

ANADROMOUS SALMONID RESOURCES
OF MENDOCINO COASTAL AND
INLAND RIVERS

1990-91 through 1991-92

An Evaluation of Rehabilitation Efforts
Based on Carcass Recovery,
and Spawning Activity

Draft
FINAL REPORT

by
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ABSTRACT

Spawning surveys were conducted for three years in Mendocino County streams. The results from the last two years of surveys are reported here. Surveys were conducted in seven watersheds in 1991 1992 and in ten watersheds in 1990-1991. These surveys were conducted to as a means to evaluate the effectiveness of restoration efforts to restore salmon populations. The populations of chinook and coho salmon were estimated by several methods including marking and recapture of carcasses, area-under-the-curve (AUC) method and from numbers of redds. A new model was developed to estimate spawning numbers which incorporated estimates of the average daily retention rate of carcasses. The reliability, of estimates are assessed in two streams where known numbers of fish were released. In both areas carcass retention and AUC estimates produced numbers_ far less than the of fish released. Redd counts were used to establish a range in spawning estimates which are believed to encompass the actual populations. It was determined that coho salmon could be expected to produce between 1 and 4 redds per female and chinook salmon from 1 to 1.8 redds per female.

The largest population of coho salmon was found in the South Fork Noyo where the run is enhanced through artificial production. This run ranged between 681 and 1197 in 1991-1992 and between 208 and 400 in 1990-91. The remainder of coho runs surveyed were supported by natural production only. The Hollowtree Creek coho run was about 120 fish in 1991-92 and between 50 and 90 in 1990-91. Caspar Creek had between 50 and 200 in 1991-1992 and between 8 and 28 in 1990-1991. Little River had less than 30 in 1991-1992 and only a single pair in 1990-1991. The Pudding Creek coho run was_ between 20 and 100 fish both years. The chinook run in Hollowtree Creek was between 270 and 420 in 1991-92 and between 18 and 53 in 1990-91. The 1991-92 chinook run was composed of 5, 5, 51, 35 and 3 percent age 2, 3, 4, 5 and 6 year old fish, respectively. Returns from hatchery released chinook composed 10 percent of the age four fish. In some areas, low water conditions are believed to have impacted the number of fish that spawned.

Of five streams into which coho were reintroduced, only one had spawning activity believed to be returning coho adults from plantings. The number in this stream was quite low- only two to eight spawning coho. It is believed that low flows conditions may have prevented coho from entering planted streams. Introductions of chinook salmon into the Ten Mile River have resulted in natural chinook production. Habitat restoration activities could not be linked to improved coho or chinook salmon production in Mendocino County streams.

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INTRODUCTION

Rivers of cool, clear water once tumbled down river valleys, propelled through deep dark pools embraced by an abundance of moss covered logs and surrounded by lush stands of giant conifers and water loving alders. These were the conditions that great abundances of salmon once met on their return from the sea. These were the conditions in which they evolved; the conditions they encountered as they emerged from the gravel of their birth and to which they returned to spawn, digging their redds perhaps, down into the exact spot from which their life began.

As conditions in the streams have deteriorated so to have the numbers of salmon declined. With river systems and salmon runs now only remnants of what they once were those who wish to see rivers alive with spawning fish try, as best they can, to restore these runs of salmon; to try and recreate again this once mighty resource. In this endeavour we have, through the expenditure of many millions of dollars, funded stream enhancement work, reared and released into the wild millions of artificially produced young fish, passed laws to limit degradation to natural environment and restrict the take of salmon. While we continue these efforts, we do it with strong desire and commitment and with the presumption that what we are doing is beneficial; that the natural production of salmon in our streams will increase from our efforts.

This three year study was initiated as part of a desire to investigate whether or not our restoration efforts have been a success. The streams of Mendocino County have undergone extensive restoration activities during the last 30 years. There have been many artificial propagation programs and fisheries have been heavily curtailed, while at the same time, little to nothing is known about the population levels of salmon and steelhead in Mendocino County. In this study, it was hoped, that through a spawning survey one could assess the extent to which salmon were

using streams for spawning. This would-tell us whether salmon runs were present in these "restored streams" and give us an indication as to the magnitude of these runs. From this information we could make a reasoned judgement as to whether our restoration activities have been of value in restoring our runs of salmon and steelhead. Streams initially selected for surveys were those that significant restoration work and/or artificial propagation had occurred. In addition, California Department of Fish and Game requested that several streams be included in surveys as part of a program to effort to establish index areas for monitoring purposes.

In the first year (the 1989-90 spawning season), over 100 tributaries and rivers sections were surveyed. An initial report (Nielsen, J.L. et. al 1990) reported the finding and included an evaluation of various spawning population estimation models. This report presents the results from the last two years of surveys and discusses the overall results from the three year survey in relation to the restoration activities that have occurred in these streams. The streams and their locations are shown on Figure 1.

METHODS

Survey methods used are described in detail in Nielsen, et al, (1990). In brief, spawning ground surveys were conducted where salmon carcasses were tagged with colored hog-rings in the lower jaw. The color was alternated weekly to establish mark-recapture data. Counts of redds and live fish were made during each survey. One change from the initial survey year is that all redds were flagged to prevent recounting on following surveys. In addition, flow and turbidity data were collected. Most streams were surveyed on a weekly basis unless conditions were unsuitable. All carcasses were tagged with hog-rings if there were jaws present. Skeletons without jaws were tagged with a hole punch if the tail was present and notes were taken where skeletons of fish were found which could no be tagged. Only jaw tagged carcasses were used to determine

recovery rates. All carcasses that. were measurable were measured (fork length) and a scale sample was taken for age determination. Fish carcasses were also examined for adipose fin clips for Coded-Wire-Tag (CWT) recovery. All stream surveyors were experienced having participated in the initial survey in 1989-90.

None of the streams sampled had sufficient number of tags or tag recoveries to preform the Jolly-Seber estimates or the nonparametric area-under-the-curve model used to estimate populations in some of the streams in the 1989-90 survey (Nielsen, et. al., 1990).

The standard area-under-the-curve (AUC) estimates (Beidler and Nickelson, 1980) of spawning population based on carcass and live fish counts are made in this report. This model estimates the population of spawning salmon by dividing the integral of the escapement curve by the average residence time of fish in the survey area. Since counts of live fish include fish that are not always identified to species, the total estimate from live fish counts is apportioned based on the fraction that each species composed within the group identified to species. No estimate of the life span of a spawning fish was estimated in this study so the 11 day average (Beidler & Nickelson, 1980) was used. The AUC procedure for carcasses data is the same only the mean carcass duration was estimated based on tag and recapture data from each streams.

A new model was developed based on the observed carcass retention rates found during this study. This model, called here the Carcass Retention Model, has one basic assumption:

That carcasses are removed or disappear from the survey area at a constant daily rate.

With this assumption and with the ability to estimate this rate

based on the recovery of tagged carcasses an estimate of the number of spawners can be made.

Carcass Retention Model.

This model assumes that there is a constant daily rate of carcass retention meaning that for each day after carcass deposition a fraction of those carcasses disappear or become untaggable and that fraction remains constant throughout the spawning period. The model assigns the number of carcasses found to each day in the interval between surveys by weighing each day by the likelihood that a carcass deposited would still be there when the survey was conducted. For example, if a survey was conducted on a seven-day interval, a carcasses deposited on the first day will have a much less chance of being found than a carcass deposited the day prior to the survey. The model calculates the number of spawners in the survey interval as:

$$\text{Carcass pop} = \frac{[D1/\text{sum}(Di-Dj)] * C}{D1} * \text{Days}$$

Where: D1 = Daily carcass retention

sum(Di-Dj)= sum of carcass retention for each day between interval

C = Number of carcasses found

Days = Days in interval between surveys

Where the daily carcass retention is .50, the interval 3 days and the number of carcasses found 90, the population of carcasses would have been:

$$\frac{[.50/ (.50 + .25 + .125)]*90}{.50} * 3 = 308$$

Or, where the daily carcass retention was .90 and number of days 4:

$$\frac{[.90/ (.90 + .81 + .73 + .66)]*90}{.90} * 4 = 116$$

In each stream or stream section, an estimate of "carcass retention" is made unless too little tagging data was collected to make this estimate. The retention curve that best represents the recovery data points is the retention curve that is selected. It was necessary to make average recovery rate estimates by lumping and weighing data. For example, where the spawning survey interval was 7 days and there were 20 fish tagged and 4 recovered and a spawning survey interval of 8 days where 2 fish were tagged and none recovered, the data was weighted as shown below:

$$\frac{7*20 + 8*2}{22} = 7.09 \text{ days and } \frac{4 \text{ recoveries}}{22 \text{ tagged}} = 18\%$$

Here, at a period of 7.09 days there was an 18 percent recovery rate. Where too little tagging data was available to estimate carcass retention, a retention estimate from a stream with a similar overall recovery rate, ie., number of recoveries divided by number tagged from the entire survey, was used.

This same procedure was used to estimate the mean carcass duration for the AUC population estimation. With an estimate of the daily retention rate., the number of days it would take to have a 50

Redd based Spawning Population Estimates

It has been observed that only female salmon do the actual redd digging (Briggs 1953; Shapovalov and Taft 1954) and since in this survey known numbers of females were going to be released at specific locations, the number of redds above that site could be used to establish the average number of redds produced per female. This ratio could then be used to estimate the number of females spawning in other areas. Since the stream surveyors rotated between survey areas no observer bias should effect the estimates of the number of redds encountered in different survey areas. The estimated number of females from redd counts are then expanded assuming that 50 percent of the runs are composed of female spawners. The exception to this is for the South Fork Noyo and Hollowtree Creek where a large sample of carcasses could be used determine male to female ratios. Over 70 carcasses were identified to sex in these areas where the 50 percent figure was not used. In other areas, carcasses numbers were below twenty. The redd based population estimates will be derived from a range in the estimated number of redds dug by female coho from 1 to 4 based on data collected in the South Fork Noyo. For chinook, 1 to 1.8 redds/female is used based on data collected in Hollowtree Creek.

Most often, only the redds found through the month of January are used to estimate salmon spawning. In February, most of the redds are made by steelhead but occasionally, where supported by carcass or live fish observations, all or partial February redd counts are also used.

Where both coho and chinook were found in relatively similar numbers, the higher number of redds/female, ie., 4, was not used and a lower figure is used instead. This is done because data collected in this study indicated that chinook were not digging,

on average, as many redds as coho, and that by using four redds/female to set the lower range of the population size it would be set too low. In the areas where both chinook and coho occurred together, the ratio of these species found in carcass samples is used to proportion the total salmon population estimates obtained from redd counts.

RESULTS AND DISCUSSION

In 1991-92, seven watersheds were surveyed which encompassed 114 miles of spawning area or a total of 443 total survey miles. In Table 1., the number of miles surveyed, the stream reach length, number of surveys conducted, number of live fish per survey mile, number of redds and number of carcasses tagged are given for each stream surveyed in 1991-92. In 1990-91, 240 miles of stream survey in ten watersheds (Figure 1.) encompassing over 90 miles of spawning area. The results for this years survey are shown in Table 2. Data for live fish and redds are divided between surveys conducted before and after February 1st. Before February 1st, few steelhead are seen. The December through January data primarily reflect salmon spawning while February is usually dominated by steelhead.

In 1991-92, there were 159 chinook, 206 coho, 8 steelhead and 5 unidentified carcasses tagged. Total redd counts were 1,159. In 1990-91, 6 chinook, 64 coho, 6 steelhead and 13 unidentified carcasses were tagged and 414 redds counted.

In both survey years, the greatest spawning density as evidenced by the number of live fish and redds was in the lower South Fork Noyo. This higher density is due to the return of hatchery reared coho. In January 1991-92 there were 16 live fish per survey mile

Figure 1. Location of Streams Surveyed

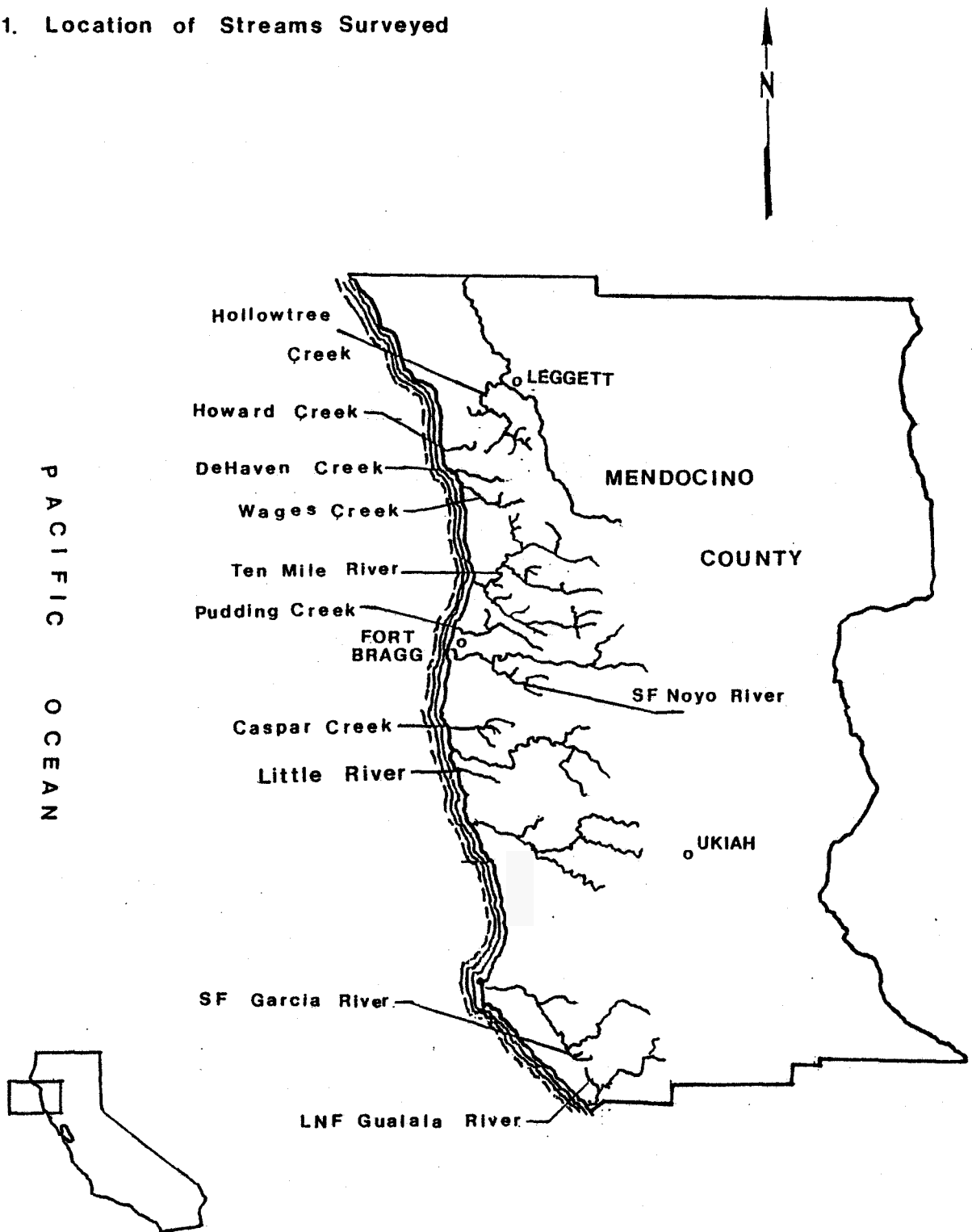


Table 1. Total number of miles surveyed, length of survey reach, number of surveys, number of live fish per survey mile, number of redds and number of carcasses jaw tagged during the 1991-1992 survey

