhang et al., 2020). Thus, NbCS projects in the Eastern US that enhance tree cover may be a “safer bet” when compared to projects in the drought- and fire-prone Western US or Alaska.

3.3.4 | Inclusivity of solutions

Developing nations, poorer communities, and black, indigenous, and other people of color (BIPOC) communities frequently bear the brunt of climate change impacts (Hardy et al., 2017; Hoffman et al., 2020), while more developed nations and privileged communities often disproportionately benefit from greater monetization of NbCS and associated research funding (Lamb et al., 2019). Yet, many indigenous regions across the globe manage large carbon stocks, especially in aboveground biomass (Walker et al., 2014), which makes these regions especially vulnerable to climate change (Ramos-Castillo et al., 2017). In addition, continuous discrimination and underrepresentation of historically minoritized groups is especially prevalent among geoscience research communities (Ali et al., 2021; Marin-Spiotta et al., 2020). These problems require structural changes within academia, starting with inclusive mentoring and fieldwork policies, cultural exchanges, more funding opportunities for BIPOC students and researchers, and changes in the focus of teaching (Ali et al., 2021).

NbCS activities, funded via emission offsets and structured in a way that does not delay meaningful decarbonization in industries whose co-pollutants inordinately impact historically disadvantaged communities, can help achieve environmental justice goals. Moreover, inclusive and equitable practices for

NbCS monitoring and implementation include: early and transparent engagement with stakeholders, incorporation of traditional knowledge and cultural values, explicit mechanisms for stakeholder self-determination, as well as continuous cross-cultural education and training of principal investigators (Ramos-Castillo et al., 2017; Reo et al., 2017; Thompson et al., 2020; Varghese & Crawford, 2021). For example, in ecosystem service markets, large emphasis has been placed on monetizing the material contributions of ecosystems to human well-being (Van Riper et al., 2017). However, for Indigenous communities, these outcomes often do not meet their objectives, highlighting the need to include social benefits and values to NbCS solutions (Olander et al., 2018). NbCS projects that ensure the rights to land ownership, as well as full transparency of accounting methods to establish accessible, scientifically sound, and sustainable market options, can help prevent exploitation.

Moreover, sustainable and equitable carbon markets require a holistic picture of the co-benefits and unintended consequences of NbCS (Seddon et al., 2020), including the biophysical impacts to local water and temperature regimes. Flux towers provide information on these impacts, and when installed for long-term deployment, towers may also offer communities with opportunities for early detection of natural disturbances, such as drought or elevated fire risk. Collectively, flux tower networks have the infrastructure and resources to contribute to building stronger communities through collaborations, outreach, and support for members from a diverse set of backgrounds; however, they remain strongly dominated by towers and personnel from the Global North. Broadening the geographic and demographic composition of the networks can help to ensure adequate global representation in datasets and the pursuit of equitable solutions.

4 | SUMMARY AND CONCLUSIONS

The scientific community certainly has not reached consensus on the realizable climate benefits of Nature-based Climate Solutions (Anderegg et al., 2020; Fleischman et al., 2020; Seddon et al., 2020). Nonetheless, current enthusiasm for NbCS across a broad set of public and private entities, will likely lead to the advancement of NbCS within local and federal climate mitigation policy moving forward. In this paper, we propose multiple strategies for a modular and structural shift in research foci that will allow us to confront the most pressing sources of NbCS uncertainty with the best available science, at both the project scale and at the regional scales where policy decisions are made. These include the following:

- Synthesize existing flux tower network data for (a) direct assessment of mitigation potential and associate biophysical impacts in paired flux tower sites, (b) creating regional NbCS mitigation and adaptation potential maps through machine-learning up-scaling and/or benchmarking of next generation remote-sensing products, and (c) answering basic questions about how much

flux tower data is necessary to improve the precision and cost-effectiveness of project-scale monitoring and verification.

- Strategic deployment of new flux towers in underrepresented biomes (e.g., intermediate age forests, ecosystems managed with understudied NbCS strategies) and to increase the number of paired sites in the network.
- The creation of “gold-standard” datasets for a representative set of sites, featuring concurrent observations of carbon stocks (e.g., soil and tree inventories), fluxes (from towers and including lateral exports in runoff), and near-surface and satellite remote sensing data. These datasets could (a) reveal biases between the biometric data typically used in NbCS assessments, and the relatively more robust information contained in flux tower and some remote sensing data streams, (b) function as a platform for a carbon market model intercomparison project, and (c) function as a test-bed for novel schemes to quantify and monitor NbCS impacts.
- Operationalizing flux tower data fusion approaches (ERF, CODAS) that (a) facilitate co-interpretable gold-standard datasets that reconcile the differing space- and time-scales among biometric, flux tower, and remote sensing observations, (b) virtually extend the flux tower footprint for robust NbCS project monitoring with favorable benefit:cost ratios, and (c) reliably nest in-situ information into the communication among remote sensing, models and tools across project- and policy-relevant scales.
- Building more demographically diverse and representative research communities that are better equipped to develop equitable solutions for NbCS implementation.

We also recognize that some sources of NbCS uncertainty are more complex, and belie the expectation that they can be confronted with “modular” or “architectural shifts” to research infrastructure (Figure 1). These include the extraordinarily complex challenge of predicting how climate feedbacks will affect future land carbon uptake, as well as the difficult question of how to value biophysical impacts in carbon market structures. These knowledge gaps may require radical changes in our data and analysis tools, and/or radical shifts in how private carbon markets are structured. In the meantime, we emphasize the need for at least first-order predictions about where NbCS biophysical impacts and permanence are likely to be most favorable.

ACKNOWLEDGMENTS

We thank Martin De Kauwe and one anonymous reviewer for their helpful feedback. We acknowledge and thank the AmeriFlux Management Project and the U.S. Department of Energy, Office of Science for providing a platform for this collaboration, and for providing the data that inform Figure 4. We thank George Burba, Ankur Desai, and David Durden for providing a range of cost scenarios. KN and BRKR recognize NSF CAREER grant funding under award numbers 1552747 and 1752083, respectively, and KN acknowledges support from the O'Neill School of Public and Environmental Affairs at Indiana University through the Fischer Faculty Fellowship and Paul H. O'Neill Chair. KSH was supported by the Stanford Woods

Institute for the Environment. This material is based in part upon work supported by the National Ecological Observatory Network—a program sponsored by the National Science Foundation and operated under cooperative agreement by Battelle.

CONFLICT OF INTEREST

The findings and conclusions in this publication are those of the authors and should not be construed to represent any official USDA or U.S. Government determination or policy. The authors declare no conflict of interest.

DATA AVAILABILITY STATEMENT

The data that support the findings of this study are openly available in from the AmeriFlux network (<https://ameriflux.lbl.gov/>).