Stream	Survey Miles	Reach Len.(mi)	# Surveys		Live/mi		Redds**		# Tagged carcasses by specie ***			
			Dec-Jan	Feb-Mar	Dec-Jan	Feb-Mar	Dec-Jan	Feb-Mar	Chinook	Coho	Steelhead	Sp.?
Caspar Creek												
Mainstem	25	3	7	2	0.8	0.0	38(12.7)	2(0.7)	0	11	0	0
North Fork	13.3	1.9	5	2	1.4	0.0	54(28.4)	5(2.7)	1	9	2	1
South Fork	2.5	.5	4	1	0.0	0.0	6(12.0)	1(2.0)	0	0	0	0
Middle Fork	.5	.5	1	0	0.0	NS	0.0	NS	0	0	0	0
Hollowtree Creek												
Below Weir	62.4	7.8	6	2	11.8	0.1	151(19)	3(0.4)	126(137)	9	0	2
Above Weir*	30.5	11.3	4	3	3.6	0.8	20(1.8)	43(3.8)	11(14)	13	1	0
Huckleberry Cr	1	1	0	1	NS	0.0	NS	0.0	0	0	0	0
Redwood Cr.	.6	.3	2	0	0.0	NS	8(26.6)	NS	0	0	0	0
Michaels Cr.	.4	.4	1	0	0.0	NS	5(12.5)	NS	0	0	0	0
Bond Cr	2	2	0	1	NS	0.5	NS	13(6.5)	0	0	1	0
Little River	32	4	5	3	0.6	0.5	17(4.3)	9(2.3)	0	8(11)	0	0
S.F. Noya River												
North Fork	31.1	3.4	5	4	1.4	0.2	40(11.8)	9(2.6)	0	5(6)	0	0
Up. S.F.	38.9	4.8	5	4	5.7	0.6	76(22.4)	26(5.4)	0	11(17)	0	0
Parlin Cr.	2.2	1	3	0	3.6	NS	2(2.0)	NS	0	0	0	0
Low. S.F.	39	3.9	7	3	16.0	0.6	275(70)	21(5.4)	0	93(99)	0	1
Kass Cr.	11.2	1.6	4	3	5.3	0.0	51(31.9)	12(7.5)	0	13(18)	1	0
Bear Cr.	.5	.5	1	0	0.0	NS	6(12.0)	NS	0	0	0	0
Pudding Creek												
Lower	29.4	4.2	4	3	1.1	1.8	31(7.4)	39(9.3)	0	9(11)	0	0
Upper	20.5	4.1	2	3	0.2	0.3	9(2.2)	18(4.4)	0	0	0	0
Ten Mile River												
North Fork*	25.6	13	3	0	0.3	NS	25(1.9)	NS	0	0	1(2)	0
Bald Hills Cr.	1.9	1.9	1	0	0.0	NS	1(0.5)	NS	0	0	0	0
Little N. Fork	11.3	3.5	2	2	0.6	0.3	7(2.0)	23(6.6)	1	0	0	0
Middle (Clark)*	29.5	14.8	2	1	0.0	0.2	32(2.2)	36(2.4)	6(8)	1(2)	0	0
Bear Haven Cr.	10	2.5	2	2	0.2	0.0	17(6.8)	12(4.8)	0	0	1	0(1)
South Fork *	13.8	14.8	2	0	0.5	NS	14(0.9)	NS	2	1	0	0
Campbell Cr. *	2.4	1.6	2	0	0.0	NS	2(1.3)	NS	0	0	0	0
Smith Cr.	2.8	2.8	1	0	0.0	NS	0.0	NS	0	0	0	0
Willits Cr	2.5	2.5	1	0	0	NS	0	NS	0	0	0	0

NS - Not surveyed

* Incomplete surveys of reach where fish use is believed likely

** Number of redds per reach mile in ()'s

*** Number in ()'s includes caudal punched carcasses in addition to jaw tagged carcasses

Table 2. Total number of miles surveyed, length of survey reach, number of surveys, number of live fish per survey mile, number of redds and number of carcasses jaw tagged during the 1990-1991 survey

Stream	Survey Miles	Reach Len. (mi)	# Surveys		Live/mi Jan.	Live/mi Feb.	Redds**		Redds**		Tagged carcasses by specie ***		
			Jan	Feb.			Jan.	Feb.	Jan.	Feb.	Chinook	Coho	Steelhead Sp.?
DeHaven Creek	9	3	2	1	0	0	0		0	0	0	0	0
Caspar Creek													
Mainstem	21	3	3	4	.05	1.8	14(4.7)	35(11.7)	0	0	1	1(2)	
North Fork	9.5	1.9	1	3	0	1.1	0	20(10.5)	0	0	0	0	0
South Fork	2.5	.5	1	4	0	0	0	0	0	0	0	0	0
Garcia River SF	6	2	0	3	HS	0	NS	2(1.0)	0	0	0	0	0
Gualala River LNF	8.7	2.9	0	3	NS	0	NS	2(0.7)	0	0	0	0	0
Hollovtree Creek													
Below Weir	46.8	7.8	2	4	.3	.6	19(2.4)	48(6.2)	2(4)	3	(0)	3(7)	
Above Weir *	48.7	11.3	3	6	.3	.2	23(2.2)	35(3.1)	1(2)	2	0	0	0
Huckelberry Cr	1	.5	2	0	0	NS	2(4.0)	HS	0	0	0	0	0
Redwood Cr.	1.5	.5	1	2	0	0	1(2.0)	4(8.0)	0	0	0)	1(2)	
Howard Creek	7.5	2.5	2	1	0	0	0	0	0	0	0	0	0
Little River	10.5	3.5	2	1	0	0	0	1(0.3)	0	0	0	0	0
South Fork Noyo													
Lower	26.1	3.9	3	4	12.7	1.1	97(24.9)	30(7.1)	0	33(51)	0	0	0
Kass Creek	.5	.5	1	0	0	NS	0	NS	0	0	0	0	0
Pudding Creek													
Lower	26.2	4.2	2	4	0	1.6	2(0.5)	54(12.9)	0	5	3	1(2)	
Upper *	7.6	2.9	0	3	NS-0	.9	NS-0	11 (3.8)	0	2	0	0	0
Wages Creek	16	4	3	2	0	.3	4(1.0)	8(2.0)	0	1	1(2)	0	0

NS- Not surveyed

NS-0 Not surveyed but assumed to be no fish use because of low flow condition

* Incomplete surveys of reach where fish use is believed likely

** Number of redds per reach mile in ()'s

*** Number in ()'s includes caudal punched carcasses in addition to jaw tagged carcasses

and 70 redds per reach mile in the lower South Fork. In 1990-91. 12.7 live fish/mile and 25 redds/mile (R/M) were found in this same reach. Similarly, 99 coho carcasses were tagged in 1991-92 while 51 were tagged in 1990-91.

Caspar Creek, in either it's North Fork or mainstem, had the second highest density of January redd counts both years. Other areas which had relatively high January redd densities were Redwood Creek, a tributary to Hollowtree Creek and Bearhaven Creek, a tributary to the Ten Mile River. The highest density of redds in the month of February was found in Pudding Creek in both years. Where normally the majority of February redd counts indicate steelhead spawning, in Pudding Creek, it was due primarily to coho spawning.

In every stream surveyed, the returns in 1991-92 were better than in 1990-91 and in many cases the difference was dramatic. In lower Hollowtree Creek, live fish counts were 11.8/mile in January 1991-92 but only 0.3 in 1990-91; the density of redds were 19/mile compared to 2.4/mile and the numbers of chinook carcasses were 137 compared to 4 in 1990-91 years. Similarly, Little River which had 0.6 live/mile, 4.3 redds/mile and 11 tagged coho carcasses in 1991-92 had no live fish, carcasses and only a single redd observed in 1990-91.

There were two counts of fish released above a counting station for which complete enough surveys were conducted to make population estimate comparisons. These were a release above the Noyo egg-taking station in 1991-92 and above the Hollowtree Creek Hatchery in 1990-91. In Table 3. the numbers of fish released are compared to the estimates based on the different population models. Clearly, the CR and AUC Carcass models under-estimate the known populations dramatically. The live fish based estimates have also under-estimated the numbers released although not as significantly. Thorough analysis of these estimates in relation to the numbers of

fish released are found in the South Fork Noyo and Hollowtree Creek sections of this report. Estimates based on carcasses do not always appear to be in error to this degree.

In Table 4 spawning escapement estimates based on the four different estimation procedures are given. The carcass derived estimates should be considered minimal estimates because coho may back downstream before dying and because tagged carcasses appear to have a higher retention rate than the general population of carcasses, especially at low spawning density levels. With all AUC live fish based estimates, it was assumed that the spawning life of a fish was eleven days based on work done in Oregon (Beidler and Nickelson, 1980). This appeared to be inappropriate for some coho streams.

As with the density of redds and live fish, population estimates show that lower numbers of fish spawned in 1990-91. The degree to which the 1990-91 populations are lower seems fairly consistent between both natural and artificially produced runs, indicating that they were likely negatively affected by ocean conditions. The surveys in Caspar Creek and Little River covered essentially all the spawning area and the low population estimates reflect very poor spawning numbers for streams with a recent history of an established coho population. The Pudding Creek run seemed to have done relatively better in 1990-91 based on the redd based range. The Hollowtree Creek estimates on Table 4 need to be added to the trap count on Table 3 for complete Hollowtree Creek estimates and indicates that this run was about half as large in 1990-91.

The timing of coho and chinook runs in 1991-92 and 1990-91 are shown in Tables 5 and 6. Here the number of carcasses tagged and redds flagged can be compared between years and streams. The months are divided into four (roughly week) periods. In 1991-92 redds and carcasses were found as early as the 3rd week of December in Caspar, Ten Mile and South Fork Noyo but not until the first week

of January in Pudding Creek. A much delayed run occurred in 1990-91 where redds and carcass were not found until the second week of January in most streams, and in Pudding Creek, not until the second week of February.

Since steelhead do not necessarily die after spawning and the stream surveys were conducted only through late February or early March which leaves out the later portion of the steelhead spawning period) steelhead populations could not be estimated in this study. From the number of steelhead carcasses, live steelhead observations and February redds some inference of steelhead abundance is possible. The highest density of live steelhead was found in Caspar Creek in February 1990-91 where 1.8 steelhead/survey mile were observed in the lower section and 1.1 steelhead/survey mile in the North Fork. In February of the same year lower Hollowtree and lower Pudding Creek had 0.32 and 0.29 live steelhead/survey mile, respectively. The following year, 1991-92, no steelhead were observed in Caspar or lower Hollowtree Creek in February. Instead, the highest density were observed in upper Hollowtree at 0.79 steelhead/survey mile where none were observed the year before. In 1991-92 the second highest February steelhead density (0.71) was observed in lower Pudding Creek. As for February redd counts not attributed to coho, the highest densities were also found in Caspar and it's North Fork in 1990-91 where 11.7 and 10.5 redds/reach mile were found. The second highest that year was in lower Pudding Creek at 6.9 redds/reach mile and third, lower Hollowtree at 4.4 redds/each mile. For 1991-92, tributaries in the Ten Mile River had highest redd densities. The Little North Fork had 6.0, Bearhaven Creek 4.8, and Clark Fork 3.2 redds/reach mile. Upper Pudding Creek also had 3.2 and that was followed by the North Fork Caspar which had 2.6 redds/reach mile.

The redd densities seem to follow steelhead counts fairly closely. One exception was the highest live steelhead density in upper Hollowtree in 1991-92 where the redd density was comparatively low

at 1 redd/reach mile. The other exception being the relatively high Ten Mile tributary densities in 1991-92. Since Ten Mile surveys, because of a lower survey priority, tended to be done after other streams were completed, the peak time of steelhead spawning probably occurred several days prior to surveys. It is interesting to note that these were the only areas surveyed in the Ten Mile River during the month of February.

The flow and turbidity data found for 1991-92 are presented in Tables 7 and 8 and in Table 9 for 1990-91. In 1991-92 flow conditions allowed fish to migrate into upstream areas in early January and heavy rains presented conditions which made spawning surveys impossible for two weeks in mid-February. In 1990-91, water flows were quite low due to drought conditions which persisted throughout most of the spawning period except the first week of February. These low flows prevented salmon from spawning in many customary spawning areas and probably delayed spawning in others. The low numbers of fish spawning in 1990-91 in some cases may have been the result of low flow conditions. The chinook run may have spawned in the South Fork Eel rather than enter Hollowtree Creek. There were salmon redds observed just below the mouths of South Fork Noyo, South Fork Garcia, Little North Fork Gualala which would indicate that fish were destined for these tributaries but because of low flows spawned in the mainstem instead. For streams that empty directly into the ocean like Caspar, Pudding, Wages, DeHaven Creek and Little River, this option was not available. Some fish did spawn in Caspar prior to January 16th before any major rain fell, indicating, at least for Caspar Creek, that the lower reaches were accessible under the low flow conditions. Access into Caspar Creek probably came during high tide periods.

Table 3. Number of trapped salmon released into areas above the Egg-taking Station on the South Fork Noyo and upper Hollowtree Creek and the populations estimated by the CR model, AUC Carcass and AUC Live methodologies and based on Redd Counts

Area	Year	Number Released	CR Model	Population Estimates		
				AUC Carcass	AUC Live	Redds
SF NOYO	91-92	216 coho	28	33	91	NE
HOLLOWTREE	91-92	100 coho	NE	NE	NE	NE
		82 chin.	NE	NE	NE	NE
" "	90-91	39 coho	6**	3**	8	12-50*
		15 chin.	1**	1**	5	
		3 Steel.	0	0	3**	NE

NE - No estimate made.

* This estimate takes into account the fact that no chinook females had been released and ignores any redds that may have been dug by the female steelhead. These are minimum estimates due to the fact that not all areas were surveyed.

** These estimates include fish that are believed to have passed into the survey area after the counting station was removed.

Table 4. Spawning estimates of Coho and Chinook Salmon
Based on CR model, AUC Carcass, Live counts
And the number of Redds by year.

Stream	Season	Species	CR	Population Model		
				AUC Carc.	AUC Live	Redd
Low. SF Noyo	1991-92	coho	151	100	210	172-688
" "	1990-91	coho	80	87	102	63-254
Low. Hollowtree	1991-92	chin.	205	261	350	189-338
" "		coho	13	10	15	11-22
" "	1990-91	chin.	4	3	3	9-38
" "		coho	5	4	3	7-28
Caspar Creek	1991-92	chin.	4	5	-	-
" "		coho	55	80	15	49-196
" "	1990-91	coho	0	0	2	8-28
Dehaven Creek	1990-91	-	0	0	0	0
SF Garcia	1990-91	coho	0	0	0	0
LNF Gualala	1990-91	coho	0	0	0	0
Howard Creek	1990-91	-	0	0	0	0
Little River	1991-92	coho	14	9	23	8-32
" "	1990-91	coho	0	0	0	2
Pudding Creek	1991-92	coho	28	45	40	26-102
" "	1990-91	coho	11	10	11	20-74
Ten Mile River	1991-92	chin.	-	-	-	51-154
" "		coho	-	-	-	14-42
Wages Creek	1990-91	coho	2	2	0	2-8
Willits Creek	1991-92	-	0	0	0	0

Table 5. Number of Carcasses Tagged and Redds Observed during the 1991-92 survey by week.

Stream	DECEMBER		JANUARY				FEBRUARY				MARCH	
	3RD	4TH	1ST	2ND	3RD	4TH	1ST	2ND	3RD	4TH	1ST	2ND
COHO CARCASSES												
Caspar	0	0	2	11	7	0	0				0	0
Holl.T.	0	0		0	1	12	9				1	1
L.River			1	5	1	1	1	0			1	0
SF.Noyo	5	6	15	30	49	11	22	0			1	0
Pudd.Cr			0	3	3		4				1	0
Ten Mile	0	0	0	1	1	1					0	
CHINOOK CARCASSES												
Caspar	0		0	0	1	0	0				0	0
Holl.T.	0	1		6	19	103	19				0	0
Ten Mile	0	0	1	0	1	9					0	0
STEELHEAD CARCASSES												
Caspar	0	0	0	0	2	0	0				0	0
Holl.T.	0	0		0	0	1	0				1	0
SF. Noyo	0	0	0	0	0	1	0	0			0	0
Ten Mile	0	0	0	1	1	0					1	0
NUMBER OF REDDS												
Caspar	10	1	53	29	5	0	2				4	2
Holl.T.	0	34		46	28	75	37				20	2
L.River		4	6	4	2	1	3	0			3	13
SF.Noyo	33	173	113	104	8	18	42	4			9	13
Pudd.Cr			6	21	13		25				24	8
Ten Mile	1	3	1	38	22	25		0			71	

Table 6. Number of Carcasses Tagged and Redds Observed during the 1990-91 survey by week.

	DECEMBER			JANUARY			FEBRUARY			MARCH	
Stream	4TH	1ST	2ND	3RD	4TH	1ST	2ND	3RD	4TH	1ST	2ND
COHO CARCASSES											
Holl.T.			0	0	0	0	1	3	1		
SF Noyo			7	19	7	12	3		3		
Pudd.Cr			0		0	0	1	4	2		
Wages Cr			0	0	0	1	0				
CHINOOK CARCASSES											
Holl.T.			0	0	0	0	3	1	2		
STEELHEAD CARCASSES											
Caspar				0	0	0	0	1	0		
Pudd.Cr			0		0	0	1	2	0		
Wages Cr			0	0	0	0	2				
NUMBER OF REDDS											
Caspar				14	0	10	12	23	10		
SF Garcia						0	1	1			
LNF Guala.						0	2	0			
Holl.T.			0	17	30	9	31	32	5		
SF Noyo			60	19	18	14	6	5	5		
Pudd.Cr			0		2	5	34	16	8		
Wages Cr			0	3	1	1		7			

Table 8. Estimated carcass visibility depth (cm.) found by survey week during 1991-92 spawning surveys

STREAM	WEEKS- SUNDAY THROUGH SATURDAY										
	DECEMBER			JANUARY			FEBRUARY			MARCH	
	15_21	22_28	29_4	5_11	12_18	19_25	26_1	2_8	9_15	23_29	8_14
Caspar Creek											
Mainstem	110	130	>100	40	>100	)120		110			80
North Fork			>100	20	>100	>120		110		70	100
South Fork				35	>100	>100				80	50
Middle Fork				25							
Hollowtree Creek											
Below Weir	>200		90	>140		>100	>100	>160			>160
Above Weir				>140	>100	>100	>140	>140		95	>140
Huckelberry Cr.										95	
Redwood Cr.						>100	>60				
Michaels Cr.							>60				
Bond Cr.										>100	
Little River			>120	25	>120	>120	45	95	20	>90	>90
South Fork Noyo											
Lower	>140	>200	70(140)	90	>140	>120	>140	>100		55	>160
Kass Creek			>100	75	>50		75	90	15	50	95
North Fork				20(100)	>140		95	>100	30		>160
Upper			85	50	120	120	40	110	35	75	90
Parlin Creek						>120					
Pudding Creek											
Lower			65	30	75	70		35		55	
Upper				45				>100		70	>100
Ten Mile River											
North Fork	130	>160	70		>120						
Bald Hills C.							>100				
Little N. Fork				70	>120		>100		50	>100	
Middle (Clark)		>100			>100	>120	85			140	
Bear Haven Cr.					>90				40	>100	
South Fork					>100		>100				
Campbell Cr.				70			>100				
Willits Creek				80							

Table 9. Flow and turbidity data taken during 1990-91 surveys. Flows are given in cubic feet per second and depth at which carcasses were visible are shown as observable depth in centimeters.

STREAM	WEEKS SUNDAY THROUGH SATURDAY													
	JANUARY							FEBRUARY						
	JANUARY							JANUARY						
	13_19	20_26	27_2	3_9	10_16	17_23	24_2	13_19	20_26	27_2	3_9	10_16	17_23	24_2
	-----FLOW(CFS)-----							-----OBSERVABLE DEPTH (cm)'s-----						
DeHaven Creek	7.8		3.9	16.2				>92		>75	70			
Caspar Creek														
Mainstem	6.0	2.8	2.1	18.1	3.9	2.8	2.1	>120	>120	120	50	>120	>100	>120
North Fork		1.4		9.5	2.1	1.8	1.1		>100		50	>140	>120	>100
South Fork		.4		1.8	.7	.4	.4		>100		47	>60	>120	>100
Garcia River, SF			1.8	6.1	2.8					>100	>100	>120		
Gualala River, LNF			2.1	6.7	4.2					>120	>130	>100		
Hollowtree Creek														
Below Weir	31.4		15.5	90.8	42.1	32.5	21.2	>220		>120	>190	>150	>200	>120
Above Weir		15.2	11.3		24.7	18.1	10.6	>110	150	>150	>180	>150	>150	>200
Huckelberry Cr.		2.8	1.1						>100					
Redwood Cr.			1.1		4.9	3.2				>100		>100	>130	
Howard Creek	4.2	3.5		8.8				>110	>95		100			
Little River	3.2	6.1		3.9				>100	>100		>120			
South Fork Noyo														
Lower	11.3	3.9	4.6	34.3	7.8	5.9	6.7	>180	>120	>150	40	>150	>150	>130
Kass Creek	.7	.7							>120					
Pudding Creek														
Lower	4.9	1.8	1.4	21.2	10.6	4.9	3.2	70		>180	35		>130	>130
Upper				12.1	3.9		1.4				73	>100		>100
Wages Creek	5.3	8.8	9.5	29.7	5.6			>100	>100	>120	90	>120		

TABLE 10. The Number of Coho Salmon Released above the Egg-taking Station in 1991-92 and the Number of Live Fish, Redds and Carcasses Tagged by Date in the North Fork, South Fork from the Egg-taking Station to Parlin Creek (SF R-1), South Fork from Parlin to Dam Spillway (SF-R2), Parlin and Bear Creeks.

	DATE																										
	DECEMBER										JANUARY										FEBRUARY					MARCH	
	28	29	30	31	1	2	3	4	5	6	7	8	9	10	11	13	14	15	20	22	28	29	4	5	11	25	10
Num. Rel.#	39	61	16	3	0	0	9	0	4	45	25	8	6	0							0	0				0	
Number of Live Fish Observed																											
North Fork					0				0		18				4					2		0		0	0	3	
SF R-1					23						14				40				16	9		6		2	3	1	
SF R-2											14						0	1			0		0	0	0	0	
Parlin Cr.											4						3	1									
Number of Redds Counted																											
North Fork					4				0		23				9					4		2		3	1	3	
SF R-1					23						16				16				4	4		10		1	3	8	
SF R-2											12						3	1		1		2	0	1	1		
Parlin Cr.											2						0	0									
Bear Creek															6												
Number of Coho Carcasses Tagged																											
North Fork					3				0		1				0					1		0		0	0	0	
SF R-1					0						1				2				2	2		8		0	1	0	
SF R-2											0						0	0			0		0	0	0	0	
Parlin Cr.											0						0	0									

* From 1/10 to 1/23 the trap was left open but it is believed that no fish passed because flow conditions were very low. From 1/23 to 2/1 coho yearlings were held in egg-taking station. No spawners were trapped although it is possible that a fish could have jumped the screen and passed upstream. From 2/2 until 2/23 the trap was left open. On 2/24 trap was operated while imprinting yearlings for 10 days. - No fish were trapped but it is possible a fish could have passed.

TABLE 11. The Number of Coho Trapped at the Egg-taking Station and the Number of Live Fish and Redds found and Carcasses Tagged in the Lower South Fork and Kass Creek in the 1991-92 Survey.

		DATE																												
		DECEMBER							JANUARY							FEBRUARY							MARCH							
		17	23	28	29	30	31	1	2	3	4	5	6	7	8	9	10	11	12	14	15	22	29	4	5	12	25	26	10	11
Coho	Trapped			134	165	55	3	0	0	11	1	10	79	34	15	2	0	0	0				0				0	0		
		Lower South Fork Noyo River																												
Live Fish		0	20			201				89							74				43	6	24	7			0		0	
# Redds		33	61			117				25							14				16	3	6	21			0		0	
Carcasses		0	5			6				8							17				18	25	5	14			0		1	
		Kass Creek																												
Live Fish								15									16				3		0		0			0		0
* Redds								39									5				4		3		7			4		1
Carcasses								4									9				4*		1		0			0		0

* Includes one tagged steelhead

The turbidity of the water can influence the ability of surveyors to see fish or carcasses. The streams which tended to have the poorest visibility were lower Pudding Creek and the upper South Fork Noyo. The turbidity in lower Pudding Creek originates in Little Valley Creek and appears to be a natural algal coloration which is present year-around. The South Fork Noyo turbidity is associated with an upstream reservoir.

SOUTH FORK NOYO RIVER

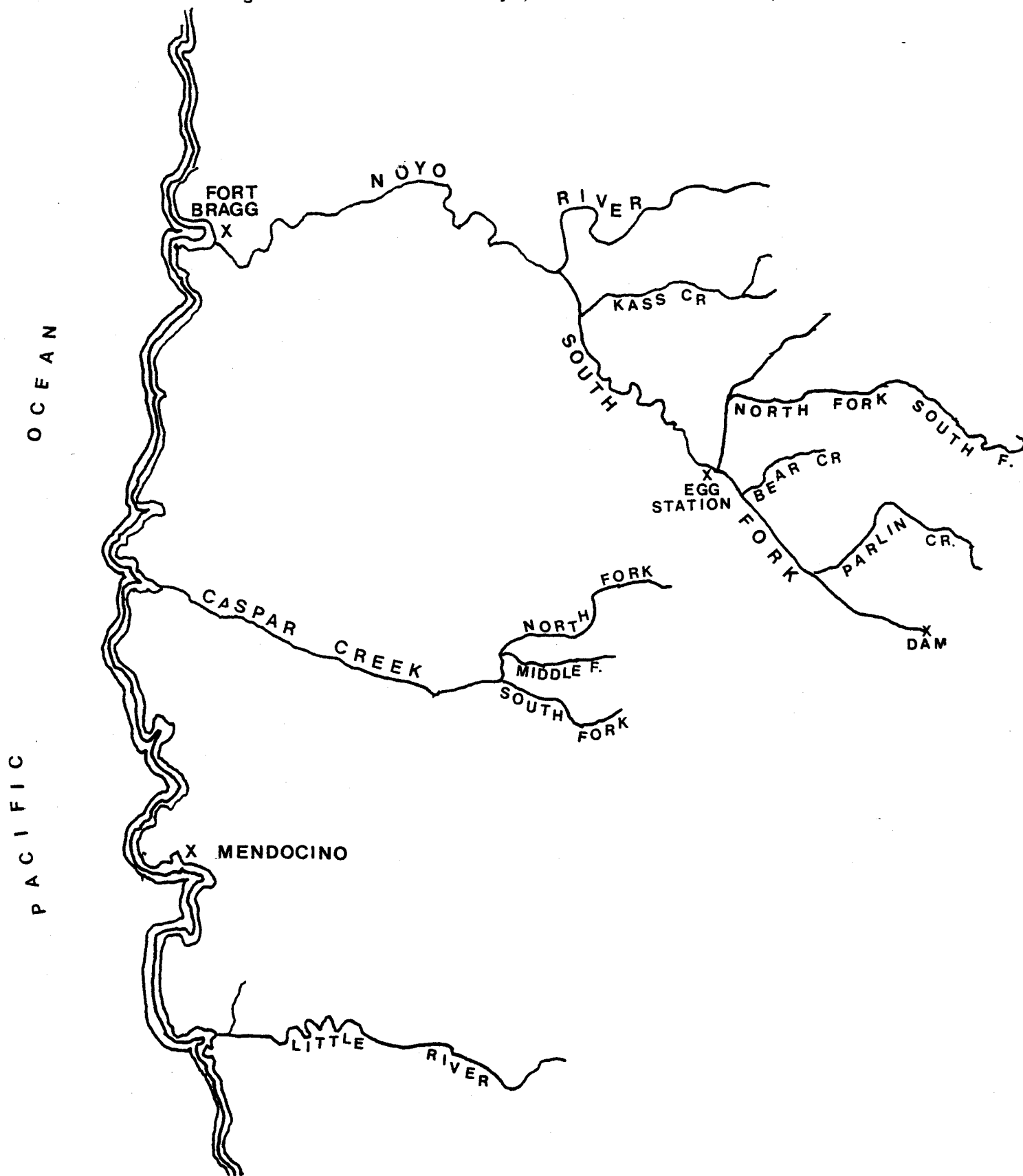
An egg-taking facility located on the South Fork is operated by the California Department of Fish & Game (CDFG). Coho salmon are trapped here to collect eggs to support an artificial propagation program. The juvenile coho are planted as yearling into the South Fork and the trapping station is used as an imprinting facility. Surveys of the on the South Fork Noyo River are divided into two areas; those above the Egg Taking Station and those below (see map Figure 2). The lower area has two reaches, Kass Creek (1.6 miles) and the lower South Fork (3.9 miles). The upper area is divided into several reaches; the North Fork (3.4 miles),. South Fork from the egg-taking station to Parlin Cr.(2.4 miles), Parlin Cr. to Dam spillway (2.4 miles), Parlin Cr.(1 mile) and Bear Cr. (0.5 miles). We did not survey the upper areas of the North Fork or Parlin Cr. this year because little spawning activity was found below these areas to indicate that spawning activity in upper areas was likely.

1991-1992 SURVEY

Between December 28th and January 9th, 216 coho (81 males, 38 females and 91 grilse) were released above the egg-taking station.

From the 10th to 23rd of January the trap was left open during a

Figure 2. South Fork Noyo, Little River and Caspar Creek



period of low water and it is believed that no fish passed through (Alan Grass CDFG, pers. comm. 1992). From the 23rd to February 2nd, no additional fish were released upstream. From February 2nd. to February 23 fish could pass the egg-taking station without being counted.

Surveys Above the Egg-taking Station

In Table 10 the chronology of coho releases can be compared to the numbers of live fish observed, number of redds counted and carcasses tagged in areas above the egg-taking station. On December 2nd the two major spawning areas above the station were first surveyed. Twenty-three live fish were observed in the South Fork Noyo between Parlin Creek and the Egg Station; no live fish were found in the North Fork. These surveys were conducted 3 to 5 days after 119 coho (19 females) had been released. There were 27 redds observed and 3 carcasses found.

By the 5th of February a peak live count of 52 fish had been observed, 142 redds counted and 20 coho carcasses tagged, including 7 adult (3-year-old) males, 2 adult females, 2 male grilse (2 year-old), and 1 female grilse. Females carcasses were 25 percent of the carcasses found. This compares to 18 percent of released being coho. After February 5th there were 21 redds counted, one coho carcass tagged and a total of nine live fish observed (1 coho and 8 steelhead). Of 16 carcasses tagged with jaw tags, 7 (44%) were recovered.

Of the fish released above the egg-taking station, eight percent were found as carcasses. The peak live fish count was only 26 percent of the number released. The measurable adult carcasses found represented only 8.8 percent of the number of adults released and the grilse carcasses represented 3.1 percent of grilse released.

Surveys Below the Egg-taking Station

The results of lower South Fork Noyo surveys and the number of fish taken at the egg-taking station are shown on Table 11. It can be seen that spawning occurred prior to the first survey on Dec. 17th where 33 redds were observed.

In the lower South Fork a peak live count of 201 was observed on December 30th. There were 275 redds by January 29th and 21 more in February. There were 99 carcasses tagged. The numbers by sex, age and lengths of these fish are shown on Table 12. including 19 adult males (mean fork length 64.5 cm), 29 adult females (63.8 cm), 22 male grilse (47.1 cm) and 2 female grilse (48.0 cm). The adult males ranged in length from 45 to 71 cm., adult females from 49 to 73 cm., male grilse from 42 to 50 cm. and female grilse ranged from 46 to 50 cm. Adult males and females were 26 and 40 percent of carcasses while male and female grilse composed 31 and 3 percent of carcasses, respectively. Of 92 carcasses tagged with jaw tags, 36 (39%) were recovered.