ORCID

Kimberly A. Novick  <https://orcid.org/0000-0002-8431-0879>
 Stefan Metzger  <https://orcid.org/0000-0002-4201-852X>
 William R. L. Anderegg  <https://orcid.org/0000-0001-6551-3331>
 Mallory Barnes  <https://orcid.org/0000-0001-8528-6981>
 Daniela S. Cala  <https://orcid.org/0000-0002-1165-1947>
 Kaiyu Guan  <https://orcid.org/0000-0002-3499-6382>
 Kyle S. Hemes  <https://orcid.org/0000-0001-5090-1083>
 David Y. Hollinger  <https://orcid.org/0000-0002-4284-1575>
 Jitendra Kumar  <https://orcid.org/0000-0002-0159-0546>
 Marcy Litvak  <https://orcid.org/0000-0002-4255-2263>
 Danica Lombardozzi  <https://orcid.org/0000-0003-3557-7929>
 Caroline P. Normile  <https://orcid.org/0000-0003-4409-470X>
 Patty Oikawa  <https://orcid.org/0000-0001-7852-4435>
 Benjamin R. K. Runkle  <https://orcid.org/0000-0002-2583-1199>
 Margaret Torn  <https://orcid.org/0000-0002-8174-0099>
 Susanne Wiesner  <https://orcid.org/0000-0001-7232-0458>

REFERENCES

- Aguilos, M., Mitra, B., Noormets, A., Minick, K., Prajapati, P., Gavazzi, M., Sun, G. E., McNulty, S., Li, X., Domec, J.-C., Miao, G., & King, J. (2020). Long-term carbon flux and balance in managed and natural coastal forested wetlands of the Southeastern USA. *Agricultural and Forest Meteorology*, 288–289, 108022. <https://doi.org/10.1016/j.agrformet.2020.108022>
- Ali, H. N., Sheffield, S. L., Bauer, J. E., Caballero-Gill, R. P., Gasparini, N. M., Libarkin, J., Gonzales, K. K., Willenbring, J., Amir-Lin, E., Cisneros, J., Desai, D., Erwin, M., Gallant, E., Gomez, K. J., Keisling, B. A., Mahon, R., Marin-Spiotta, E., Welcome, L., & Schneider, B. (2021). An actionable anti-racism plan for geoscience organizations. *Nature Communications*, 12, 3794. <https://doi.org/10.1038/s41467-021-23936-w>
- Amiro, B. D., Barr, A. G., Barr, J. G., Black, T. A., Bracho, R., Brown, M., Chen, J., Clark, K. L., Davis, K. J., Desai, A. R., Dore, S., Engel, V., Fuentes, J. D., Goldstein, A. H., Goulden, M. L., Kolb, T. E., Lavigne, M. B., Law, B. E., Margolis, H. A., ... Xiao, J. (2010). Ecosystem carbon dioxide fluxes after disturbance in forests of North America. *Journal of Geophysical Research: Biogeosciences*, 115, G00K02. <https://doi.org/10.1029/2010JG001390>
- Anderegg, W. R. (2021). Gambling with the climate: How risky of a bet are natural climate solutions? *AGU Advances*, 2, e2021AV000490. <https://doi.org/10.1029/2021AV000490>
- Anderegg, W. R., Chegwidden, O. S., Badgley, G., Trugman, A. T., Cullenward, D., Abatzoglou, J. T., Hicke, J. A., Freeman, J., & Hamman, J. J. (2021). Climate risks to carbon sequestration in US forests. *bioRxiv*. <https://doi.org/10.1101/2021.05.11.443688>
- Anderegg, W. R. L., Trugman, A. T., Badgley, G., Anderson, C. M., Bartuska, A., Ciais, P., Cullenward, D., Field, C. B., Freeman, J., Goetz, S. J., Hicke, J. A., Huntzinger, D., Jackson, R. B., Nickerson, J., Pacala, S., & Randerson, J. T. (2020). Climate-driven risks to the climate mitigation potential of forests. *Science*, 368, eaaz7005. <https://doi.org/10.1126/science.aaz7005>
- Anderson, C. M., DeFries, R. S., Litterman, R., Matson, P. A., Nepstad, D. C., Pacala, S., Schlesinger, W. H., Shaw, M. R., Smith, P., Weber, C., & Field, C. B. (2019). Natural climate solutions are not enough. *Science*, 363(6430), 933–934. <https://doi.org/10.1126/science.aaw2741>
- Anderson, R. G., Canadell, J. G., Randerson, J. T., Jackson, R. B., Hungate, B. A., Baldocchi, D. D., Ban-Weiss, G. A., Bonan, G. B., Caldeira, K., Cao, L., Diffenbaugh, N. S., Gurney, K. R., Kueppers, L. M., Law, B. E., Luyssaert, S., & O'Halloran, T. L. (2011). Biophysical considerations in forestry for climate protection. *Frontiers in Ecology and the Environment*, 9(3), 174–182. <https://doi.org/10.1890/090179>
- Arias-Ortiz, A., Oikawa, P. Y., Carlin, J., Masqué, P., Shahan, J., Kanneg, S., Paytan, A., & Baldocchi, D. D. (2021). Tidal and nontidal marsh restoration: A trade-off between carbon sequestration, methane emissions, and soil accretion. *Journal of Geophysical Research: Biogeosciences*, 126, e2021JG006573. <https://doi.org/10.1029/2021JG006573>
- Arora, V. K., Katavouta, A., Williams, R. G., Jones, C. D., Brovkin, V., Friedlingstein, P., Schwinger, J., Bopp, L., Boucher, O., Cadule, P., Chamberlain, M. A., Christian, J. R., Delire, C., Fisher, R. A., Hajima, T., Ilyina, T., Joetzier, E., Kawamiya, M., Koven, C. D., ... Ziehn, T. (2020). Carbon-concentration and carbon-climate feedbacks in CMIP6 models and their comparison to CMIP5 models. *Biogeosciences*, 17(16), 4173–4222. <https://doi.org/10.5194/bg-17-4173-2020>
- Azzari, G., Grassini, P., Edreira, J. I. R., Conley, S., Mourtzinis, S., & Lobell, D. B. (2019). Satellite mapping of tillage practices in the North Central US region from 2005 to 2016. *Remote Sensing of Environment*, 221, 417–429. <https://doi.org/10.1016/j.rse.2018.11.010>
- Badgley, G., Freeman, J., Hamman, J. J., Haya, B., Trugman, A. T., Anderegg, W. R., & Cullenward, D. (2022). Systematic over-crediting in California's forest carbon offsets program. *Global Change Biology*, 28, 1433–1445. <https://doi.org/10.1111/gcb.15943>
- Baker, J. M., & Griffis, T. J. (2005). Examining strategies to improve the carbon balance of corn/soybean agriculture using eddy covariance and mass balance techniques. *Agricultural and Forest Meteorology*, 128(3–4), 163–177. <https://doi.org/10.1016/j.agrformet.2004.11.005>
- Baldocchi, D. (2008). 'Breathing' of the terrestrial biosphere: Lessons learned from a global network of carbon dioxide flux measurement systems. *Australian Journal of Botany*, 56(1), 1. <https://doi.org/10.1071/BT07151>
- Barnes, M. L., Farella, M. M., Scott, R. L., Moore, D. J., Ponce-Campos, G. E., Biederman, J. A., MacBean, N., Litvak, M. E., & Breshears, D. D. (2021). Improved dryland carbon flux predictions with explicit consideration of water-carbon coupling. *Communications Earth & Environment*, 2, 248. <https://doi.org/10.1038/s43247-021-00308-2>
- Barnes, M. L., Yoder, L., & Khodaei, M. (2021). Detecting winter cover crops and crop residues in the midwest US using machine learning classification of thermal and optical imagery. *Remote Sensing*, 13(10), 1998. <https://doi.org/10.3390/rs13101998>
- Bechtold, W. A., & Patterson, P. L. (2005). *The enhanced forest inventory and analysis program—national sampling design and estimation procedures* (Vol. 80). USDA Forest Service, Southern Research Station. <https://doi.org/10.2737/SRS-GTR-80>