Table 12. The Number (%), Average Length and Range (fork length cm.) of coho salmon carcasses found in the lower South Fork Noyo in 1991-92 by age and sex.

Age	MALES			FEMALES		
	Number	Ave.Len.	Range	Number	Ave.Len.	Range
2	22	47.1	42-50	2	48.0	46-50
3	19	64.5	45-71	29	63.8	49-73

The peak count of 201 live fish represents fish that did not enter the facility during that early run since only 3 fish entered the station over the next three days. A second run of fish came through

from the 3rd to 9th of January, coincident with rain and increased flows. Relatively few additional fish spawned in the lower section after the initial period, evidenced by low numbers of new redds found. A surprisingly few live fish were observed on January 3rd, just 4 days after observing 201 even though visibility improved markedly. Between the live fish counts in lower South Fork and Kass, carcasses tagged and fish entering the Egg Station, only 130 fish can be accounted for on January 3rd. A minimal estimate (minimal due to turbidity conditions when the 201 live fish were observed) of 71 fish are unaccounted for after just 4 days. It is possible that some of these fish backed out of the South Fork and into the Noyo River.

In Kass Creek, surveys were conducted 8 times from January 2nd to March 11th. The peak live count was on January 10th at 16 coho. There were 51 redds were counted in January and 12 in February. A total of 15 carcasses were tagged including two adult males, two adult females, two male grilse and one steelhead (48 cm.in length). Of 14 tagged with jaw tags, 5 (36%) were recovered

Population Estimation

In the area above the egg-taking station 38 females (adults and grilse) were released. Redds found in the upper South Fork surveys through February 5th are considered to have been made by fish counted through the egg-taking station. Those found after that date are considered to have been made by fish that entered upper South Fork uncounted. With a count of 142 redds and a release of 38 females, the number of redds constructed per female would be 3.7 if all of the females released did spawn above the egg-taking station.

In the lower South Fork, there were 296 redds by February 5th (none after this date). If this ratio (3.7:1) were to hold for areas

below the station, the number of spawning female coho would be 80. For comparison purposes, it would be good to be able to convert the number of females into a total run estimate. Females composed about 43 percent of the carcasses found in the lower South Fork this year, indicating that the total coho run was about 186 fish. This number is slightly below the peak live fish count of 201, indicating that the 3.7 figure may be too high for this area.

In the area above the egg-taking station, 21 redds were found after February 5th and during this period one live coho and 8 live steelhead were observed. Using the percent composition of coho in live fish counts to indicate the percentage of February redds that were coho redds, two coho redds would be estimated. Only a single pair of coho would be expected to have spawned in this area after February 5th.

To use the carcass retention (CR) model to estimate the number of spawners above the egg-taking station, an estimate of the average daily carcass retention must be made. Table 13 shows for the combined areas above the egg-taking station the number of carcasses tagged and recovered under the survey intervals experienced. By lumping the days from 5-9, the recovery rate is 40 percent after seven days. The longer intervals had no recoveries although it is apparent that had more fish been tagged, some recoveries would have occurred. A daily retention rate of .86 was selected as it would pass near the 40 percent retention estimate after a seven day interval.¹ With this retention rate the CR model estimates a spawning population of 28 fish.

¹ The .86 retention rate was selected here since it matches that found for the lower South Fork Noyo this year which was based on significantly greater numbers of tagged and recovered carcasses.

Table 13. The Number of Days between Surveys, Number of Jaw Tags and the Number of Tags recovered in the Upper South Fork Noyo in 1991-92

Days	Tagged	Recoveries
5	1	0
7	13	6
9	1	0
12	1	0
14	1	0
16	1	0
19	1	0

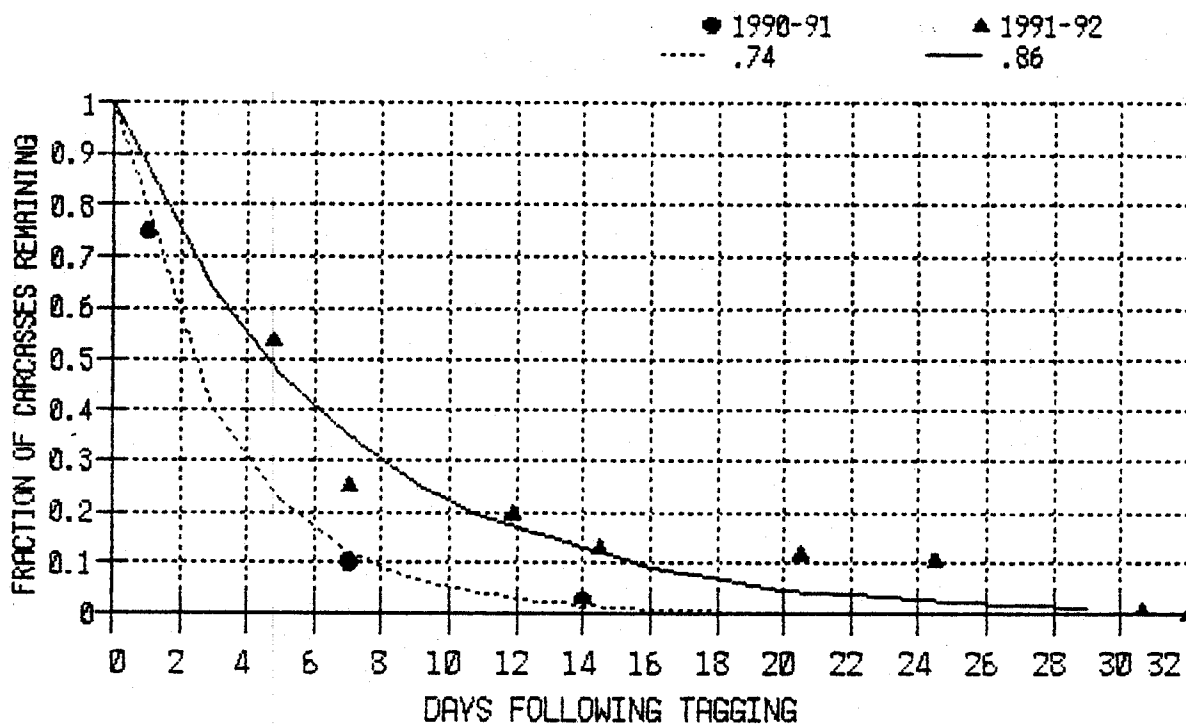
In the lower South Fork Noyo the carcass retention rate was .86 as well. The fit of the retention curve to the South Fork data is shown in Figure 3, where the retention curves for both the 1991-92 and 1990-91 South Fork data are shown. Using the retention rate selected, the (CR) model estimate for the lower South Fork is 151 coho. In Kass Creek, this model would estimate that 24 coho spawned.

The spawning population above the egg-taking station based on the AUC method is 33 coho where a average carcass duration of 4.7 days was estimated. An AUC estimate based on live fish is 91 coho.

In the lower South Fork, the AUG carcass method would estimate 100 coho spawners and from live counts an estimate of 210 is obtained. In Kass Creek, the estimate of coho spawning is 22 fish with both live and carcass methods.

The CR and AUC carcass model estimates are about 170 fish short of the 216 coho released above the station and the live fish estimate

Figure 3. Carcass Retention Data for the Lower South Fork Noyo and associated Daily Retention Curves for years 1991-92 and 1990-91.



is 125 fish short.

Part of the problem may be due to the ability of fish released above the station to back down and drop over the dam. In the initial planning of this survey, it was intended that fish released above the egg-taking station would be marked so that if they dropped back down they could be identified. As this did not happen, it could be assumed that fish released did spawn above but drifted downstream to die, or, they did not spawn above and instead backed down to spawn below. A study conducted in three tributaries to the Alsea River in Oregon found that of the coho released above a trap only 47, 57 and 61 percent of the coho originally released upstream eventually spawned in those same tributaries (Moring 1975). The majority of these coho returned back downstream as live unspent fish. The possibility that these fish did construct redds but did not actually spawn is quite possible. Briggs (1954) found that when examining coho redds 54 percent were false or trial redds containing no eggs.

Another potential problem is that the carcass retention rates of tagged carcass may not represent the retention rates for the population as a whole. If a high proportion of carcasses are being removed from the stream by a predator or scavenger which has a particular territory, the life of a carcass within that territory may be much shorter than in an area which doesn't have that predation. This would be especially true where the spawning population is small and the predator/ scavenger can remove most or all of the carcasses. The surveyor in this instance would be most likely to find fish carcasses outside the territory of the animal and the carcasses found would have a longer retention rate than a carcasses found within the territory.

The utility of the AUC live fish estimate based on an assumed "residence time" (Ames and Phinney 1977) or "survey area residence time" (Irvine et al., 1992) for coho is limited. The residence

time of coho have been found to vary considerably (Bocking et al. 1988; Perrin and Irvine 1990; Irvine et al. 1992; English et al. 1992) and AUC estimates are very sensitive to observer efficiency (English et al. 1992). Here though, the AUC live estimate comes closer to the number released than does the carcass estimates.

The estimates for the lower South Fork Noyo are 151 and 100 based on the CR and AUC carcass method. The carcass based estimates for the area above the egg-taking station are about 170 less than the number released. This indicates that the low estimates based on carcasses in the area above the egg-taking station is not simply due to spawners dropping back down over the dam since carcass estimates in the lower section are lower than the 170 fish shortage to start. Assuming the redds/female estimate of 186 coho is correct, there are about 200 carcasses that are unaccounted for by the CR model and 256 using the AUC carcass method.

1990-91 SURVEY

In the lower South Fork Noyo, surveys were conducted seven times from January 16th, through February 27th. No surveys were conducted above the egg-taking station because there were no coho released above the station in 1990-91. The egg-taking station did receive 89 adult coho and 56 grilse. (Alan Grass, CDFG, pers. comm., 1992). A peak live count of 138 was taken on December 16th. There were 51 coho carcasses tagged including those in Table 14 that were aged for males and females. One of the males (60 cm.) appeared to be a four-year-old fish having had two years of freshwater growth. Redd counts were 97 in January and 30 in February. Of 30 jaw tagged carcasses, 6 (20%) were recovered.

The estimated carcass retention in 1990-91 was considerably lower than in 1991-92, 74 compared to 86 percent (Figure 2.). This may have been due to the lower number of fish spawning in 1990-91, resulting in a higher predation rate. The estimated spawning

population in the lower South Fork Noyo based on CR model is 80 coho. The AUC carcass and live coho based estimates are 87 and 102, respectively.

Table 14. The Number(%), Average Length and Range (fork length cm.)of coho carcasses found in the Lower South Fork Noyo in 1990-91 by age and sex

Age	MALES			FEMALES		
	Number	Ave.Len.	Range	Number	Ave.Len	Range
2	2	45	43-46	1	40	-
3	3	66	62-70	5	59	52-62
4	1	60	-	0		

Since there were no live steelhead or steelhead carcasses observed, 11 redd counts in the lower South Fork are assumed to be coho. The 127 redds would indicate 34 females (@ 3.7redds/female) or 68 total (half of carcasses found were female) coho spawned. Two surveys in Kass Creek in the later half of January found no evidence of fish use.

The fact that the redd/female estimate of spawners is about 80 percent of that produced by carcass models is not be expected since carcass models are expected to produce a minimum estimates. To produce similar numbers to carcass and live fish models, the number of redds/female would have to be between 2.5 to 3.2. The number in the lower South Fork this year is not influenced by spawners backing down from above the egg-taking station. This may give an indication that the redds/female ratio estimated for the 1991-92 season may not be appropriate for this section of the river or for this year's run and that perhaps an estimate as low as 2.5 may be more appropriate.

REVIEWING POPULATION ESTIMATES

Using the number of fish released above the egg-taking station as a test of population estimation models has not been conclusive. There is no way to explain away all of the discrepancies found. If the 3.7 redds/female is correct and 186 coho spawned below the station in 1991-92 the 151 CR estimate and the 210 AUC live fish estimates are quite close. But because the estimates above are low by 188 fish with the CR model and AUC live fish estimates are 125 fish low these models appear to underestimate populations significantly. The CR model estimate was only 13 percent of the number released and the AUC live fish estimate was only 42 percent. If the population estimates below the station are as low as above, the spawning population could be 620 based on CR model shortfall and 500 based on AUC live fish shortfall. For this to be correct, female coho would only be making 1.1 redds on average which does not come close to that observed above the egg-taking station.

There appears to be no way to resolve this discrepancy at this time. For each of the streams surveyed, this report will give the results from each model and to account for the possible error that these models produce, the redd/female estimates will vary using estimates ranging from 1 to 4 redds/female with an exception for streams where chinook make up a major portion of spawning (see Hollowtree Creek section of this report). In the lower South Fork this range produces estimates from 172 to 688 in 1991-92 and 63 to 254 in 1990-91.

RESTORATION EVALUATION

The runs entering the South Fork Noyo to spawn were the largest of any of the coho runs surveyed. It is not known what portion, if any, of these coho are produced through natural production because they were not marked nor identifiable through scale analysis,

therefore, the benefits in terms of increased natural production from these relatively large South Fork Noyo spawning escapements can not be evaluated. Clearly, in terms of value to fisheries, these planting of yearling coho are beneficial.

The number of coho trapped in the South Fork Noyo were 1,200 to 3,000 during the first 13 years of operation at the egg-taking station (Snyder and Sanders 1979). Since 1977-78, the numbers have averaged much less and exceeded 1200 only twice (Brown et al. 1993). Reduced coho populations since 1976 have occurred throughout Oregon, Washington and California and have been associated with poor ocean survival conditions (Gunsolous 1978; and Bottom et al. 1986) which may be the result of a change in the ocean circulation pattern which occurred in the winter of 1977 (Norton et al. 1985). The 1011 coho that were trapped in the South Fork Noyo in 1989-90 was the forth largest run since 1976 while the 145 in 1990-91 was the second lowest on record.

The South Fork has undergone habitat restoration work as have many salmon streams in California. In the early 1960's, of an estimated 16 miles of habitat in the South Fork and it's tributaries, only one mile was classified as in satisfactory condition (Holman, G., & W.A. Evans, 1964) and 15 miles were "improved", or in other words, had log jams removed. Holman (1964) reported that the main benefit of the log jam removal was not, contrary to popular belief, the removal of impassable barriers but instead creating conditions in which winter lows could remove silt and gravel deposited behind log jams. He also reported a problem that there was a tendency to be over meticulous in clearing of small unimportant debris. These activities probably had a negative impact on coho production by removing from the stream the large woody debris and cover; the critical habitat required by coho (Hoar 1951; Hartman 1965; Ruggles 1966; Bustard & Narver 1975; Nickelson et al 1992; McMahon and Holtby 1992).

HOLLOWTREE CREEK

Surveys of Hollowtree Creek are divided into two primary areas; 1, below the Hollowtree Fish Hatchery and 2, areas above the fish hatchery (Figure 4.). The area below consists of a 7.8 mile reach of the mainstem of Hollowtree Creek from it's mouth to the hatchery. At the fish hatchery site is a weir where fish can be collected for fish propagation or counted and released above. The area above is divided into several reaches. Some upstream areas were surveyed fairly consistently while others were not surveyed at all.

The mainstem Hollowtree Creek from the hatchery to the confluence of Huckleberry Creek is 11.3 miles. No surveys were conducted on Hollowtree Creek or any tributary above Huckleberry Creek. The two primary reaches were from the hatchery to Eastside Rd. sign (2.5 miles) and from Redwood Creek up to Huckleberry Creek (4.6 miles). The 4.2 mile reach between the Eastside Rd. sign and Redwood Creek was not surveyed due the difficult survey conditions combined with the lack of spawning activity found in prior surveys. This area consists of steep canyons and deep pools where swimming or floating the stream was usually required. In the upper reach (Redwood to Huckleberry), surveyors did not always have sufficient time to complete the entire reach or came out at the wrong location so some surveys did not cover the entire reach. Several tributaries were surveyed, but none on a consistent basis. These included Redwood, Huckleberry and Michaels Creek.

1991-92 SURVEY

Lower Hollowtree Creek

The lower 7.8 miles of Hollowtree Creek was surveyed 9 times from November 25th to March 12th. No evidence of fish use was found until a survey on December 31st. The number of carcasses tagged

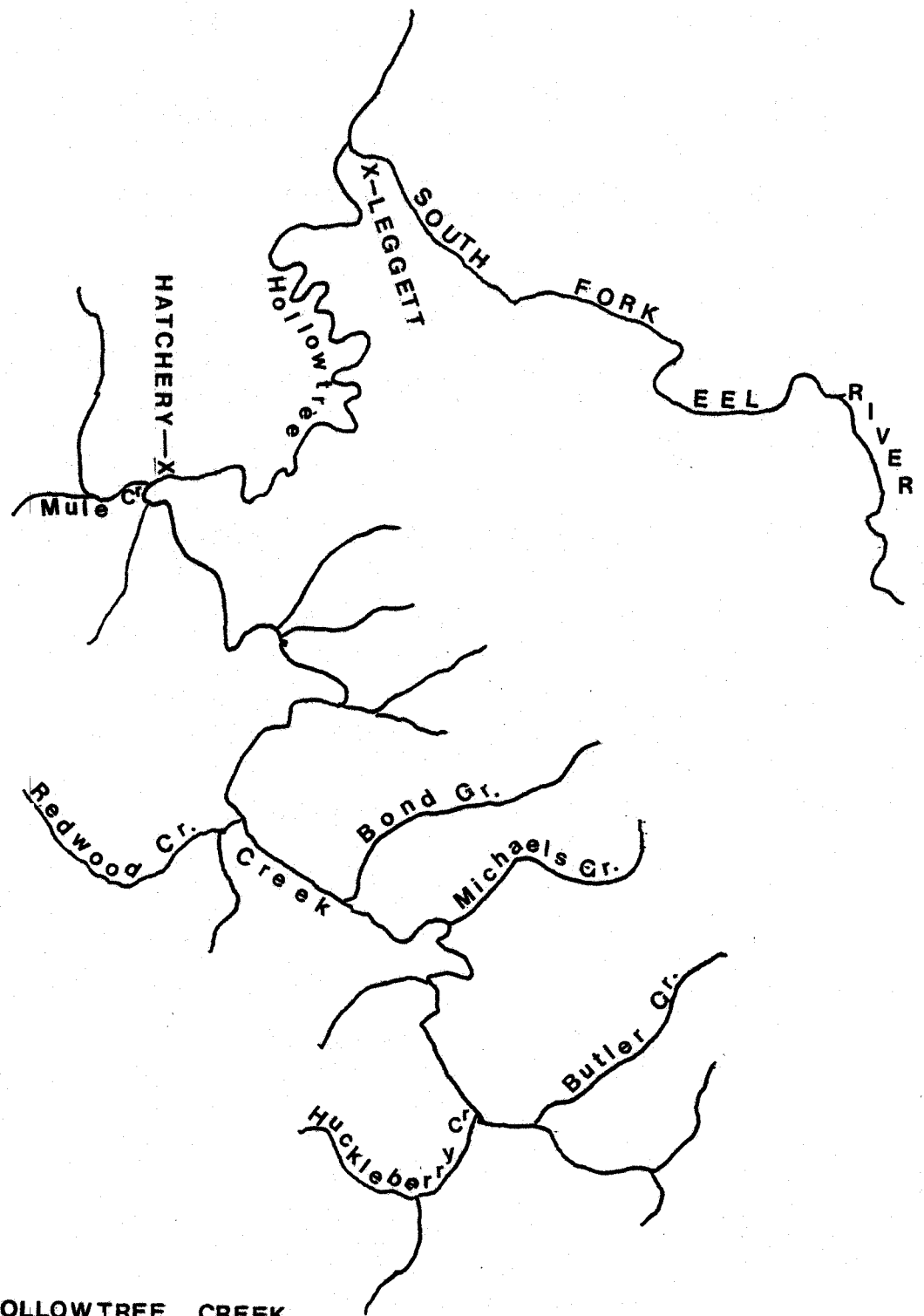


Figure 4. HOLLOW TREE CREEK

were 134 chinook, 9 coho and 2 steelhead. The peak live count was 189 on January 10th and there were 151 redds in January and 3 in February. Of 135 jaw tagged carcasses, 49 (36%) were recovered. The number, average length and range in length of aged chinook and coho carcasses found in Hollowtree Creek are shown on Table 15. The chinook run was heavily dominated 4 and 5 year old fish.

There was a single fish found with an adipose clip. (Tag number B61509). This was a 1987 brood Hollowtree Creek Hatchery release. The hatchery released a total of 51,670 adipose clipped chinook in 1987 out of a total of 189,914 fingerlings. Expanding for untagged fish would indicate that four hatchery released fish from the 1987 brood were in the sample checked for fin clipped fish; about 10% of the total returns of age four fish. This would indicate that between 12 and 18 of these fish spawned in Hollowtree Creek this year.

This section of river had a lot of bear sign. Fresh bear tracks were often observed. In one survey, about 15 or so carcasses were found along a bear trail which followed the stream. Some of these fish were badly decomposed and uneaten. There was one whole carcass found which had been previously tagged in the stream which the bear had hauled out but did not eat. Many of the carcasses had the top of the head and body missing with the remainder of the head, skin, backbone and tail still attached.

The estimated population of chinook salmon in lower Hollowtree is 205 fish based on the CR model based on a carcass retention estimate of 0.89. The coho estimate is 13 fish. The AUC carcass method estimates 261 kings and 10 coho. The AUC live method estimates 350 chinook and 15 coho spawning in lower Hollowtree Creek using the assumed 11 day average spawner residence time.

The 151 January redds, using the 1 to 4 redds per female range,

would estimate between 38 and 151 females. The percentage females in the carcass sample was 42 percent. This would estimate between 90 and 360 total salmon. Clearly, because 189 fish were actually counted on a survey, estimates below this figure are too low. This may indicate that chinook, at least in lower Hollowtree Creek, do not dig as many redds as do the Noyo River coho. With an absolute lower limit of 189 fish and since about 6 percent of carcasses were coho, the range based on redds is 11 to 22 coho and 178 to 338 chinook. The lower range is produced with a redd/female estimate of 1.8.

Upper Hollowtree Creek

At the Hollowtree Creek Hatchery, 76 adults and 6 jack chinook salmon were trapped. Of these 77, including 7 females, were released upstream. There were also 100 coho adults, including 52 females and 3 grilse, trapped and released. The first fish was trapped on December 28th and the last on January 11th. (Louis Hans, Salmon Restoration Ass. pers. comm., 1992). The trap was taken down on January 31st and fish could pass without being counted after this date.

The area immediately above the hatchery was surveyed 6 times from January 10th to March 12th, including one incomplete survey on January 23rd. There were 6 chinook and 3 coho carcasses tagged. The peak live fish count was 23 on January 10th. The redd counts were 6 in January, zero in February and 2 in March. Of the 9 jaw tagged fish, 5 (56%) were recovered.

In the area between Redwood and Huckleberry Creek 6 surveys were conducted between January 10th and March 12th. Two of these surveys were from Huckleberry to Bond (3.4 miles), two were from Huckleberry to Michaels Creek (2.2 miles) and one covered the area from Bond to Redwood Creek (1.2 miles). There were 7 chinook, 10

coho and 1 steelhead tagged. The steelhead was a 33 cm. 2-year old male with one year of ocean growth. The peak live count was on March 12 where 11 steelhead were seen. The total redd count was 14 in January and 39 in February. Of the February redd count, 32 of these redds were found on an early February survey that included

Table 15. The Number (%), Average Length and Range (cm.) of Salmon Carcasses found in 1991-92 Hollowtree Creek Surveys by age.

MALES				FEMALES		
Chinook Salmon						
Age	Number	Ave.Len.	Range	Number	Ave.Len.	Range
2	4 (5%)	49	46-52	0		
3	3 (4%)	71	10-71	1 (1%)	73	
4	16(21%)	91	83-98	23(30%)	91	78-94
5	19(25%)	94	82-102	8(10%)	93	85-98
6	1 (1%)	104		2(2%)	112	110-115
Coho Salmon						
2	4(25%)	45	41-48	0		
3	7(44%)	62	54-68	5(31%)	62	57-65

the area between Redwood and Michaels Creek (2.4 miles) which was not surveyed in the previous survey period; for this reason, these redds are considered to be January redds. Of 15 jaw tagged carcasses only one (7%) was recovered.

In Redwood Creek, surveys were conducted three times. In one of these (January 30th) a distance of 0.3 miles was surveyed above the bridge and the other two, January 23rd and February 6th, the area surveyed was from the bridge to the mouth (0.3 miles). No carcasses or live fish were observed. The total redd count was 6

in January and 2 in February.

In Bond Creek, a single survey on February 27th. found 13 redds, one steelhead carcass and one live steelhead. The steelhead was 74 cm. male. This fish could not be aged because all scales were regenerated but did appear to have had 3 years in ocean. This fish died as it tried to pass a falls but fell between boulders where it became trapped.

One survey on the lower one mile of Huckleberry Creek on February 27th found no evidence of fish use. It was also reported that this Creek was heavily silted.

The single survey in Michaels Creek on January 30th found five redds. No live fish or carcasses were found.

In the combined survey areas above the hatchery, the redd count was 63 in January and 24 after February 6th. There were 13 chinook and 1.6 coho tagged. These represent 17 and 16 percent of the numbers of chinook and coho released. One of these coho carcasses was found on February 27th and may have come up after the trap was taken down. Out of 24 jaw tags, 6 (25%) were recovered. The low recovery rate of tagged fish (25% compared to 36% below the hatchery) is due in some part, to the incomplete surveys.

Since Hollowtree was not surveyed above Huckleberry Creek, the 4.2 mile middle section, nor in tributaries such as Butler, Waldron and Mule Creeks, the numbers of fish released above the hatchery can't justly be used to test population estimates through the population models. Also, since the numbers of tagged fish could not always be identified as to which portion of the reach they were tagged (because of the incomplete surveys), no recovery rate could be established nor could the number of days between surveys be established for case in carcass models. What can be said is that about one redd was found for each female (coho+chinook) released.

For Hollowtree Creek the combined trap count and lower river estimates produce estimates that range from a low of 260 to 432 chinook and 111 to 122 coho this year.

1990-91 SURVEY

Lower Hollowtree Creek

Lower Hollowtree Creek was surveyed 6 times from January 18th to February 28th. Chinook salmon carcasses or remains were found from the first survey until February 20th and coho from February 8th to the 20th. There were 4 chinook, 3 coho and 3 unidentified carcasses tagged. The peak live count was 9 on February 20th, 6 of which, were identified as steelhead. The number of redds were 19 in January and 48 in February. Of 7 jaw tagged carcasses, three (43%) were recovered. There was only one measurable chinook salmon. It was an unusual fish, a 38 cm. 2-year old female. The two measurable coho were males 67 and 59 cm. in length. One of these was a three-year-old male the other possibly a four-year-old fish.

The CR model would estimate 5 coho, 4 chinook and 5 unidentified fish spawning in lower Hollowtree Creek. AUC carcasses method would estimate 4 coho, 3 chinook and 4 unidentified and AUC live estimates are 3 coho and 3 chinook.

For lower Hollowtree in 1991-92, the lower range of redd based spawning estimate was produced with a 1.8 redd/female estimate. The proportion of carcasses that were coho that year was only 6 percent. In 1990-91, of 7 salmon carcasses 3 (42%) were coho carcasses and so the 1.8 redd/female figure is likely too low since chinook produce considerably fewer redds than coho. The 1 to 4 range would be more appropriate here. The 19 January redds would indicate that between 10 and 38 salmon spawned. In February, out

of 14 identified live fish, 2 were coho and 2 were chinook and ten were steelhead. Using these proportions to allot redds, 14 would be salmon and 34 steelhead redds. The total number of salmon redds is estimated at 33, indicating the range for salmon is between 16 and 66. The coho range is from 7 to 28 fish and the chinook range is from 9 to 38 fish.

Upper Hollowtree Creek

At the Hollowtree Creek Hatchery trapping facility, 15 chinook were trapped and released upstream. There were no females trapped so all chinook were released. There were 39 coho trapped and released upstream: 17 males, 21 females and 1 grilse. There were also 2 male and 5 female steelhead trapped and released upstream. The trap began operation on January 6th and ended on February 10th. Chinook were trapped from January 14th to February 9th. and coho from the January 13th to February 8th (Louis Hans, Salmon Restoration Ass., pers. comm., 1992).

In this survey season, upper Hollowtree was divided into three sections; the trap to Eastside (2.5 miles), Eastside to Bond Cr.(5.4 miles) and Bond Cr. to Huckleberry Cr. (3.4 miles). Upper Hollowtree Creek surveys began on January 18th and continued until February 27th.

In the trap to Eastside section, one live fish was seen on February 1st, two redds by January 8th and two more on February 27th. One coho carcass was jaw tagged and a chinook tail punched on the 27th of February. In the middle section, one live coho was seen on February 13th and a total of 18 redds counted. No carcasses were tagged. In the upper section, Bond Cr. to Huckleberry Creek, the peak live count was 7 coho on January 28th. The redd count for this section was 38 of which 23 were found February 13th or earlier. One coho and one chinook carcass were tagged; one (50%) of these was recovered. The chinook carcass was recovered on February 27th

and was a two-year-old female 60 cm. fork length.

The two coho carcass found were females 63 and 65 cm. in length. Both of these coho, based on scales, appeared to be four-year-old fish having two years of fresh water growth. Four-year-old coho are not common (Moring 1975; Shapovalov and Taft 1954) and having both these females and possibly one of the two males found below the hatchery trap all being four-years-old is noteworthy. The only other four-year coho identified in this survey was of the South Fork Noyo this same season. Perhaps some environmental factor has lead to this unusual incidence of four-year coho this year.

In Huckleberry Creek, two surveys in the lower half mile of the Creek found no live fish or carcasses in the later part of January. There two redds were counted.

Three surveys were conducted in Redwood Creek between January 28th and February 20th. The first two surveys were in the lower 0.3 mile from the mouth to the bridge and the last extended upstream an additional 0.3 of a mile. No live spawners were seen. Two carcasses were tagged; one tail punch and one jaw tag. Neither fish could be identified to species. The single jaw tagged carcasses was recovered for a 100% recovery rate. A total of five redds were found.

The Hollowtree Creek trap was operated until February 10th. It is assumed that spawners and redds found through the February 13th survey were accountable to fish counted through the trap. Through this date 50 redds were found. This compares to 26 females (21 coho and 5 steelhead) released at the trap. At a very minimum, there were about 2 redds per female. Some tributaries and Hollowtree Creek above Huckleberry were not surveyed. With the low flow conditions that existed, most of the spawning would be expected to be in the mainstem of Hollowtree Creek. At least one fish spawned in the relatively smaller Huckleberry Creek so it is

likely that some spawning occurred in the section of Hollowtree above Huckleberry Creek. Some spawning would be expected in the lower portions of some tributaries like Bond, Michaels and Mule Creeks.

If the 3.7 redds/female ratio estimated for Noyo coho applies to Hollowtree Creek, the surveys and the 50 redds found account for 14 female spawners or 66 percent of those released, but if the redds/female ratio were closer to 2, all of the females could be accounted for. The one to four redds/female range would produce salmon estimates of from 25 to 100 salmon. Since half the salmon carcasses were coho, the range for coho spawners would be 12 to 50.