- Bogard, M. J., Bergamaschi, B. A., Butman, D. E., Anderson, F., Knox, S. H., & Windham-Myers, L. (2020). Hydrologic export is a major component of coastal wetland carbon budgets. *Global Biogeochemical Cycles*, 34, e2019GB00643. <https://doi.org/10.1029/2019GB006430>
- Bonan, G. B., & Doney, S. C. (2018). Climate, ecosystems, and planetary futures: The challenge to predict life in Earth system models. *Science*, 359, 533. <https://doi.org/10.1126/science.aam8328>
- Bond-Lamberty, B., Christianson, D. S., Malhotra, A., Pennington, S. C., Sihi, D., AghaKouchak, A., Anjileli, H., Altaf Arain, M., Armesto, J. J., Ashraf, S., Ataka, M., Baldocchi, D., Andrew Black, T., Buchmann, N., Carbone, M. S., Chang, S.-C., Crill, P., Curtis, P. S., Davidson, E. A., ... Zou, J. (2020). COSORE: A community database for continuous soil respiration and other soil-atmosphere greenhouse gas flux data. *Global Change Biology*, 26(12), 7268–7283. <https://doi.org/10.1111/gcb.15353>
- Campioli, M., Malhi, Y., Vicca, S., Luyssaert, S., Papale, D., Peñuelas, J., Reichstein, M., Migliavacca, M., Arain, M. A., & Janssens, I. A. (2016). Evaluating the convergence between eddy-covariance and biometric methods for assessing carbon budgets of forests. *Nature Communications*, 7, 13717. <https://doi.org/10.1038/ncomms13717>
- Chu, H., Luo, X., Ouyang, Z., Chan, W. S., Dengel, S., Biraud, S. C., Torn, M. S., Metzger, S., Kumar, J., Arain, M. A., Arkebauer, T. J., Baldocchi, D., Bernacchi, C., Billesbach, D., Black, T. A., Blanken, P. D., Bohrer, G., Bracho, R., Brown, S., ... Zona, D. (2021). Representativeness of Eddy-Covariance flux footprints for areas surrounding AmeriFlux sites. *Agricultural and Forest Meteorology*, 301–302, 108350. <https://doi.org/10.1016/j.agrformet.2021.108350>
- Churkina, G., & Running, S. W. (1998). Contrasting climatic controls on the estimated productivity of global terrestrial biomes. *Ecosystems*, 1(2), 206–215. <https://doi.org/10.1007/s100219900016>
- CNRA. (2021). Natural and working lands climate-smart strategy: Nature-based climate solutions. California Natural Resources Agency. Draft for Public Comment. October 11, 2021. Retrieved from https://resources.ca.gov/-/media/CNRA-Website/Files/Initiatives/Expanding-Nature-Based-Solutions/FINAL_DesignDraft_NWL_100821_508-opt.pdf
- Craig, M. E., Mayes, M. A., Sulman, B. N., & Walker, A. P. (2021). Biological mechanisms may contribute to soil carbon saturation patterns. *Global Change Biology*, 27(12), 2633–2644. <https://doi.org/10.1111/gcb.15584>
- Coffield, S. R., Hemes, K. S., Koven, C. D., Goulder, M. L., & Randerson, J. T. (2021). Climate-driven limits to future carbon storage in California's wildland ecosystems. *AGU Advances*, 2, e2021AV000384. <https://doi.org/10.1029/2021AV000490>
- Collier, N., Hoffman, F. M., Lawrence, D. M., Keppel-Aleks, G., Koven, C. D., Riley, W. J., Mu, M., & Randerson, J. T. (2018). The International Land Model Benchmarking (ILAMB) system: Design, theory, and implementation. *Journal of Advances in Modeling Earth Systems*, 10(11), 2731–2754. <https://doi.org/10.1029/2018MS001354>
- Cook-Patton, S. C., Leavitt, S. M., Gibbs, D., Harris, N. L., Lister, K., Anderson-Teixeira, K. J., Briggs, R. D., Chazdon, R. L., Crowther, T. W., Ellis, P. W., Griscom, H. P., Herrmann, V., Holl, K. D., Houghton, R. A., Larrosa, C., Lomax, G., Lucas, R., Madsen, P., Malhi, Y., ... Griscom, B. W. (2020). Mapping carbon accumulation potential from global natural forest regrowth. *Nature*, 585(7826), 545–550. <https://doi.org/10.1038/s41586-020-2686-x>
- Cox, P. M., Betts, R. A., Jones, C. D., Spall, S. A., & Totterdell, I. J. (2000). Acceleration of global warming due to carbon-cycle feedbacks in a coupled climate model. *Nature*, 408(6809), 184–187. <https://doi.org/10.1038/35047138>
- Cui, Y. Y., Jacobson, A. R., Feng, S., Wesloch, D., Barkley, Z. R., Zhang, L., Gerken, T., Keller, K., Baker, D., & Davis, K. J. (2021). Evaluation of CarbonTracker's inverse estimates of North American net ecosystem exchange of CO₂ from different observing systems using ACT-America airborne observations. *Journal of Geophysical Research: Atmospheres*, 126(12), e2020JD034406. <https://doi.org/10.1029/2020jd034406>
- Curtis, P. S., & Gough, C. M. (2018). Forest aging, disturbance and the carbon cycle. *New Phytologist*, 219(4), 1188–1193. <https://doi.org/10.1111/nph.15227>
- Detto, M., Verfaillie, J., Anderson, F., Xu, L., & Baldocchi, D. (2011). Comparing laser-based open-and closed-path gas analyzers to measure methane fluxes using the eddy covariance method. *Agricultural and Forest Meteorology*, 151(10), 1312–1324. <https://doi.org/10.1016/j.agrformet.2011.05.014>
- Dietze, M. (2017). *Ecological forecasting*. Princeton University Press. <https://doi.org/10.1515/9781400885459>
- Duman, T., Huang, C. W., & Litvak, M. E. (2021). Recent land cover changes in the Southwestern US lead to an increase in surface temperature. *Agricultural and Forest Meteorology*, 297, 108246. <https://doi.org/10.1016/j.agrformet.2020.108246>
- Duveiller, G., Caporaso, L., Abad-Viñas, R., Perugini, L., Grassi, G., Arneth, A., & Cescatti, A. (2020). Local biophysical effects of land use and land cover change: Towards an assessment tool for policy makers. *Land Use Policy*, 91, 104382. <https://doi.org/10.1016/j.landusepol.2019.104382>
- Eldering, A., Wennberg, P. O., Crisp, D., Schimel, D. S., Gunson, M. R., Chatterjee, A., Liu, J., Schwandner, F. M., Sun, Y., O'Dell, C. W., Frankenberg, C., Taylor, T., Fisher, B., Osterman, G. B., Wunch, D., Hakkarainen, J., Tamminen, J., & Weir, B. (2017). The Orbiting Carbon Observatory-2 early science investigations of regional carbon dioxide fluxes. *Science*, 358, 188. <https://doi.org/10.1126/science.aam5745>
- Fargione, J. E., Bassett, S., Boucher, T., Bridgman, S. D., Conant, R. T., Cook-Patton, S. C., Ellis, P. W., Falcucci, A., Fourqurean, J. W., Gopalakrishna, T., Gu, H., Henderson, B., Hurteau, M. D., Kroeger, K. D., Kroeger, T., Lark, T. J., Leavitt, S. M., Lomax, G., McDonald, R. I., ... Griscom, B. W. (2018). Natural climate solutions for the United States. *Science Advances*, 4(11), eaat1869. <https://doi.org/10.1126/sciadv.aat1869>
- Fisher, J. B., Lee, B., Purdy, A. J., Halverson, G. H., Dohlen, M. B., Cawse-Nicholson, K., Wang, A., Anderson, R. G., Aragon, B., Arain, M. A., Baldocchi, D. D., Baker, J. M., Barral, H., Bernacchi, C. J., Bernhofer, C., Biraud, S. C., Bohrer, G., Brunsell, N., Cappelaere, B., ... Hook, S. (2020). ECOSTRESS: NASA's next generation mission to measure evapotranspiration from the International Space Station. *Water Resources Research*, 56, e2019WR026058. <https://doi.org/10.1029/2019WR026058>
- Fisher, R. A., & Koven, C. D. (2020). Perspectives on the future of land surface models and the challenges of representing complex terrestrial systems. *Journal of Advances in Modeling Earth Systems*, 12, e2018MS001453. <https://doi.org/10.1029/2018MS001453>
- Fisher, R. A., Koven, C. D., Anderegg, W. R. L., Christoffersen, B. O., Dietze, M. C., Farrior, C. E., Holm, J. A., Hurtt, G. C., Knox, R. G., Lawrence, P. J., Lichstein, J. W., Longo, M., Matheny, A. M., Medvigy, D., Muller-Landau, H. C., Powell, T. L., Serbin, S. P., Sato, H., Shuman, J. K., ... Moorcroft, P. R. (2018). Vegetation demographics in Earth System Models: A review of progress and priorities. *Global Change Biology*, 24(1), 35–54. <https://doi.org/10.1111/gcb.13910>
- Fleischman, F., Basant, S., Chhatre, A., Coleman, E. A., Fischer, H. W., Gupta, D., Güneralp, B., Kashwan, P., Khatri, D., Muscarella, R., Powers, J. S., Ramprasad, V., Rana, P., Solorzano, C. R., & Veldman, J. W. (2020). Pitfalls of tree planting show why we need people-centered natural climate solutions. *BioScience*, 70, 947–950. <https://doi.org/10.1093/biosci/biaa094>
- Friedlingstein, P., Meinshausen, M., Arora, V. K., Jones, C. D., Anav, A., Liddicoat, S. K., & Knutti, R. (2014). Uncertainties in CMIP5 climate projections due to carbon cycle feedbacks. *Journal of Climate*, 27(2), 511–526. <https://doi.org/10.1175/JCLI-D-12-00579.1>
- Friedlingstein, P., O'Sullivan, M., Jones, M. W., Andrew, R. M., Hauck, J., Olsen, A., Peters, G. P., Peters, W., Pongratz, J., Sitch, S., Le Quéré,