In the combined areas of upper Hollowtree Creek, there were four jaw tagged carcasses. Of these, two were tagged on the last survey, so they had no opportunity to be recovered. For the two that had recovery surveys, both were recovered although one was not recovered the following survey but instead on the second survey, 20 days after tagging. With the CR model and a high carcass retention rate the model would estimate that there were 6 coho and 1 chinook. Similarly, AUC carcass method results in only 3 coho and 1 chinook. The AUC live estimate is also quite low, only 8 coho and 5 chinook. These estimates compare to 54 chinook and coho salmon being released. The majority of the fish estimated by carcass data are due to carcasses found on February 27th. These carcasses could have included, or have been totally due to, fish that passed the trap uncounted since the trap was removed 17 days earlier. One of those carcasses was identified as a female chinook, and since none were released, it must be assumed that it did pass by the trap site after the trap was removed.

As with the South Fork Noyo, population estimates in the upper Hollowtree Creek basin underestimated the population. The degree depends on what fraction of the spawning areas were surveyed. It

is not believed that more than 50 percent of spawning areas could have been missed. If only 50 percent were surveyed, then the number of spawners may have been 27 instead of 54. Even at that, the highest model estimate (based on live fish) is only 48 percent of this number, although 54 is well within the redd based range. When accounting for the fact that only two of the carcasses tagged would likely have been fish that were counted at Hollowtree Creek Hatchery, that, similar to the South Fork Noyo, only about 8 percent or less of the spawners were found as taggable carcasses. In the 1991-92 upper Hollowtree Creek survey, even though surveys covered less spawning area, 16 percent of the number of fish released were found as taggable carcasses.

For the spawners in the upper South Fork Noyo, a distance of 0 to 3 miles separated most spawners from a lower river section where carcasses would be counted against areas below the release site and a distance 4 miles further, spawners could enter areas not surveyed at all. But in upper Hollowtree, most of the spawning occurred three to eleven miles above the release site, malting the likelihood that fish backed down into areas below the release site less likely. In addition, areas below the release site are surveyed for a distance of nearly 8 miles and only 5 coho were estimated to spawn in this area based on carcass recovery. So clearly, in this case, the discrepancy can not be placed on movement of spawners out of the area.

It appears that there must be a difference in the retention rate of fish carcasses found by surveyors compared to those carcasses that are never seen. This is evidenced by the fact that both tagged carcasses in upper Hollowtree were recovered but only a small fraction of carcasses are ever seen. This may not be the case where relatively larger populations of spawners are found.

In this case, all females (the 21 coho and 5 steelhead) could be accounted for by a ratio of 2 redds per female but it is most

likely that some fish spawned in areas not surveyed and that some of these females were taken by predators prior to spawning. With these considerations, 2.5 redds/female would account for 20 females and would be believed to be a good estimate of the average number of redds produced.

For Hollowtree Creek the combined trap and lower river estimates produce a range in spawning escapements from 18 to 53 chinook and from 103 to 128 coho this year.

RESTORATION EVALUATION

Many areas within the Hollowtree Creek basin have undergone considerable restoration work. A major portion of this work has been barrier/log jam removal. There has also been a considerable amount of structure added such as logs and rootwads as well as bank stabilization projects. Much of this work has occurred in tributaries such as Bond, Bulter and Michaels Creek. The vast majority of chinook spawning in Hollowtree has occurred in the mainstem, mainly below the trap site. Only in one tributary, Butler Creek, were live chinook or carcasses found and then only on a single survey. Because chinook were not found to use Hollowtree Creek tributaries, restoration activities are not believed to be having a significant effect of chinook production.

Coho salmon typically spawn in smaller streams than chinook and it would be expected that Hollowtree Creek tributaries would be utilized by coho salmon. The only extensive surveys in Hollowtree Creek tributaries occurred in the initial survey year. Of 17 coho carcasses found throughout all of Hollowtree Creek in 1989-90, 5 (29%) were found in tributaries (4 in Redwood Creek and 1 in Huckleberry.) Only two redds were found in each of two tributaries Redwood and Huckleberry Creek in 1990-91 and only 8 to 10 coho are

estimated to have spawned in tribs the next year. Extensive restoration work has occurred in Bond and Butler Creeks but no coho carcasses were found in either. Not a single redd was found in Bond Creek in 1989-90 even though eight surveys were conducted between December 12th and February 27th. In 1991-92, there were 13 redds were found in late February but most of these were likely steelhead redds. Results from this survey do not indicate that tributaries were being extensively used by coho for spawning and that most spawning activity is occurred in the mainstem. The coho spawning in tributaries was primarily limited to Redwood and Huckleberry Creek. Both of these tributaries, especially Redwood Creek, have ideal rearing habitat for coho, typified by slow moving water and extensive amounts of cover. Even though the amount of spawning activity in Redwood and Huckleberry Creek was minor, the upstream rearing areas may be utilized by fry which move upstream looking for suitable habitat (Neave 1949).

The Hollowtree Creek Hatchery, with a release of about 190,000 fingerlings in the spring of 1988, contributed about 10 percent to the brood year production of chinook. This compares to releases of about 19,000 fingerlings from the 1985 and 1986 brood (CWT# B61515 and 065017) in Redwood Creek a tributary to the South Fork Eel which contributed 1.2 and 1.6 percent to the 3 and 4 chinook carcasses recovered in the South Fork Eel in the 1989-90 survey. The impact of a release of hatchery stock into a river is difficult to assess. If a river system is under-seeded, an increase in production will occur from additional spawners although if on the other hand the estuary or other river area is limiting production through food or space limitations, the overall production will not increase and could even decrease. Solazzi et al. (1990) found that increasing the rearing density of a stream with coho presmolts did not result in a change in the average spawning density of returning adults whereas supplemental feeding of under-yearling coho salmon increased the carrying capacity of a small Vancouver stream 6-7 fold (Mason 1976).

SUMMARY POPULATION ESTIMATION TESTING

The AUC live fish estimates based on an assumed 11 day spawning life of a coho salmon given in this report did produce closer estimates than did carcass models to the known population in upper South Fork Noyo drainage although in Hollowtree Creek it estimated only 8 compared to 6 where 54 fish were released. Carcass based estimates appear to be useful in that the estimates can be considered minimum estimates of spawning. The AUC carcass method has not received the review in the literature that the AUC live fish method has nor has the author found the method being used elsewhere. There is considerable variation between estimates of the two carcass models which is surprising since the carcass retention rate was used to determine the average carcass duration for the AUC model. Since the method used to determine the average carcass duration for use in the AUC method was not discussed by Beidler and Nickelson (1980) the method used in this report may not be appropriate for use in that model. For this reason the author believes that the CR model estimates are more valid than the AUC estimates derived here.

No point estimates for spawning populations are to be intended in this report and instead the ranges from a low determined from the CR model and a high based on the upper end of the redd based estimates are believed to encompass the likely range in the spawning populations for the streams surveyed. Additional study is needed to better determine the relationship between female spawners and the number of redds constructed. Also, a study of the retention rate of carcasses in streams where known numbers of carcasses were deposited would be of great interest. Perhaps in the South Fork Noyo where carcasses could be obtained from the egg-taking station, the carcasses could be distributed into various streams in a controlled study to see if the retention rates

determined in this study reflect the actual rates or if they are somehow biased.

CASPAR CREEK

Caspar Creek surveys are divided into four reaches: lower Caspar from mouth to confluence of the North and South Forks (3.0 miles), the South Fork from weir to mouth (0.5 miles), the North Fork from weir to mouth (1.9 miles) and the Middle Fork (0.5 miles). The weirs on the North and South Forks are operated by the U.S Forest Service but not as a fish counting station.

1991-92 SURVEY

The lower section, was surveyed 10 times from December 20th to March 9th. A total of eleven coho carcasses were tagged, eight of which were found on January 13th. Of The peak live count was 7 on January 8th. Of eight jaw tagged coho none were recovered in subsequent surveys. A total of 38 redds were found through January 21st and no more were found until the final survey on March 9th where 2 were counted.

In the North Fork, 8 surveys were conducted from January 3rd to March 9th. There was 1 king salmon, 9 coho, 2 steelhead and 1 unidentified carcass tagged. The peak live count was 8 on January 3rd. Redd counts were 54 in January and 5 in February. Of 13 jaw tagged carcasses, only 1 (8%) was recovered.

In the South Fork, 6 surveys were conducted from January 6th to March 9th. No live fish or carcasses were found. In January, 6 redds were counted and one more on February 24th.

In the Middle Fork, one survey on January 6th was conducted. No

evidence of fish use was found.

The king salmon, the first reported from Caspar Creek, was a female 78 cm in length. The measurable coho consisted of 5 adult males (averaging 62.8 and ranging from 58 to 66 cm.), and 4 adult females (averaging 64.3 and ranging from 62 to 67 cm.) Two female steelhead carcasses were 43 and 69 cm. in length; the larger was a five year old fish, the smaller couldn't be aged.

In total, out of 21 jaw tagged fish in Caspar Creek, only 1 (5%) was recovered. This is the lowest recovery for any stream surveyed. Roughly, these areas were surveyed on a six day interval. To obtain a 5% retention after 6 days, a carcass retention rate of 60 percent would apply. The CR model would estimate that 26 coho spawned in lower Caspar and 29 in the North Fork. In addition 4 king salmon, 4 unidentified fish and 7 steelhead carcasses would have also been deposited in the North Fork. No fish would be estimated in the South or Middle Forks.

The estimates for the AUG method was 33 fish in the lower section and 47 in the North Fork from carcasses and 9 and 6 in the lower section and North Fork based on live fish counts.

There 98 January redds. This would indicate that between 49 and 196 coho spawned in Caspar Creek this season of which two were likely chinook.

The estimate based on live fish is much lower than the estimate based on carcasses which is unusual. The carcass retention rate was also unusually low. If Caspar Creek has a population of predators which reduces carcass retention significantly, these predators may also reduce the spawning life of a coho significantly. To bring the population estimate up enough to approximate the other population estimates, the spawning life of a coho would have to be reduced from 11 to 3 days. Such a rate of

loss would indicate that many fish would probably be killed prior to spawning.

1990-91 SURVEY

In this year, the, lower section was surveyed 7 times from January 16th to February 25th. Only two carcass were tagged: a 72 cm. steelhead on February 18th and the other, unidentified to species, on February 25th. The steelhead had been shot with a gun. Neither of the tagged fish were recovered. A peak live count of 9 was made on February 25th, eight of which were identified as steelhead. Redd counts were 14 in January and 35 in February. One female coho was reportedly caught by tourist from Idaho on Caspar Beach on February 3rd as it tried to swim the bar at the mouth of Caspar Creek.

The North Fork was surveyed 4 times from January 21st to February 25th. No carcasses were tagged. A peak live count of 4 steelhead was made on February 11th. No redds were found in January, 20 were found in February. Surveys were conducted above the weir by U.S. Forest personnel. These surveys were conducted once a week from January 18th to April 29th from the weir to a distance 3524 feet upstream (Karen West, USFS, pers. comm., 1991). Live steelhead were observed from March 6th to April 29th. The peak count of live steelhead was 7 on April 17th. Most of the steelhead seen were in a small reservoir formed by the dam on the North Fork. Only one redd was reported in these surveys.

In the South Fork, 5 surveys were conducted from January 21st to February 25th. No evidence of fish use was found. Additional surveys were conducted above the South Fork weir by the U.S. Forest Service. Surveys included a distance of 2008 feet above the weir. They found a single fish on April 2nd. No other evidence of fish use was found.

Since only two carcasses were tagged and since they were both steelhead, no coho would be estimated to have spawned. From the 14 redds in January, between 4 and 7 females or between 8 and 28 coho spawned. Based on live fish, the AUC method would estimate 2 coho. Surveyors in Caspar Creek did note lots of otter and racoon tracks and did find bear sign in one survey.

RESTORATION EVALUATION

In 1960 Federal and State agencies initiated a study of long-term effects of logging and road building on stream flow, sedimentation, aquatic habitat, and fish populations on two watersheds of Caspar Creek - the North and South Forks (Krammes and Burns 1973). No restoration projects have occurred in the drainage since the early 1970's which would interfere with on-going study (Fay Yee, U.S. Forest Service per. comm. Ft. Bragg, 1993).

DEHAVEN, HOWARD AND WAGES CREEK 1990-1991

All three of these streams (Figure 1.) were planted with yearling coho in the Spring of 1988. The purpose of planting these Noyo River coho stocks in these drainages was to reintroduce coho into these streams since they no longer had native coho runs. These spawning surveys were conducted for the purpose of determining whether or not adult coho were returning to spawn.

1990-91 SEASON

DeHaven Creek

DeHaven Creek was surveyed from the mouth to three miles upstream. There were three surveys from January 17th to February 4th. There was no evidence of spawning activity during any of these surveys.

Howard Creek

Howard Creek was surveyed three times between January 17th. and February 4th. The distances ranged from two to three miles. As for DeHaven, no evidence of fish use was found. The stream was described as having a lot of large rock and no areas considered to be good spawning gravel.

Wages Creek

This stream was surveyed five times between January 17th. and February 13th. There were two carcasses tagged; one a coho on February 4th and two steelhead on the 13th. The peak live count was two steelhead on February 13th. Four redds were found in January and eight in February. Surveys were done from the mouth to the forks, a distance of three miles. One survey was continued an additional mile upstream. Only one jaw tagged carcass had a recovery survey and it was subsequently recovered for a 100 percent recovery rate. There was bear sign along this stream.

Based on January redds, there were 2 to 8 coho spawning in this stream. The single carcass would have estimated no more than one pair of coho spawners.

RESTORATION EVALUATION

It appears that only in Wages Creek was the release of coho smolts successful. Wages Creek flows were greater than DeHaven or Howard Creek. This may have allowed the returning salmon to enter Wages Creek where they may not have been unable to enter the other streams. Even though Wages Creek did have coho spawning, the number was very low. In Howard and DeHaven Creeks, this planting effort was not successful.

LITTLE RIVER

Little River flows through Van Damme State Park in Mendocino County. There is a road that follows the stream through the park which provides easy access to the stream. The stream surveys were conducted from the mouth to four miles upstream.

1991-92 SEASON

There were nine surveys conducted from December 31st. to March 9th. There were eleven coho carcasses tagged. The peak live count was eleven on January 5th. Redd counts were 16 in January and 9 in February. Of the 7 jaw tags applied (one additional was tagged on the last survey), 4 (57%) were recovered. There was one female (63 cm.), three males (60 to 61 cm.) and one male grilse (38 cm.) coho found during surveys. It was noted that at least one redd was out of the water because of low flow conditions.

The CR model would estimate that 14 coho spawned in Little River this year. The AUC carcass and live fish estimates would be 9 and 23 coho respectively. Based on redd counts in January, 8 to 32 coho would have spawned.

The recovery rate of carcasses on Little River was relatively high compared to other streams. This would indicate a low rate of predation and/or scavenging on this stream.

There has been, concern that culverts along the creek are an impediment to fish passage and the State Park personnel had sandbagged several of these to improve passage prior to the spawning run starting. In the first survey, a redd and a live female coho were, found Just below the upper culvert indicating that passage was possible at all the other culverts. The other redds

were below the sixth culvert down from the upper end. A survey on January 28th also found fish just below the upper culvert. It appears that the sandbagging did allow fish to pass the culverts. While no redds were found above the upper culvert, it did not appear as if it would have prevented fish passage.

1990-91 SURVEY

Little River was surveyed three times from January 16th to February 9th. Only one redd was found. It was found on the last survey and was located between 3 and 3.5 miles upstream. Previous surveys had not surveyed this far upstream and the redd appeared not to be fresh so it is believed that this was a January redd. Only one pair of coho could have been expected to have spawned. No carcasses or live fish were seen.

RESTORATION EVALUATION

Adult spawning migrations are believed to have been hindered in Little River by a series of culverts. The spawning survey in 1991 92 found fish above all but the upper culvert and here it appeared that fish could have passed this culvert without difficulty. It appeared that the modifications that were made prior to the 1991 92 survey allowed fish to pass with the flows as they existed. There appeared to be very little quality spawning gravel in Little River. The stream was described as consisting of fast water with a large component of the substrate consisting of boulders. The low coho population estimate of 1 to 3 fish in 1989-90 (Nielsen et al., 1990) and only 2 in 1990-91 indicates very poor production levels. The lack of low velocity habitat in all but the very lowest section of the creek probably limits the coho rearing potential of this stream.

S. F. GARCIA AND L. N. F. GUALALA RIVER 1990-1991

Both the South Fork Garcia and Little North Fork Gualala Rivers were planted with yearling coho in the spring of 1988. The purpose for planting these fish (Noyo River stock) was to reintroduce coho into these areas since native stocks no longer existed (Weldon Jones, CDFG, per. comm., 1989). These surveys were undertaken to find out whether coho were returning from this effort.

The South Fork Garcia River was surveyed 3 times in February 1991 from the 1st to the 15th. No live fish or carcasses were seen. Two redds were observed, one on each of the last two surveys. An additional redd was reported on the Garcia River 100 feet below the mouth of the South Fork.

The Little North Fork of the Gualala River was surveyed 3 times in February 1991 from the 1st to the 15th. No live fish or carcasses were seen. Only two redds were observed in the Little North Fork, these on the 8th, but on the 15th, five redds were found on the North Fork just downstream from the mouth of the Little North Fork. These redds were, in all likelihood, from fish that would have spawned in the Little North Fork but did not due to the low water conditions that existed at the time.

Common to both the S.F. Garcia and L. N. F. Gualala were reports of high numbers of juvenile steelhead. Schools of 20 to 100 fish 4-7 cm. in length were common. Also, low flows prevented spawners from entering both these streams during the 1990-91 season. No surveys were conducted in 1991-92. It appears that even though these streams have not been producing coho salmon, steelhead production is relatively high. There were as many as 33 redds in

the SF Garcia and 17 in the LNF Gualala in 1989-90 (Nielsen et al., 1990), many of which were in the month of January. There were at the same time several steelhead carcasses and live fish observed while there was no evidence of coho spawning. Together, the juvenile populations and spawning steelhead in these streams indicate that steelhead production in these streams is quite good.

Only one pair of fish is estimated to have spawned in these drainages this year. Since these fish did not spawn until February, it is assumed that these spawners were steelhead. In both streams, there was spawning activity just below the mouths of these streams, indicating that there would have been more fish spawning had the flow conditions been better.

RESTORATION EVALUATION

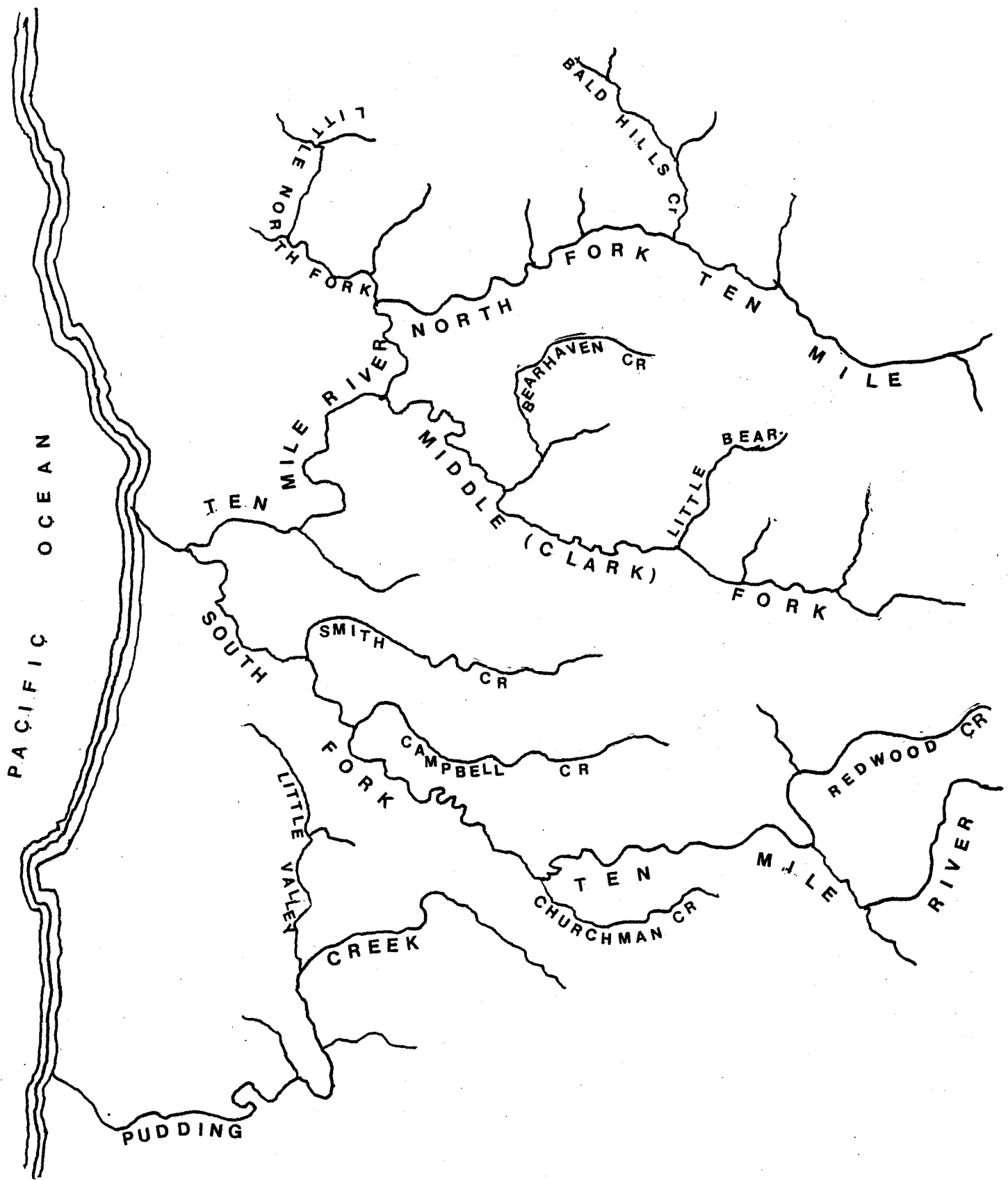
Both of these streams have been planted with yearling coho in the Spring of 1988. There were only two redds found in each stream this year and additional redds were found just below the mouths of these streams. This spawning activity is not believed to be due to returning adult coho from the yearling releases since they were not found until the second February survey.

TEN MILE RIVER

1991-92 SURVEY

The Ten Mile River (Figure 5.) surveys were done sporadically this season. This was due to the low priority rating given the Ten Mile by the California Department of Fish and Game. Surveys were conducted only if the other stream surveys had been completed in

Figure 5- Ten Mile River and Pudding Cr.



each survey period. The upper reaches of the South Fork and North Fork and associated tributaries were not surveyed nor were Mill, Redwood, Churchman, or Little Bearhaven Creek. Most areas that were surveyed were only surveyed once or twice. With this low effort, no population estimates could be made except with the redd based method.

The mainstem of the Ten Mile River from the confluence of the Middle (Clark) Fork down was surveyed twice; once on December 20th and once on January 22nd. The first survey extended from Baxman's Gravel Plant to Clark Fork (3.0 miles) and from the South Fork to Clark Fork the second survey. No fish, redds or carcasses were found the first survey. On the second, 3 steelhead and 3 other unidentified fish were observed and 10 redds were found.

North Fork

The North Fork was surveyed from Clark Fork to the Little North Fork (1.9 miles) twice, once on December 20th and once on January 22nd. No fish, redds, or carcasses were observed the first survey but a coho and two steelhead were tagged on the 22nd and three redds were counted. The Steelhead was an 89 cm. male at least six years of age.

The area between the Little North Fork (LNF) and Bald Hills Creek (5.2 miles) was surveyed three times between December 23rd and January 29th. The last survey covered only the lower 2.5 miles of this reach. Three redds were found on the first survey. Another 7 redds were found in later surveys and a single live steelhead was observed.

One survey was conducted above Bald Hills Creek for a distance of 2.3 miles on January 29th. No fish, redds, or carcasses were found although water visibility was not ideal.

In all, 23 redds were found in the area from 2.3 miles above Bald Hills Creek to the South Fork and a single coho carcass was found. This would indicate that 11 to 46 salmon spawned in this area.

Clark Fork

The Lower section of Clark Fork from the mouth to Bearhaven Creek (3.1 miles) was surveyed once on the 22nd of December. One redd was found. This section was surveyed again one month later and one coho was tagged and six redds counted. A final survey on February 29th found a single live steelhead and 2 additional redds.

The area above Bearhaven Creek (10.1 miles) was surveyed in mid to late January. In this area, for three survey dates combined, two coho, one steelhead and eight chinook carcasses were tagged. There were 25 redds counted but no live fish observed. Two carcasses were measurable; one was a 103 cm. male and the other a 96 cm. female chinook (5 years of age). On the 28th and 29th of February, 7.1 miles of this area was again surveyed. In this survey 34 redds were counted and a single live steelhead observed. No carcasses were found.

In January, there were 32 redds in the Clark Fork and there were 3 coho and 8 chinook carcasses. The redd count would indicate that from 16 to 64 salmon spawned.

South Fork

The South Fork was surveyed on January 17th and 27th from Smith Creek to 1.3 miles above the confluence of Redwood Creek (13.8 miles). Two chinook and one coho were tagged. There were 13 redds counted. The peak live count was on the 27th. These were steelhead (7) in a pool just above the lower end of the survey area. The January redds would estimate between 6 and 26 salmon spawned. In

February, no surveys were conducted.

Campbell Creek

Two surveys were conducted in Campbell Creek in January; one on the 7th and one on the 26th. The first covered the lower 1.6 miles and the second the lower 0.8 miles. One redd was found on each survey. A coho skeleton was found which would indicate that there was coho spawning. No survey was conducted in February. A single pair of spawning salmon would be estimated.

Smith Creek

A single survey was conducted in the lower 2.8 miles on January 27th. No fish or signs of spawning activity were found.

Bald Hills Creek

A single survey was conducted in the lower 1.9 miles on January 29th. A single redd was found, or one pair of fish spawned.

Little North Fork

The Little North Fork was surveyed three times in January. The first survey covered the lower 1.8 miles, the second 2.7 miles and the last only the lower 0.3 miles. A total of nine redds were counted. The peak live count was three coho, one adult and two grilse, on the 29th. An 84 cm. 4-year old male chinook carcass was tagged and a coho skeleton seen. The redd count would indicate that between 5 and 18 salmon spawned.

In February, a survey was conducted on the 13th and the 28th. One live steelhead was seen on each survey and 21 redds were counted. The first February survey extended from the mouth upstream 3.5

miles and the second 3.0 miles. Since no live salmon or salmon carcasses were found, all February redds are considered steelhead redds.

Bearhaven Creek

In January, a survey on the 14th found 17 redds in the lower 2.5 miles. An survey on the 24 surveyed a mile above the previous survey; one redd was found. One live coho was found on the 14th. These 18 redds estimate 9 to 36 spawners.

Two surveys were conducted in February. One steelhead and one unidentifiable carcass were tagged. No live fish were seen and twelve redds were counted.

Ten Mile River System

In the Ten Mile River there were 98 January redds. This would estimate between 49 and 196 salmon spawners. There were 11 chinook and 3 coho tagged. Assuming that this is the relative abundance of the two species, the chinook run would be between 38 and 154 and the coho run between 11 and 42. It doesn't appear that chinook females dig more than 2 redds based on Hollowtree Creek data whereas coho appear likely to dig between 2.5 to 3.5 redds. From this information and the high proportion of chinook the lower portion of the range appears unlikely and that minimal estimates would be better estimated by a 3 redds/female estimate. At this ratio, the lower estimate for chinook and coho would be 51 and 14, respectively.

1990-91 Surveys

No surveys were conducted in the 1990-91 season.

RESTORATION EVALUATION

Surveys in 1989-90 and 1991-92 surveyed a substantial proportion of the watershed; about 70 and 55 miles respectively. These compare to an estimated total area of salmon habitat of 103 miles (California Wildlife Plan, 1965). The only area not surveyed in 1989-90 which likely had spawning activity was in the upper Clark Fork where in 1991-92 half of all chinook carcasses in the Ten Mile River system were found. In 1991-92, even though fewer miles were surveyed, surveys of spawning areas were fairly complete because they included the upper Clark Fork and all of the areas where salmon carcasses were found in 1989-90.

Chinook salmon are not considered to be native to the Ten Mile River although chinook salmon have been reported caught in the river several decades ago. Chinook salmon were introduced into the Ten Mile River in the early 1980's. The last major introduction of chinook was in 1982. Some eggs were taken from chinook trapped in the Ten Mile River in the mid-1980's. The last and largest release from this group was 9000 fingerlings released in the spring of 1987. The chinook carcasses found in the Ten Mile River have been composed of various age groups and the four-year-old chinook found in 1991-92 were undoubtedly of natural origin. This indicates that the chinook introduced into the Ten Mile River are successfully reproducing. While these runs are not large, from 34 to 54 in 1989-90 and 51 to 154 in 1991-92, they are widely scattered throughout the basin, being found in the Little North, South, Clark and North Forks. In addition the author observed chinook spawning in upper Clark Fork in the winter of 1993. Together, these observations indicate successful natural production of chinook salmon in the Ten Mile River. If this natural production continues, this would be one of the few successful introductions of a salmon species into a watershed along the west coast (Withler 1982).

Oregon coho stocks were planted into Ten Mile in the early 1970's. From 1974 to 1977 approximately 200,000 coho were reared and released into the Ten Mile River annually (Taylor 1978). The last of these imports of out of basin coho stocks was in 1978 when 44,000 fish were released. The only other release of coho salmon was 6,000 fingerlings in June 1987 which were the offspring of coho trapped in the Ten Mile River. Fifteen male and ten female coho were trapped that year (Salmon Restoration Ass. unpublished data).

The coho run estimates ranging from 32 to 52 in 1989-90 and 14 to 42 in 1991-92 are quite low for the size of the Ten Mile River. In the early 1960's, the Ten Mile River was estimated to a coho run of 6,000 fish (California Wildlife Plan 1965). The effort to restore this run in the mid-1970's by artificial propagation was unsuccessful due the inappropriateness of the Oregon coho stocks propagated and also the habitat problems and limitations that existed.

The California Wildlife Plan (1965) considered fishery habitat conditions in the Ten Mile River to be severely degraded by logging activity. The Ten Mile River has undergone extensive restoration activities primarily in the form of barrier removal. Areas surveyed that have had extensive work are the upper South Fork, Bearhaven and Bald Hills Creek. In the surveys conducted, the upper South Fork and Bald Hills Creek show very little salmon spawning activity. No evidence of salmon spawning was found in the South Fork above the confluence of Redwood Creek even though habitat appeared to in good condition and spawning activity in the 5 mile reach below this area was sparse; only two redds in January 1991 92 and only 1 in 1989-90. Bald Hills Creek had a single redd in 1991-92 and none in 1989-90. Bearhaven Creek had a relatively healthy coho run in 1991-92 but had only three redds in January 1989-90. Of seven sites samples for juvenile populations in the Ten Mile River in October 1991 only two had coho present (Maahs, M.

unpublished report, Salmon Restoration Ass. 1992). Bearhaven Creek had a density of 0.08 coho/m² and Bald Hills Creek 0.01 coho/m².

It appears that coho in the Ten Mile River have been unable to recover from low population levels even though extensive restoration work has been conducted. The upper reaches of the South Fork and North Forks did not see salmon spawning during the two survey years nor likely in 1990-91 because of the low water conditions. For these areas to be reseeded the Ten Mile will need to have large escapement levels that allow sufficient straying or artificial propagation projects. The low population level of coho found in these Ten Mile River surveys has prompted a trapping and artificial rearing project in the Ten Mile which hopefully can help restore coho stocks in the basin.