- C., Canadell, J. G., Ciais, P., Jackson, R. B., Alin, S., Aragão, L. E. O. C., Arneeth, A., Arora, V., Bates, N. R., ... Zaehele, S. (2020). Global carbon budget 2020. *Earth System Science Data*, 12(4), 3269–3340. <https://doi.org/10.5194/essd-12-3269-2020>
- Fung, I. Y., Doney, S. C., Lindsay, K., & John, J. (2005). Evolution of carbon sinks in a changing climate. *Proceedings of the National Academy of Sciences of the United States of America*, 102(32), 11201–11206. <https://doi.org/10.1073/pnas.0504949102>
- Ge, J., Guo, W., Pitman, A. J., De Kauwe, M. G., Chen, X., & Fu, C. (2019). The nonradiative effect dominates local surface temperature change caused by afforestation in China. *Journal of Climate*, 32(14), 4445–4471. <https://doi.org/10.1175/JCLI-D-18-0772.1>
- Gough, C. M., Bohrer, G., Hardiman, B. S., Nave, L. E., Vogel, C. S., Atkins, J. W., Bond-Lamberty, B., Fahey, R. T., Fotis, A. T., Grigri, M. S., Haber, L. T., Ju, Y., Kleinke, C. L., Mathes, K. C., Nadelhoffer, K. J., Stuart-Haëntjens, E., & Curtis, P. S. (2021). Disturbance-accelerated succession increases the production of a temperate forest. *Ecological Applications*, 31, e02417. <https://doi.org/10.1002/eap.2417>
- Graham, M. W., Thomas, R. Q., Lombardozzi, D. L., & O'Rourke, M. E. (2021). Modest capacity of no-till farming to offset emissions over 21st century. *Environmental Research Letters*, 16(5), 054055. <https://doi.org/10.1088/1748-9326/abe6c6>
- Griscom, B. W., Adams, J., Ellis, P. W., Houghton, R. A., Lomax, G., Miteva, D. A., Schlesinger, W. H., Shoch, D., Siikamäki, J. V., Smith, P., Woodbury, P., Zganjar, C., Blackman, A., Campari, J., Conant, R. T., Delgado, C., Elias, P., Gopalakrishna, T., Hamsik, M. R., ... Fargione, J. (2017). Natural climate solutions. *Proceedings of the National Academy of Sciences of the United States of America*, 114(44), 11645–11650. <https://doi.org/10.1073/pnas.1710465114>
- Guan, K., Wu, J., Kimball, J. S., Anderson, M. C., Froking, S., Li, B., Hain, C. R., & Lobell, D. B. (2017). The shared and unique values of optical, fluorescence, thermal and microwave satellite data for estimating large-scale crop yields. *Remote Sensing of Environment*, 199, 333–349. <https://doi.org/10.1016/j.rse.2017.06.043>
- Günther, A., Barthelmes, A., Huth, V., Joosten, H., Jurasinski, G., Koebisch, F., & Couwenberg, J. (2020). Prompt rewetting of drained peatlands reduces climate warming despite methane emissions. *Nature Communications*, 11, 1644. <https://doi.org/10.1038/s41467-020-15499-z>
- Günther, A., Böther, S., Couwenberg, J., Hüttel, S., & Jurasinski, G. (2018). Profitability of direct greenhouse gas measurements in carbon credit schemes of peatland rewetting. *Ecological Economics*, 146, 766–771. <https://doi.org/10.1016/j.ecolecon.2017.12.025>
- Hardy, R. D., Milligan, R. A., & Heynen, N. (2017). Racial coastal formation: The environmental injustice of colorblind adaptation planning for sea-level rise. *Geoforum*, 87, 62–72. <https://doi.org/10.1016/j.geoforum.2017.10.005>
- Harper, A. B., Powell, T., Cox, P. M., House, J., Huntingford, C., Lenton, T. M., Sitch, S., Burke, E., Chadburn, S. E., Collins, W. J., Comyn-Platt, E., Daioglou, V., Doelman, J. C., Hayman, G., Robertson, E., van Vuuren, D., Wiltshire, A., Webber, C. P., Bastos, A., ... Shu, S. (2018). Land-use emissions play a critical role in land-based mitigation for Paris climate targets. *Nature Communications*, 9(1). <https://doi.org/10.1038/s41467-018-05340-z>
- Harris, N. L., Gibbs, D. A., Baccini, A., Birdsey, R. A., de Bruin, S., Farina, M., Fatoyinbo, L., Hansen, M. C., Herold, M., Houghton, R. A., Potapov, P. V., Suarez, D. R., Roman-Cuesta, R. M., Saatchi, S. S., Slay, C. M., Turubanova, S. A., & Tyukavina, A. (2021). Global maps of twenty-first century forest carbon fluxes. *Nature Climate Change*, 11(3), 234–240. <https://doi.org/10.1038/s41558-020-00976-6>
- Heavens, N. G., Ward, D. S., & Natalie, M. M. (2013). Studying and Projecting Climate Change with Earth System Models. *Nature Education Knowledge*, 4(5), 4.