PUDDING CREEK

Pudding Creek surveys were divided into two sections. The lower section extends from the railroad tunnel upstream to the confluence of Little Valley Creek (4.2 miles). The upper section extends upstream from Little Valley Creek for 4.1 miles. The upper end of the upper reach is very brushy and difficult to survey often requiring surveyors to get out of stream and try to find ways around brush. Areas farther upstream were not surveyed primarily because of the difficulty of this area to survey. While bad for surveys, these areas provide good habitat for rearing and spawning. The water in the lower section is often cloudy because of the addition of water from Little Valley Creek. This seems to be a natural condition which exists year-around. This discoloration reduces visibility and probably reduces the chances of finding carcasses and live fish. Because of the very poor visibility, Little Valley Creek was not surveyed. This creek was surveyed once

in 1989-90. Besides the poor visibility, the creek was quite deep and slow moving with large amounts of woody material. Very little spawning habitat was found in the lower 0.8 miles that was surveyed. The survey conditions in Pudding Creek probably result in surveys that miss more carcasses, live fish and redds than in other survey areas.

1991-92 SEASON

Lower Pudding Creek was surveyed seven times from January 2nd. to March 13th. Ten coho carcasses were tagged. The peak live count was 17 (11 coho and 7 steelhead) on February 3rd. The redd counts were 31 in January and 23 in February. Of 9 jaw tagged coho carcasses there were four adult males (averaging 68 cm. and ranging from 61 to 77 cm. fork length. None of the tagged carcass were recovered.

In the upper section on Pudding Creek, surveys were conducted five times from January 9th to March 12th. Only one coho carcass was tagged. The peak live count was on February 5th where one coho and three steelhead were observed. There were 9 redds in January and 18 in February.

Since none of the nine jaw tagged carcasses in lower Pudding Creek were recovered, estimating the retention rate is somewhat subjective. A retention rate of .74 was selected for use in modeling. This is the retention estimate selected the for South Fork Noyo in 1990-91 where out of 30 tags, six were recovered. At this rate, the CR model would estimate 28 spawning coho. The AUC carcass estimate is 45 coho and the live AUC estimate is 37. The live fish estimate is lower than the AUC carcass estimate but higher than the CR estimate.

The 31 redds in lower Pudding Creek in January would estimate from

16 to 62 spawning coho salmon. In February, 64 percent of live fish counts were coho. At this rate, 15 of the 23 February redds would also be coho redds, bringing the redd based range from 23 to 92 coho.

In the upper section, 25 percent of live courts in February were coho, bringing the coho redd estimate in the upper section to 5. Coho spawners in upper Puddings Creek would be estimated to range between 3 and 10 fish. The AUC live fish estimate is 3 coho. No coho would be estimated to have spawned with carcass data since no fish were tagged with jaw tags.

The estimates for all surveyed areas of Pudding Creek are between 28 and 45 for AUC and CR models and from 26 to 102 coho based on redd counts.

1990-91 SEASON

Lower Pudding Creek was surveyed seven times from January 17th to February 25th. There were 11 tagged carcasses of which 5 were coho, 4 steelhead and 1 unidentified. The peak live count was 12 on February 11th of which half were coho and half steelhead. There were two redds in January and 54 in February. Out of seven jaw tagged carcasses, 3 (43%) were recovered. There were four measurable carcasses, one was a 70 cm. male and two were 45 and 66 cm. coho that the sex could not be identified. There was also a 80 cm. steelhead estimated to be four years old with three years of ocean growth.

In upper Pudding Creek, surveys were conducted three times between the 6th and 25th of February. Two coho carcasses were tagged and the peak live count was 4 coho on the first survey. There were 11 redds found. Only one of the jaw tagged fish had a recovery survey and it was recovered for a 100% recovery rate.

Estimates of spawning in lower Pudding Creek are 8, 7 and 9 coho based on CR model, AUC carcass and AUC live counts, respectively. Population estimates for coho based on January redd counts are one pair. Some proportion of the 54 redds in February are due to coho spawning. Of the February live fish counts 46 percent were coho. This would indicate that 25 of these redds were dug by female coho, and that between 12 and 50 coho spawned in February. The additional pair of coho spawning in January brings the range to 14 to 52 coho spawners.

In upper Pudding Creek the CR model, with a 92 percent retention rate, would estimate 3 coho and 3 coho is the AUC carcass estimate. The AUC live model estimates is 2 coho (even though there were actually four seen). There were 11 redds and all live fish and carcasses were coho. The redd based range is from 6 to 22 coho spawners.

The total Pudding Creek estimate for 1990-91 is 10 to 11 based on CR and AUC methods and 20 to 74 based on redd counts.

The difference between carcass and redd based estimates are reminiscent of the upper South Fork Noyo and Hollowtree Creek where the populations of coho were known. The low AUC live fish estimate in relation to the carcass estimates is likely due to the poor visibility and large amount of cover which is inherent with Pudding Creek surveys.

Restoration Summary

The Pudding Creek coho run estimates varying from 6 to 50 in 1989 90, 20 to 74 in 1990-91 and 26 to 102 in 1991-92 are considerably

where from 1957-58 through 1960-61 (the first four years of operation) minimal counts of 1257, 628, 442 and 484 were made (Strohschein 1961). The dam and associated reservoir has probably impacted coho production in Pudding Creek. The fish ladder is not operational until rains have raised the reservoir level. This can prevent an early run component from spawning in Pudding Creek. A similar structure would have prevented the spawning activity that occurred in Caspar Creek and South Fork Noyo in 1990-91 where some spawning activity occurred prior to significant rainfall. The heavy growth of aquatic plants and exotic warm water fish in the reservoir would not indicate advantageous conditions for coho production. Pudding Creek above the impounded water appear to provide very good rearing habitat as does Little Valley Creek.

Pudding Creek has had several log jam removal projects. These jams were not barriers to fish migration and overall these projects have not improved this streams capacity to produce coho salmon.

WILLITS CREEK

Willits Creek was surveyed only once on January 11th, 1992. No carcasses, redds or live fish were found. No fish were found in this area during the 1989-90 surveys as well. Low flow conditions have likely impacted spawning runs in this area.

CONCLUSIONS

The results of these spawning surveys indicate that coho runs are presently much smaller (if they exist at all) than they were in the early sixties even though extensive restoration work has been done since that time. Two primary factors can explain the apparent low abundance of coho spawners. Either inadequate numbers of spawners have been available to repopulate these areas since

restoration or the restoration activities have been generally ineffective at improving the habitat; restoration activities may even have had negative effects overall by removal of the primary habitat needed by coho salmon.

In streams which empty directly into the ocean such as Howard, DeHaven and Wages Creek where coho runs were lost probably due to the aggravated conditions of poor logging practices and drought nothing short of transplants will bring back coho production in these streams. On the other hand the Ten Mile River appears that it could sustain larger coho production than it presently has and that increased numbers of spawners may increase coho production. Ten Mile River does lack habitat that would be considered ideal coho habitat but generally would be considered to have some good habitat. But for other streams surveyed this is not necessarily true.

In Little River, even in the year where only a single redd was observed, young-of-the-year (YOY) coho are found outmigrating and in Caspar Creek YOY coho are also regularly captured in outstream-migrate traps (Weldon Jones, CDF&G, per. comm. 1993). Mason and Chapman (1965) found a positive correlation between levels of aggression and the rate of emigration of YOY coho in streams and that levels of aggressive and territory size of rearing coho was effected by the abundance of food and that sustained higher levels of food brought about greater density (reduced migration) of fish. LeCren (1965) states that territorial behavior acts as a density-determining mechanism in salmonids and Chapman (1962) found that coho fry migrating downstream in the spring were smaller than residual coho and that downstream movement would cease if fish were placed in a suitable environment. Together these studies indicate that downstream migrations of coho fry in Little River and Caspar Creek indicate that insufficient habitat quantity or quality is available. Various studies (Shapovalov 1954; Crone et al. 1966; Crone and Bond 1976; Hartman et al. 1982; Washington 1991) have

found that unless emigrating coho reach one year of age and sufficient size coho juveniles will not survive to become adults. There is considerable habitat below the outmigrant trap site on Caspar Creek and some number of these YOY migrants may find suitable habitat to rear. In Little River very little rearing habitat is available below trapping site.

Presently, many coastal streams have high levels of fines in the sediment resulting from poor land use practices. Increased levels of fines have been shown (Pennak & Van Gerben, 1947; Sprules, 1947; and Kimble and Wesche 1975) to reduce the aquatic life which effects juvenile coho's ability to feed which in turn increases territory size of coho fry and reduces the carrying capacity of the stream. Stream restoration activities had concentrated for years of the removal of barriers to salmon migration and in the process has removed the vast majority of log jams and associated habitat which tend to produce the highest densities of coho (Hartman, 1959). In general, where juvenile salmonid cover is removed salmonid abundance declines (Boussu 1954; Peters & Alvord 1964; Elser 1968). How food abundance and habitat conditions can limit coho populations are demonstrated in studies attempting to increase coho production. Mason (1976) found that by supplementing food in a natural stream, the density of coho that could rear was increased 7 fold but, that after supplemental feeding, the populations dropped back to normal levels due to habitat limitations. Mason (1976) also found that outmigration was directly related to stocking density and that survival, growth and biomass yield were inversely related to stocking density in unfed populations. In another study where coho presmolts were released into streams, the densities of juvenile coho increased but the supplementation did not result in a change in the average spawning density of returning adults (Solazzi, 1990). Shapovalov and Taft (1954) found that increased egg deposition resulted in increased population sizes of downstream migrant coho, decreases in the average size of emigrates, reduced survival rates and increased

rate of straying among returning adults.

The degree to which density-dependant factors limit coho salmon production in streams is not often studied although many studies have been done which indicate density-dependent factors, are influencing production. Currently, ocean salmon fisheries harvests are being significantly reduced to provide increased numbers spawning coho salmon into Oregon coastal streams to enhance the production of coho salmon (PFMC 1993). The success of this effort depends upon the availability of under-utilized coho rearing habitat throughout the juvenile rearing stage prior to emigration to the marine environment. Greater efforts need to be made to determine where and to what extent density-dependent factors in the fresh-water environment are limiting production. The degree to which YOY coho are emigrating the streams may be a good indicator of limitations in the first several months of the coho rearing period. Since these juvenile coho are not expected to survive and become adults, these fry could be used to stock areas believed to be underseeded.

LITERATURE CITED

- Beidler, W. A. and T. E. Nickelson. 1980. An evaluation of the Oregon Department of Fish and Wildlife standard spawning fish survey system for coho salmon. Oreg. Dep. Fish Wildl. Info. Rep. Ser. 80-9: 23p.
- Boussu, M. F. 1954. Relationship between trout populations and cover on a small stream. J. Wildl. Manage. 18:227-239.

- Briggs, J. S. 1953. The behavior and reproduction of salmonid fishes in a small coastal stream. California Fish & Game Bull. No.94: 62p.
- Brown, L. R. , P. B. Moyle, and R. M. Yosiyama. 1993. Status of coho salmon in California. North Amer. J. Fish. Man.
- Bustard, D. R., and D. W. Narver. 1975. Preferences of juvenile coho salmon and cutthroat trout relative to simulated alteration of winter habitat. J. Fish. Res. Board. Can. 32:681-687.
- Conte, F. P., H. H. Wagner, J. Fessler, and C. Gnose. 1966. Development of osmotic and ionic regulation in juvenile coho salmon. Comp. Biochem. Physiol. 18:1-15.
- Crone, R. A., and C. E. Bond. 1976. Life history of coho salmon. in Sashin Creek, Southeast Alaska. Fish Bull. 74:897-923.
- Elser, A. A. 1968. Fish populations of a trout stream in relation to major habitat zones and channel alterations. Trans. Am. Fish. Soc. 97(4):389-397.
- English, K.K., R.C. Bocking & J.R. Irvine (1992). A robust procedure for estimating salmon escapement based on the area-under-the-curve method. Ca. J. Fish. Aquat. Sci. Vol. 49:1982-1989.
- Hartman, G. F. 1965. The role of behavior in the ecology and interaction of under-yearling coho salmon and steelhead trout. J. Fish. Res. Board Can. 22:1035-1081.

- Hartman, G. F., B. C. Anderson, and J. C. Scrivner. 1982. Seaward movement of coho salmon fry in Carnation Creek, an unstable coastal stream in British Columbia. Can. J. Fish. Aquatic Sci. 39:588-597.
- Kammes, J. S. and D. M. Burns. 1973. Road construction on Caspar Creek watersheds. Ten year impact report. Pac. Southwest Forest and Range Station. USDA Forest Ser. Res. Paper PSW-93
- Kimble, L. A., and T. A. Wesche. 1975. Relationship between selected physical parameters and benthic community structure in a small mountain stream. Univ. Wyo. Laramie. Water Resour. Res. Inst. Series 55. 64 p.
- McMahon, T., E., L. B Holtby. 1992. Behavior, habitat use, and movement of coho salmon smolts during seaward migration. Can J. Fish. Aquat. Sci. 49:1478-1485.
- Moring, J. R., and J. R. Lantz. 1974. The Alsea watershed study: effects of logging on the aquatic resources of three headwater streams of the Alsea River, Oregon. Part 1- Biological studies. Oreg. Dep. Fish Wildl., Fish. Res. Rep. No. 9.
- Neave, F. 1949. Game fish populations of the Cowichan River. Fish. Res. Board Ca. Bull. 4. 76p.
- Nickelson, T. E., J. D. Rodgers, S. L. Johnson and M. F. Solazzi. 1992. Seasonal changes in habitat use by juvenile coho salmon in Oregon coastal streams. Can. J. Aquat. Sci. 49:783-789.

- Nielsen, J. L., M., Maahs, & G. Balding. 1990) Anadromous salmonid resources of Mendocino coastal and inland rivers 1989 1990. An evaluation of rehabilitation efforts based on carcass recovery and spawning activity. Report submitted to California Dept. of Fish and Game. Fisheries Restoration Program, Work Progress Report 1990. 98 p.
- Norton, J., D. McLain, R. Brainard & D. Husby (1985). The 1982-83 El Nino event off Baja California and its ocean climate context. p. 44-72. In W.S. Wooster & D.L. Fluharty (ed.), El Nino North. Washington Sea Grant Program, Univ. of Wash.
- Perrin, C.J., and J.R. Irvine. 1990. A review of survey life history estimates as they apply to the area-under-the-curve method the spawning life escapement of Pacific salmon. Can. Tech. Rep. Fish. Aquat. Sci. 1733: 49 p.
- Pennak, R. W., and E. D. Van Gerpen. 1947. Bottom fauna production and physical nature of the substrate in a northern Colorado trout stream. Ecology 28:42-48.
- Peters, J. C., and W. Alvord. 1964. Man-made channel alterations in thirteen Montana streams and rivers. Trans. North Am. Wildl. Conf. 29:93-102.
- Pacific Fisheries Management Council. 1993. Draft amendment 11 to the Fishery Management Plan for commercial and recreational salmon fisheries off the coasts of Washington, Oregon, and California commencing in 1978. Portland, Oregon.
- Ruggles, C. P. 1966. Depth and velocity as a factor in stream rearing and production of juvenile coho salmon. Can. Fish Cult. 38:37-53

- Shapovalov, L. & A.C. Taft. 1954. The life histories of the steelhead rainbow trout (*Salmo gairdneri gairdneri*) and silver salmon. Cal. Dep. Fish & Game, Fish Bull. No. 98.
- Snyder, R. U. and S. W. Sanders. 1979. Summary of operations Noyo River egg collecting station 1964-65 through 1977-78. Anad. Fish. Branch Admin. Rep. 79-7.
- Solazzi, M. F., T. E. Nickelson and S. L. Johnson. 1990. An evaluation of the use of coho salmon presmolts to supplement wild production in Oregon coastal streams. Oreg. Dep. Fish. Wildl. Rep. No. 10. 22 p.
- Taylor S. 1978. California's private rearing pond programs. Cal. Dep. Fish & Game. Sacramento, 18 p.
- Willis, R. A. 1954. The length of time that silver salmon spent spent before death on the spawning grounds at Spring Creek, Wilson River in 1951-52. Fish Comm. Oreg. Res. Briefs 5:27-31.