- Helbig, M., Gerken, T., Beamesderfer, E. R., Baldocchi, D. D., Banerjee, T., Biraud, S. C., Brown, W. O. J., Brunsell, N. A., Burakowski, E. A., Burns, S. P., Butterworth, B. J., Chan, S., Davis, K. J., Desai, A. R., Fuentes, J. D., Hollinger, D. Y., Kljun, N., Mauder, M., Novick, K. A., ... Richardson, A. D. (2021). Integrating continuous atmospheric boundary layer and tower-based flux measurements to advance understanding of land-atmosphere interactions. *Agricultural and Forest Meteorology*, 307, 108509. <https://doi.org/10.1016/j.agrformet.2021.108509>
- Hemes, K. S., Chamberlain, S. D., Eichelmann, E., Knox, S. H., & Baldocchi, D. D. (2018). A biogeochemical compromise: The high methane cost of sequestering carbon in restored wetlands. *Geophysical Research Letters*, 45(12), 6081–6091. <https://doi.org/10.1029/2018GL077747>
- Hemes, K. S., Runkle, B. R., Novick, K. A., Baldocchi, D. D., & Field, C. B. (2021). An ecosystem-scale flux measurement strategy to assess natural climate solutions. *Environmental Science & Technology*, 55(6), 3494–3504. <https://doi.org/10.1021/acs.est.0c06421>
- Henderson, R. M., & Clark, K. B. (1990). Architectural innovation: The reconfiguration of existing product technologies and the failure of established firms. *Administrative Science Quarterly*, 35(1), 9. <https://doi.org/10.2307/2393549>
- Hicks Pries, C. E., Castanha, C., Porras, R. C., & Torn, M. S. (2017). The whole-soil carbon flux in response to warming. *Science*, 355(6332), 1420–1423. <https://doi.org/10.1126/science.aal1319>
- Hinckley, E.-L., Anderson, S. P., Baron, J. S., Blanken, P. D., Bonan, G. B., Bowman, W. D., Elmendorf, S. C., Fierer, N., Fox, A. M., Goodman, K. J., Jones, K. D., Lombardozzi, D. L., Lunch, C. K., Neff, J. C., SanClements, M. D., Suding, K. N., & Wieder, W. R. (2016). Optimizing available network resources to address questions in environmental biogeochemistry. *BioScience*, 66(4), 317–326. <https://doi.org/10.1093/biosci/biw005>
- Hoffman, J. S., Shandas, V., & Pendleton, N. (2020). The effects of historical housing policies on resident exposure to intra-urban heat: A study of 108 US urban areas. *Climate*, 8(1), 12. <https://doi.org/10.3390/cli8010012>
- Hollinger, D. Y., Davidson, E. A., Fraver, S., Hughes, H., Lee, J. T., Richardson, A. D., Savage, K., Sihi, D., & Teets, A. (2021). Multi-decadal carbon cycle measurements indicate resistance to external drivers of change at the Howland forest AmeriFlux site. *Journal of Geophysical Research: Biogeosciences*, 126, e2021JG006276. <https://doi.org/10.1029/2021JG006276>
- Huntzinger, D. N., Schwalm, C., Michalak, A. M., Schaefer, K., King, A. W., Wei, Y., Jacobson, A., Liu, S., Cook, R. B., Post, W. M., Berthier, G., Hayes, D., Huang, M., Ito, A., Lei, H., Lu, C., Mao, J., Peng, C. H., Peng, S., ... Zhu, Q. (2013). The North American carbon program multi-scale synthesis and terrestrial model intercomparison project—part 1: Overview and experimental design. *Geoscientific Model Development*, 6(6), 2121–2133. <https://doi.org/10.5194/gmd-6-2121-2013>
- Jackson, R. B., Jobbágy, E. G., Avissar, R., Roy, S. B., Barrett, D. J., Cook, C. W., Farley, K. A., Le Maitre, D. C., McCarl, B. A., & Murray, B. C. (2005). Trading water for carbon with biological carbon sequestration. *Science*, 310(5756), 1944–1947. <https://doi.org/10.1126/science.1119282>
- Jung, M., Koira, S., Weber, U., Ichii, K., Gans, F., Camps-Valls, G., Papale, D., Schwalm, C., Tramontana, G., & Reichstein, M. (2019). The FLUXCOM ensemble of global land-atmosphere energy fluxes. *Scientific Data*, 6, 74. <https://doi.org/10.1038/s41597-019-0076-8>
- Kaye, J. P., & Quemada, M. (2017). Using cover crops to mitigate and adapt to climate change. A review. *Agronomy for Sustainable Development*, 37, 4. <https://doi.org/10.1007/s13593-016-0410-x>
- Konings, A. G., Saatchi, S. S., Frankenberg, C., Keller, M., Leshy, V., Anderegg, W. R. L., Humphrey, V., Matheny, A. M., Trugman, A., Sack, L., Agee, E., Barnes, M. L., Binks, O., Cawse-Nicholson, K., Christoffersen, B. O., Entekhabi, D., Gentile, P., Holtzman, N. M., Katul, G. G., ... Zuidema, P. A. (2021). Detecting forest response to droughts with global observations of vegetation water content. *Global Change Biology*, 27(23), 6005–6024. <https://doi.org/10.1111/gcb.15872>
- Kroeger, K. D., Crooks, S., Moseman-Valtierra, S., & Tang, J. (2017). Restoring tides to reduce methane emissions in impounded wetlands:

- A new and potent Blue Carbon climate change intervention. *Scientific Reports*, 7, 11914. <https://doi.org/10.1038/s41598-017-12138-4>
- Lamb, W. F., Creutzig, F., Callaghan, M. W., & Minx, J. C. (2019). Learning about urban climate solutions from case studies. *Nature Climate Change*, 9(4), 279–287. <https://doi.org/10.1038/s41558-019-0440-x>
- Law, B. E., Sun, O. J., Campbell, J., Van Tuyl, S., & Thornton, P. E. (2003). Changes in carbon storage and fluxes in a chronosequence of ponderosa pine. *Global Change Biology*, 9(4), 510–524. <https://doi.org/10.1046/j.1365-2486.2003.00624.x>
- Lee, X., Goulden, M. L., Hollinger, D. Y., Barr, A., Black, T. A., Bohrer, G., Bracho, R., Drake, B., Goldstein, A., Gu, L., Katul, G., Kolb, T., Law, B. E., Margolis, H., Meyers, T., Monson, R., Munger, W., Oren, R., Paw, U. K. T., ... Zhao, L. (2011). Observed increase in local cooling effect of deforestation at higher latitudes. *Nature*, 479(7373), 384–387. <https://doi.org/10.1038/nature10588>
- Liu, T., Yu, L., & Zhang, S. (2019). Land surface temperature response to irrigated paddy field expansion: A case study of semi-arid western Jilin Province, China. *Scientific Reports*, 9, 5278. <https://doi.org/10.1038/s41598-019-41745-6>
- Lombardozi, D. L., Lu, Y., Lawrence, P. J., Lawrence, D. M., Swenson, S., Oleson, K. W., Wieder, W. R., & Ainsworth, E. A. (2020). Simulating agriculture in the Community Land Model version 5. *Journal of Geophysical Research: Biogeosciences*, 125, e2019JG005529. <https://doi.org/10.1029/2019JG005529>
- Magney, T. S., Barnes, M. L., & Yang, X. (2020). On the covariation of chlorophyll fluorescence and photosynthesis across scales. *Geophysical Research Letters*, 47, e2020GL091098. <https://doi.org/10.1029/2020GL091098>
- Malhotra, A., Todd-Brown, K., Nave, L. E., Batjes, N. H., Holmquist, J. R., Hoyt, A. M., Iversen, C. M., Jackson, R. B., Lajtha, K., Lawrence, C., & Vinduškova, O. (2019). The landscape of soil carbon data: Emerging questions, synergies and databases. *Progress in Physical Geography: Earth and Environment*, 43, 707–719. <https://doi.org/10.1177/0309133319873309>
- Marín-Spiotta, E., Barnes, R. T., Berhe, A. A., Hastings, M. G., Mattheis, A., Schneider, B., & Williams, B. M. (2020). Hostile climates are barriers to diversifying the geosciences. *Advances in Geosciences*, 53, 117–127. <https://doi.org/10.5194/adgeo-53-117-2020>
- Matthes, J. H., Sturtevant, C., Verfaillie, J., Knox, S., & Baldocchi, D. (2014). Parsing the variability in CH₄ flux at a spatially heterogeneous wetland: Integrating multiple eddy covariance towers with high-resolution flux footprint analysis. *Journal of Geophysical Research: Biogeosciences*, 119, 1322–1339. <https://doi.org/10.1002/2014JG002642>
- Metzger, S. (2018). Surface-atmosphere exchange in a box: Making the control volume a suitable representation for in-situ observations. *Agricultural and Forest Meteorology*, 255, 68–80. <https://doi.org/10.1016/j.agrformet.2017.08.037>
- Metzger, S., Ayres, E., Durden, D., Florian, C., Lee, R., Lunch, C., Luo, H., Pingth-Durden, N., Roberti, J. A., SanClements, M., Sturtevant, C., Xu, K. E., & Zulueta, R. C. (2019). From NEON field sites to data portal: A community resource for surface-atmosphere research comes online. *Bulletin of the American Meteorological Society*, 100(11), 2305–2325. <https://doi.org/10.1175/BAMS-D-17-0307.1>
- Metzger, S., Junkermann, W., Mauder, M., Butterbach-Bahl, K., Trancón y Widemann, B., Neidl, F., Schäfer, K., Wieneke, S., Zheng, X. H., Schmid, H. P., & Foken, T. (2013). Spatially explicit regionalization of airborne flux measurements using environmental response functions. *Biogeosciences*, 10(4), 2193–2217. <https://doi.org/10.5194/bg-10-2193-2013>
- Mohammed, G. H., Colombo, R., Middleton, E. M., Rascher, U., van der Tol, C., Nedbal, L., Goulas, Y., Pérez-Priego, O., Damm, A., Meroni, M., & Joiner, J. (2019). Remote sensing of solar-induced chlorophyll fluorescence (SIF) in vegetation: 50 years of progress. *Remote Sensing of Environment*, 231, 111177. <https://doi.org/10.1016/j.rse.2019.04.030>
- Nakhavali, M., Lauerwald, R., Regnier, P., Guenet, B., Chadburn, S., & Friedlingstein, P. (2021). Leaching of dissolved organic carbon from mineral soils plays a significant role in the terrestrial carbon balance. *Global Change Biology*, 27(5), 1083–1096. <https://doi.org/10.1111/gcb.15460>
- Nolan, C. J., Field, C. B., & Mach, K. J. (2021). Constraints and enablers for increasing carbon storage in the terrestrial biosphere. *Nature Reviews Earth & Environment*, 2(6), 436–446. <https://doi.org/10.1038/s43017-021-00166-8>
- Novick, K. A., Biederman, J. A., Desai, A. R., Litvak, M. E., Moore, D. J., Scott, R. L., & Torn, M. S. (2018). The AmeriFlux network: A coalition of the willing. *Agricultural and Forest Meteorology*, 249, 444–456. <https://doi.org/10.1016/j.agrformet.2017.10.009>
- Novick, K. A., & Katul, G. G. (2020). The duality of reforestation impacts on surface and air temperature. *Journal of Geophysical Research: Biogeosciences*, 124. <https://doi.org/10.1029/2019JG005543>
- Novick, K. A., Oishi, A. C., Ward, E. J., Siqueira, M. B., Juang, J. Y., & Stoy, P. C. (2015). On the difference in the net ecosystem exchange of CO₂ between deciduous and evergreen forests in the southeastern United States. *Global Change Biology*, 21, 827–842. <https://doi.org/10.1111/gcb.12723>
- Olander, L. P., Johnston, R. J., Tallis, H., Kagan, J., Maguire, L. A., Polasky, S., Urban, D., Boyd, J., Wainger, L., & Palmer, M. (2018). Benefit relevant indicators: Ecosystem services measures that link ecological and social outcomes. *Ecological Indicators*, 85, 1262–1272. <https://doi.org/10.1016/j.ecolind.2017.12.001>
- Osaka, S., Bellamy, R., & Castree, N. (2021). Framing “nature-based” solutions to climate change. *Wiley Interdisciplinary Reviews: Climate Change*, 12, e729. <https://doi.org/10.1002/wcc.729>
- Peters, W., Jacobson, A. R., Sweeney, C., Andrews, A. E., Conway, T. J., Masarie, K., Miller, J. B., Bruhwiler, L. M. P., Pétron, G., Hirsch, A. I., Worthy, D. E. J., van der Werf, G. R., Randerson, J. T., Wennberg, P. O., Krol, M. C., & Tans, P. P. (2007). An atmospheric perspective on North American carbon dioxide exchange: CarbonTracker. *Proceedings of the National Academy of Sciences of the United States of America*, 104(48), 18925–18930. <https://doi.org/10.1073/pnas.0708986104>
- Poeplau, C., & Don, A. (2015). Carbon sequestration in agricultural soils via cultivation of cover crops—A meta-analysis. *Agriculture, Ecosystems & Environment*, 200, 33–41. <https://doi.org/10.1016/j.agee.2014.10.024>
- Pongratz, J., Dolman, H., Don, A., Erb, K. H., Fuchs, R., Herold, M., Jones, C., Kuemmerle, T., Luyssaert, S., Meyfroidt, P., & Naudts, K. (2018). Models meet data: Challenges and opportunities in implementing land management in Earth system models. *Global Change Biology*, 24(4), 1470–1487. <https://doi.org/10.1111/gcb.13988>
- Quegan, S., Le Toan, T., Chave, J., Dall, J., Exbrayat, J.-F., Minh, D. H. T., Lomas, M., D'Alessandro, M. M., Paillou, P., Pathanassiou, K., Rocca, F., Saatchi, S., Scipal, K., Shugart, H., Smallman, T. L., Soja, M. J., Tebaldini, S., Ulander, L., Villard, L., & Williams, M. (2019). The European Space Agency BIOMASS mission: Measuring forest above-ground biomass from space. *Remote Sensing of Environment*, 227, 44–60. <https://doi.org/10.1016/j.rse.2019.03.032>
- Ramos-Castillo, A., Castellanos, E. J., & McLean, K. G. (2017). Indigenous peoples, local communities and climate change mitigation. *Climatic Change*, 140(1), 1–4. <https://doi.org/10.1007/s10584-016-1873-0>
- Reo, N. J., Whyte, K. P., McGregor, D., Smith, M. A., & Jenkins, J. F. (2017). Factors that support Indigenous involvement in multi-actor environmental stewardship. *AlterNative: An International Journal of Indigenous Peoples*, 13, 58–68. <https://doi.org/10.1177/1177180117701028>
- Rogelj, J., Shindell, D., Jiang, K., Fifita, S., Forster, P., Ginzburg, V., Handa, C., Kheshgi, H., Kobayashi, S., Kriegler, E., Mundaca, L., Séférian, R., & Vilariño, M. V. (2018). Mitigation pathways compatible with 1.5°C in the context of sustainable development. In V. Masson-Delmotte, P. Zhai, H.-O. Pörtner, D. Roberts, J. Skea, P. R. Shukla, A. Pirani, W. Moufouma-Okia, C. Péan, R. Pidcock, S. Connors, J. B. R. Matthews, Y. Chen, X. Zhou, M. I. Gomis, E. Lonnoy, T. Maycock, M. Tignor, & T. Waterfield (Eds.), *Global warming of 1.5°C* (pp. 93–174). IPCC. Retrieved from <https://www.ipcc.ch/sr15/>

- Rodríguez-Veiga, P., Wheeler, J., Louis, V., Tansey, K., & Balzter, H. (2017). Quantifying forest biomass carbon stocks from space. *Current Forestry Reports*, 3(1), 1–18. <https://doi.org/10.1007/s40725-017-0052-5>
- Rosentreter, J. A., Al-Haj, A. N., Fulweiler, R. W., & Williamson, P. (2021). Methane and nitrous oxide emissions complicate coastal blue carbon assessments. *Global Biogeochemical Cycles*, 35, e2020GB006858. <https://doi.org/10.1029/2020GB006858>
- Runkle, B. R., Suvočarev, K., Reba, M. L., Reavis, C. W., Smith, S. F., Chiu, Y. L., & Fong, B. (2019). Methane emission reductions from the alternate wetting and drying of rice fields detected using the eddy covariance method. *Environmental Science & Technology*, 53, 671–681. <https://doi.org/10.1021/acs.est.8b05535>
- Schimel, D., Schneider, F. D., and JPL Carbon and Ecosystem Participants. (2019). Flux towers in the sky: Global ecology from space. *New Phytologist*, 224(2), 570–584. <https://doi.org/10.1111/nph.15934>
- Schwingshackl, C., Hirschi, M., & Seneviratne, S. I. (2017). Quantifying spatiotemporal variations of soil moisture control on surface energy balance and near-surface air temperature. *Journal of Climate*, 30(18), 7105–7124. <https://doi.org/10.1175/jcli-d-16-0727.1>
- Seddon, N., Chausson, A., Berry, P., Girardin, C. A., Smith, A., & Turner, B. (2020). Understanding the value and limits of nature-based solutions to climate change and other global challenges. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 375(1794), 20190120. <https://doi.org/10.1098/rstb.2019.0120>
- Seddon, N., Smith, A., Smith, P., Key, I., Chausson, A., Girardin, C., House, J., Srivastava, S., & Turner, B. (2021). Getting the message right on nature-based solutions to climate change. *Global Change Biology*, 27(8), 1518–1546. <https://doi.org/10.1111/gcb.15513>
- Smith, P., Soussana, J.-F., Angers, D., Schipper, L., Chenu, C., Rasse, D. P., Batjes, N. H., Egmond, F., McNeill, S., Kuhnert, M., Arias-Navarro, C., Olesen, J. E., Chirinda, N., Fornara, D., Wollenberg, E., Álvaro-Fuentes, J., Sanz-Cobena, A., & Klumpp, K. (2020). How to measure, report and verify soil carbon change to realize the potential of soil carbon sequestration for atmospheric greenhouse gas removal. *Global Change Biology*, 26(1), 219–241. <https://doi.org/10.1111/gcb.14815>
- Sun, Y., Frankenberg, C., Wood, J. D., Schimel, D. S., Jung, M., Guanter, L., Drewry, D. T., Verma, M., Porcar-Castell, A., Griffis, T. J., Gu, L., Magney, T. S., Köhler, P., Evans, B., & Yuen, K. (2017). OCO-2 advances photosynthesis observation from space via solar-induced chlorophyll fluorescence. *Science*, 358, eaam574. <https://doi.org/10.1126/science.aam5747>
- Teuling, A. J., Seneviratne, S. I., Stöckli, R., Reichstein, M., Moors, E., Ciais, P., Luyssaert, S., van den Hurk, B., Ammann, C., Bernhofer, C., Dellwik, E., Gianelle, D., Gielen, B., Grünwald, T., Klumpp, K., Montagnani, L., Moureaux, C., Sottocornola, M., & Wohlfahrt, G. (2010). Contrasting response of European forest and grassland energy exchange to heatwaves. *Nature Geoscience*, 3(10), 722–727. <https://doi.org/10.1038/ngeo950>
- Thompson, K. L., Lantz, T., & Ban, N. (2020). A review of Indigenous knowledge and participation in environmental monitoring. *Ecology and Society*, 25, 10. <https://doi.org/10.5751/ES-11503-250210>
- Torn, M. S., & Chapin III, F. S. (1993). Environmental and biotic controls over methane flux from arctic tundra. *Chemosphere*, 26(1–4), 357–368. [https://doi.org/10.1016/0045-6535\(93\)90431-4](https://doi.org/10.1016/0045-6535(93)90431-4)
- Turner, J., Desai, A. R., Thom, J., & Wickland, K. P. (2021). Lagged wetland CH₄ flux response in a historically wet year. *Journal of Geophysical Research: Biogeosciences*, 126, e2021JG006458. <https://doi.org/10.1029/2021JG006458>
- Ustin, S. L., & Middleton, E. M. (2021). Current and near-term advances in Earth observation for ecological applications. *Ecological Processes*, 10, 1. <https://doi.org/10.1186/s13717-020-00255-4>
- Valach, A. C., Kasak, K., Hemes, K. S., Anthony, T. L., Dronova, I., Taddeo, S., Silver, W. L., Szutu, D., Verfaillie, J., & Baldocchi, D. D. (2021). Productive wetlands restored for carbon sequestration quickly become net CO₂ sinks with site-level factors driving uptake variability. *PLoS One*, 16(3), e0248398. <https://doi.org/10.1371/journal.pone.0248398>
- Van Riper, C. J., Landon, A. C., Kidd, S., Bitterman, P., Fitzgerald, L. A., Granek, E. F., Ibarra, S., Iwaniec, D., Raymond, C. M., & Toledo, D. (2017). Incorporating sociocultural phenomena into ecosystem-service valuation: The importance of critical pluralism. *BioScience*, 67, 233–244. <https://doi.org/10.1093/biosci/biw170>
- Varghese, J., & Crawford, S. S. (2021). A cultural framework for Indigenous, Local, and Science knowledge systems in ecology and natural resource management. *Ecological Monographs*, 91, e01431. <https://doi.org/10.1002/ecm.1431>
- Vermaat, J. E., Palt, M., Piffady, J., Putnins, A., & Kail, J. (2021). The effect of riparian woodland cover on ecosystem service delivery by river floodplains: A scenario assessment. *Ecosphere*, 12, e03716. <https://doi.org/10.1002/ecs2.3716>
- Walker, W., Baccini, A., Schwartzman, S., Ríos, S., Oliveira-Miranda, M. A., Augusto, C., Ruiz, M. R., Arrasco, C. S., Ricardo, B., Smith, R., Meyer, C., Jintach, J. C., & Campos, E. V. (2014). Forest carbon in Amazonia: The unrecognized contribution of indigenous territories and protected natural areas. *Carbon Management*, 5(5–6), 479–485. <https://doi.org/10.1080/17583004.2014.990680>
- Wang, H., Jiang, F., Wang, J., Ju, W., & Chen, J. M. (2019). Terrestrial ecosystem carbon flux estimated using GOSAT and OCO-2 XCO₂ retrievals. *Atmospheric Chemistry and Physics*, 19, 12067–12082. <https://doi.org/10.5194/acp-19-12067-2019>
- Wang, X., Wang, C., & Bond-Lamberty, B. (2017). Quantifying and reducing the differences in forest CO₂-fluxes estimated by eddy covariance, biometric and chamber methods: A global synthesis. *Agricultural and Forest Meteorology*, 247, 93–103. <https://doi.org/10.1016/j.agrformet.2017.07.023>
- Williams, C. A., Gu, H., & Jiao, T. (2021). Climate impacts of US forest loss span net warming to net cooling. *Science Advances*, 7, eaax8859. <https://doi.org/10.1126/sciadv.aax8859>
- Windisch, M. G., Davin, E. L., & Seneviratne, S. I. (2021). Prioritizing forestation based on biogeochemical and local biogeophysical impacts. *Nature Climate Change*, 11(10), 867–871. <https://doi.org/10.1038/s41558-021-01161-z>
- Xu, K., Metzger, S., & Desai, A. R. (2017). Upscaling tower-observed turbulent exchange at fine spatio-temporal resolution using environmental response functions. *Agricultural and Forest Meteorology*, 232, 10–22. <https://doi.org/10.1016/j.agrformet.2016.07.019>
- Yin, Y. I., Byrne, B., Liu, J., Wennberg, P. O., Davis, K. J., Magney, T., Köhler, P., He, L., Jeyaram, R., Humphrey, V., Gerken, T., Feng, S., Digangi, J. P., & Frankenberg, C. (2020). Cropland carbon uptake delayed and reduced by 2019 Midwest floods. *AGU Advances*, 1, e2019AV00014. <https://doi.org/10.1029/2019AV000140>
- Zhang, Q., Barnes, M., Benson, M., Burakowski, E., Oishi, A. C., Quimette, A., Sanders-DeMott, R., Stoy, P. C., Wenzel, M., Xiong, L., Yi, K., & Novick, K. A. (2020). Reforestation and surface cooling in temperate zones: Mechanisms and implications. *Global Change Biology*, 26(6), 3384–3401. <https://doi.org/10.1111/gcb.15069>

SUPPORTING INFORMATION

Additional supporting information may be found in the online version of the article at the publisher's website.

How to cite this article: Novick, K. A., Metzger, S., Anderegg, W. R. L., Barnes, M., Cala, D. S., Guan, K., Hemes, K. S., Hollinger, D. Y., Kumar, J., Litvak, M., Lombardozzi, D., Normile, C. P., Oikawa, P., Runkle, B. R. K., Torn, M., & Wiesner, S. (2022). Informing Nature-based Climate Solutions for the United States with the best-available science. *Global Change Biology*, 28, 3778–3794. <https://doi.org/10.1111/gcb.16156>